

# **Aquatic Species Biological Assessment for the Bear and York On/Off Allotments**

**BLUE MOUNTAIN RANGER DISTRICT**

**MALHEUR NATIONAL FOREST**

**GRANT COUNTY, OREGON**

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Signature: /s/ Ian Reid

Date: October 14, 2011

AQUATIC BIOLOGICAL ASSESSMENT  
(9/21/2011) for  
Endangered Species Act Section 7 and  
Magnuson-Stevens Act Consultation

TABLE 1. ESA ACTION AREA HYDROLOGIC UNIT CODE NAMES AND NUMBERS

Bear Allotment

4 <sup>th</sup> Fields	4 <sup>th</sup> Field HUC Numbers	6 <sup>th</sup> Fields	6 <sup>th</sup> Field HUC Numbers
Middle Fork John Day	17070203	Bear Creek-Middle Fork John Day River	170702030301
5 <sup>th</sup> Fields	5 <sup>th</sup> Field HUC Numbers		
Big Creek-Middle Fork John Day	1707020303		

York On/Off Allotment

4 <sup>th</sup> Fields	4 <sup>th</sup> Field HUC Numbers	6 <sup>th</sup> Fields	6 <sup>th</sup> Field HUC Numbers
Middle Fork John Day	17070203	Slide Creek	170702030304
5 <sup>th</sup> Fields	5 <sup>th</sup> Field HUC Numbers		
Big Creek-Middle Fork John Day	1707020303		

**Name and Location of Administrative Unit:** Blue Mountain Ranger District, Malheur National Forest, P.O. Box 909, John Day, OR 97845

TABLE 2. ESA & EFH EFFECT DETERMINATIONS (ENDANGERED (E) AND THREATENED (T) SPECIES)

Bear Allotment

Common Name	Scientific Name	Known to Occur?	Potential to Occur?	ESA <sup>1</sup> Status	Critical Habitat Status	Essential Fish Habitat Status	ESA <sup>1</sup> Effect Determination Species/CH <sup>2</sup>	EFH <sup>3</sup> Effect Determination
<b>MCR steelhead</b>	<i>Oncorhynchus mykiss</i>	Yes	Yes	Threatened	Designated	Not applicable	NLAA <sup>3</sup> /NLAA	Not applicable
<b>Spring Chinook</b>	<i>Oncorhynchus tshawytscha</i>	Yes	Yes	Not warranted	Not applicable	Established	Not applicable/Not applicable	Will not adversely affect
<b>CR bull trout</b>	<i>Salvelinus confluentus</i>	No	Yes (migration/winter rearing)	Threatened	Designated	Not applicable	NLAA <sup>4</sup> /NLAA	Not applicable

York On/Off Allotment

Common Name	Scientific Name	Known to Occur?	Potential to Occur?	ESA Status	Critical Habitat Status	Essential Fish Habitat Status	ESA Effect Determination Species/CH	EFH Effect Determination
<b>MCR steelhead</b>	<i>Oncorhynchus mykiss</i>	Yes	Yes	Threatened	Designated	Not applicable	LAA/LAA <sup>5</sup>	Not applicable
<b>Spring Chinook</b>	<i>Oncorhynchus tshawytscha</i>	No	No, but occur downstream in Slide Creek	Not warranted	Not applicable	Established	Not applicable/Not applicable	May adversely affect
<b>CR bull trout</b>	<i>Salvelinus confluentus</i>	No	No	Threatened	Designated	Not applicable	No effect / No effect	Not applicable

<sup>1</sup> Endangered Species Act

<sup>2</sup> Critical Habitat

<sup>3</sup> Essential Fish Habitat under the Magnuson-Stevens Fishery Conservation and Management Act

<sup>4</sup> May Affect, Not Likely to Adversely Affect

<sup>5</sup> May Affect, Likely to Adversely Affect

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## 1 INTRODUCTION

The Blue Mountain Ranger District of the Malheur National Forest (MNF) proposes to authorize livestock grazing for the next five seasons, 2012-16, on the York and Bear Allotments. Consistent with the Endangered Species Act (ESA) and its implementing regulations, this Biological Assessment (BA) documents the analysis and conclusions of the Forest Service regarding the effects of implementing the livestock it intends to authorize during that period. The analysis in the BA evaluates the effects on: (1) the Middle Columbia River Steelhead Distinct Population Segment (DPS) listed by the National Marine Fisheries Service (NMFS) as Threatened, and the Columbia River bull trout DPS listed by the US Fish and Wildlife Service (FWS) as Threatened; (2) designated critical habitat (CH) for both of these DPSs; and (3) Essential Fish Habitat (EFH) established for Chinook salmon, a species regulated under a Federal fisheries management plan. It is prepared in compliance with the requirements of Forest Service Manual (FSM) 2630.3, FSM 2672.4, ESA regulations, and regulations promulgated pursuant to the Magnuson-Stevens Fishery Conservation Act (MSA) as amended by the Sustainable Fisheries Act of 1996 (Public Law 104-267) (MSA §305(b)(2)).

### 1.1 CONSULTATION HISTORY

Past and ongoing informal and formal consultations that overlap the ESA action area and the 6<sup>th</sup> field HUC subwatersheds of York and Bear Allotments are described in this section.

#### 1.1.1 INFORMAL CONSULTATIONS (RECENT AND ONGOING)

##### 1.1.1.1 LIVESTOCK GRAZING

Consultation with NMFS was concluded on the 2006 grazing season for all MNF allotments. A letter of concurrence (LOC) was completed for the 2006 grazing season for the York and Bear allotments since the effects were determined to not likely to adversely affect (NLAA) MCR steelhead or designated CH.

##### 1.1.1.2 BLUE MOUNTAIN EXPEDITED SECTION 7 CONSULTATION PROCESS

The MNF received a concurrence letter in June 2007 (NMFS 2007a) from NMFS (2007/02970) for a consultation on the effects of the Blue Mountain Expedited Section 7 Consultation Process (Process). The LOC is currently active, and applies through June 2012. The Process evaluates consistency of an action with a set of project design criteria (PDC). Among the categories of actions specifically considered for application of the Process are: (1) vegetation management (mechanized and non-mechanized); (2) livestock grazing (range improvements); (3) wildlife, fish or watershed improvement projects; (4) road maintenance; (5) low impact permits; and, (6) recreation and administrative sites. Other types of projects can be covered under the programmatic as long as all of the PDC are met. The Level I team will determine if the use of the expedited process is appropriate for that project.

The action agency prepares documentation evaluating whether or not the action is consistent with the PDC covered by the concurrence letter. If an action is determined to be consistent with all of the PDC after evaluation by the Level 1 team, then an ESA effect determination of “May Affect, NLAA” applies. The Level 1 team then sends a letter to the respective action agency

official, documenting its finding regarding consistency with the Process and stating that the letter tiers to the concurrence letter.

Actions occurring in the allotments that are covered by the Process:

### **MNF Road Maintenance**

The MNF has consulted with the NMFS and USFWS on Forest wide road maintenance. On January 29, 2010 a letter was sent to the Malheur National Forest Supervisor from the interagency members of the Malheur Level 1 Team (FS, BLM, NMFS and FWS). The letter tiered to the 2007 NMFS concurrence letter for the Process and stated that the team had reviewed the PDC documentation package for the MNF Road Maintenance program for consistency with the Process. The team agreed with the MNF finding that the project “may affect, but is not likely to adversely affect (NLAA) the species and their designated CH for Mid-Columbia River Steelhead and Bull Trout based on the rationale that was presented for consistency with all PDC in the documentation package.” The Malheur Level 1 Team also concluded that the documentation package demonstrated that the action would adequately avoid, minimize or otherwise offset potential effects to designated EFH and fulfilled requirements under the MSA. The letter provided ESA and MSA coverage for the Forest’s Road Maintenance program from 2010 to 2015.

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## **1.1.2 FORMAL CONSULTATION (RECENT AND ONGOING)**

### **1.1.2.1 AQUATIC AND RIPARIAN RESTORATION PROGRAMMATIC CONSULTATION**

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The Forest Service and BLM concluded formal consultation on June 27, 2008 with NMFS (2008/03505) on Forest Service and BLM aquatic restoration activities for administrative units in Oregon and Washington including the MNF. The biological opinion (BO) applies through CY 2012, and provides coverage for 19 aquatic restoration program activity types:

1. Large Wood, Boulder, and Gravel Placement
2. Reconnection of Existing Side Channels and Alcoves
3. Head-cut Stabilization and Associated Fish Passage
4. Bank Restoration
5. Fish Passage Culvert and Bridge Projects
6. Irrigation Screen Installation and Replacement
7. In-channel Nutrient Enhancement
8. Floodplain Overburden Removal
9. Reduction of Recreation Impacts
10. Estuary Restoration
11. Riparian Vegetation Treatment (non-commercial, mechanical)
12. Riparian and Upland Juniper Treatment (non-commercial)
13. Riparian Vegetation Treatment (controlled burning)
14. Riparian Area Invasive Plant Treatment
15. Riparian Exclusion Fencing (with water gaps and stream crossings)
16. Riparian Vegetation Plantings
17. Road Treatments
18. Removal of Legacy Structures

## 19. Fisheries, Hydrology, Geomorphology Wildlife, Botany, and Cultural Surveys in Support of Aquatic Restoration

Actions occurring in the ESA action area that are covered by the Aquatic and Riparian Restoration Programmatic Consultation:

### York On/Off Allotment:

Plantation Maintenance – Long Creek Thinning Project. This project includes pre-commercial thinning and controlled burning of plantations from previously harvested units, some of which occur within the York On/Off Allotment. The project is expected to begin in 2011 and includes design criteria specified in aquatic restoration program activity types #11 and #13.

### Bear Allotment:

Plantation Maintenance – Lower Middle Fork Project. This project includes pre-commercial thinning and controlled burning of plantations from previously harvested units, some of which occur within the Bear Allotment. The project began in 2010 and includes design criteria specified in aquatic restoration program activity types #11 and #13.

## 2 DESCRIPTION OF PROJECT AREA

### 2.1 BEAR ALLOTMENT

Grazing activities on the Bear Allotment are conducted within the Middle Fork John Day (HUC 4# 17070203) subbasin. This allotment lies within the Big Creek-Middle Fork John Day (HUC 5# 1707020303) watershed. Elevations within the Allotment range from 3,400 feet along the Middle Fork John Day River to 4,500 feet.

Overstory vegetation in the Allotment consists of Ponderosa Pine, Douglas Fir, Western Larch, and Lodgepole Pine. Dominant grass species throughout the allotment are bluebunch wheatgrass, Idaho fescue, various wheatgrass species with lesser components of elk sedge, and pine grass. Riparian overstory vegetation generally consists of a mix of hardwood and conifer species along the stream. The majority of Mosquito Creek is dominated by hawthorn.

TABLE 3. MILES OF STEELHEAD CRITICAL HABITAT – BEAR ALLOTMENT

Stream Name	Miles of Steelhead Critical Habitat	Miles of MSRA
Middle Fork John Day River	1.38	0.96
Mosquito Creek	1.60	
<b>TOTAL</b>	<b>2.98</b>	<b>0.96</b>

TABLE 4. MILES OF BULL TROUT CRITICAL HABITAT – BEAR ALLOTMENT

Stream Name	Miles of Bull Trout Critical Habitat
Middle Fork John Day River	1.38
<b>TOTAL</b>	<b>1.38</b>

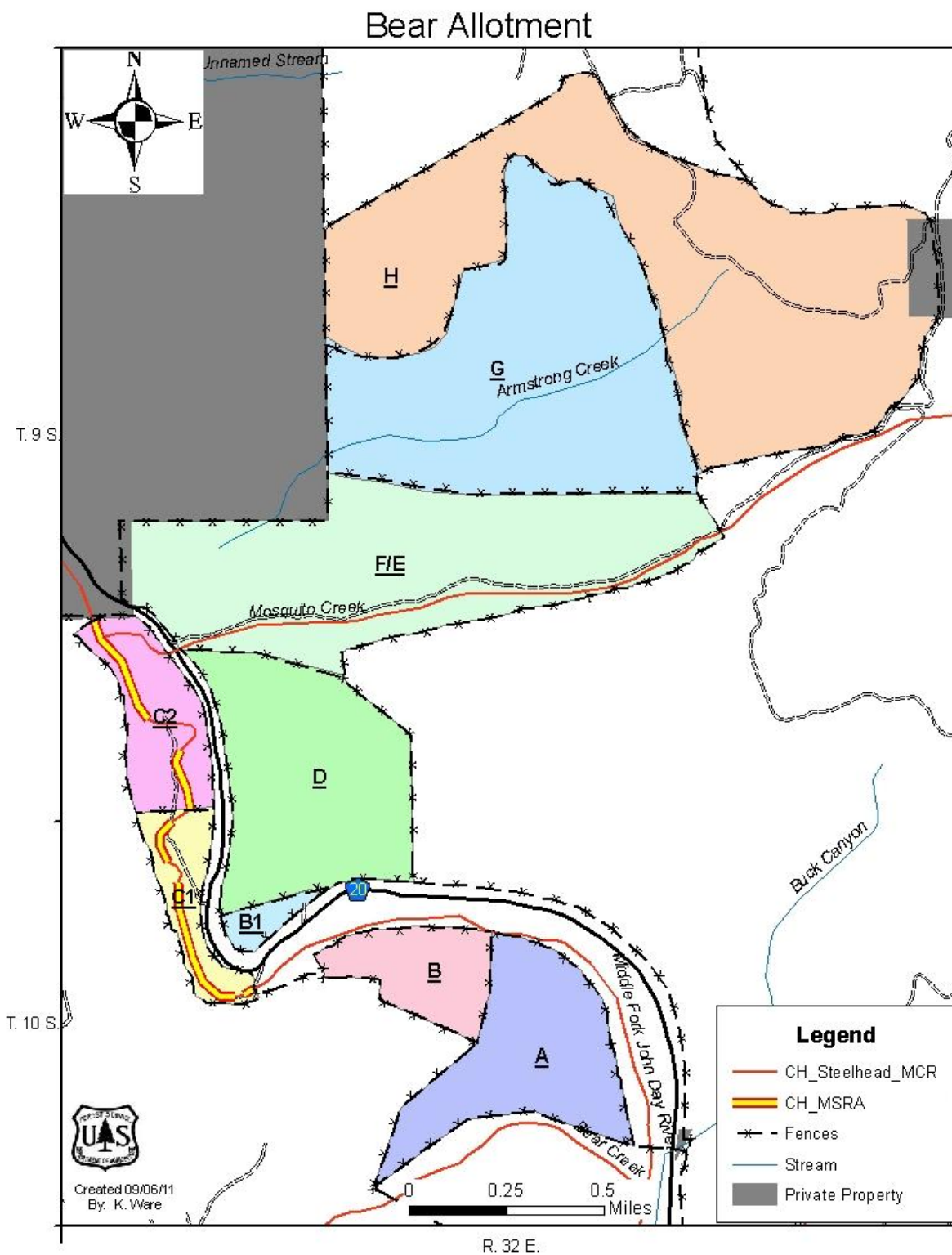


FIGURE 1. BEAR ALLOTMENT AND PASTURE MAP WITH STEELHEAD DCR AND MSRA

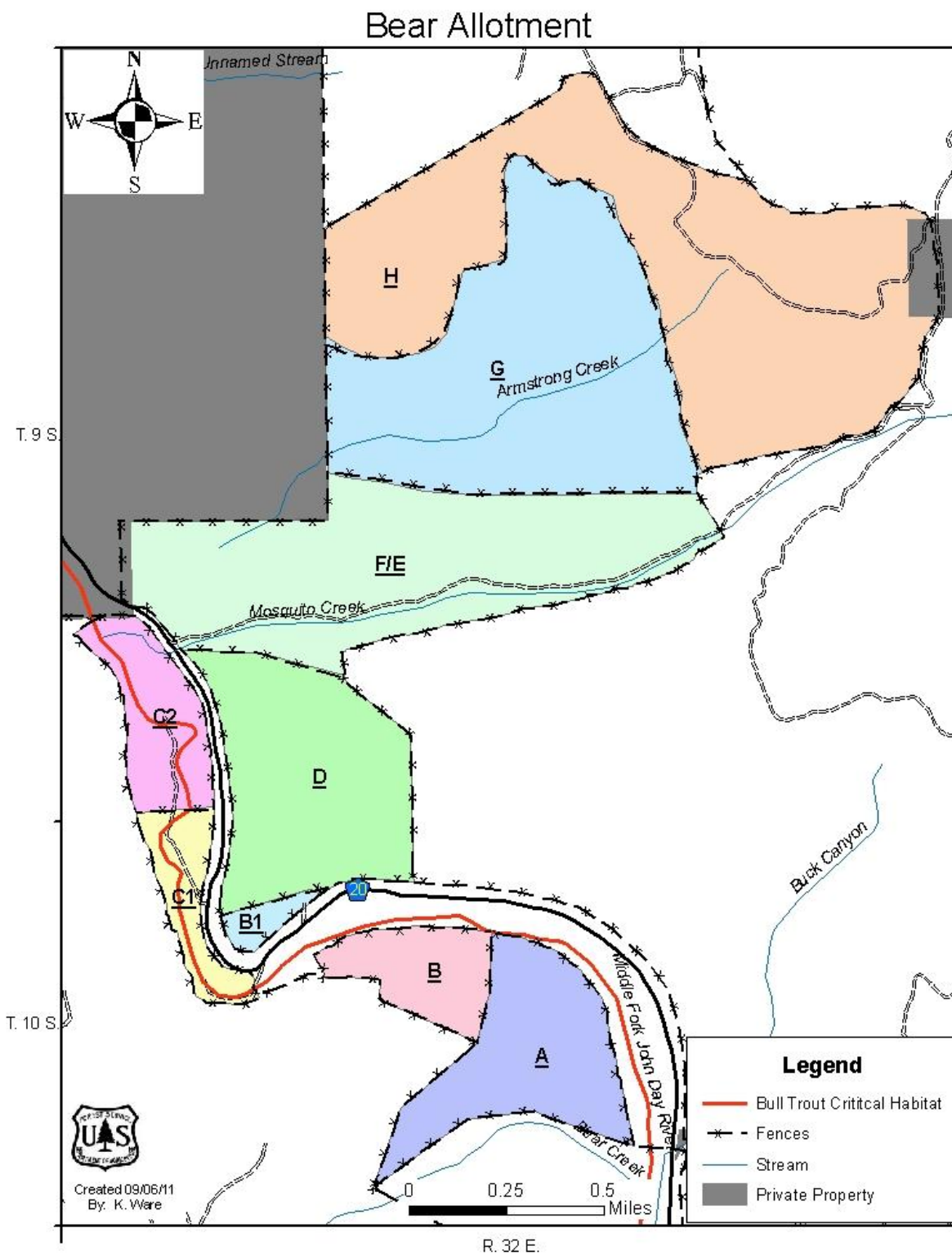


FIGURE 2. BEAR ALLOTMENT AND PASTURE MAP WITH BULL TROUT DCR



This allotment consists of 8 pastures (Figures 1 and 2) and contains 2.83 miles of steelhead critical habitat (Table 3) and 1.38 miles of bull trout critical habitat (Table 4). Within streams containing steelhead critical habitat, the Forest Service has identified areas that are the most accessible and sensitive to livestock impacts. These areas are typically production flats and generally contain valuable spawning grounds. They are located in low gradient, open meadow reaches of a stream. Because of the life-cycle stages of MCR steelhead relevant to the streams within these allotments, the Forest Service decided to focus the identification of the “Most Sensitive Riparian Areas” (MSRA) on the known and likely spawning areas for MCR steelhead. Riparian areas adjacent to spawning areas can be considered relatively more sensitive for the species because they occur on low gradient sections of a stream and often prove to be particularly attractive to grazing livestock as a water and shade source for much the same reasons. MSRAs are not land-use designations and do not have any legally independent force or effect, but instead are being referenced in this Proposed Action to facilitate the Forest Service’s analysis of impacts and provide a useful basis for distinguishing among areas within the York, and Bear Allotments so as to focus special attention on those areas the agency has already determined may be the most susceptible to causing adverse impacts to the listed fish from grazing. The process for determining MSRA can be found in Appendix G. The application of the criteria in respect to the Bear Allotment is displayed in Table 3 and Figure 1.

#### *C1 and C2 Pastures*

The C1 and C2 pastures contain the Middle Fork John Day River and short section of Mosquito Creek, both of which support steelhead critical habitat. The Middle Fork John Day River also supports bull trout critical habitat, used primarily as a migration corridor and for winter rearing.

The C1 and C2 pastures were historically used for hay production and consist of a mixture of meadow grass species, including Kentucky bluegrass and meadow foxtail. To maximize hay production the river was channelized and moved to the base of the hill on the western edge of the pastures. In 2009, restoration of the channel relocated the river to its historic location through the center of meadow. The restoration work primarily took place within the C2 pasture. The river remains channelized and along the western edge of the C1 pasture.

A portion of Mosquito Creek, after it crosses under County Road 20, runs through the C2. Mosquito Creek does not have a defined channel between County Road 20 and the MFJD River. The creek spreads out across a flat 200 foot floodplain and filters down to the MFJDR via overland sheet flow in the spring and primarily as sub-surface flow during the later summer months. ODFW steelhead spawning surveys have not located any steelhead redds or identified adult steelhead in Mosquito Creek. They attribute this to the lack of a defined channel in the C2 pasture.

#### *E/F Pasture*

The E/F pasture contains 2.45 miles of Mosquito Creek. Elevations within this pasture range from 3,400 feet to 4,000 feet. This pasture is predominately south facing with opens stands of ponderosa pine in the overstory and Idaho fescue/bluebunch wheatgrass communities in the understory. Mosquito Creek runs east to west along the southern edge of the pasture. It is dominated by dense Hawthorne and alder limiting cattle access to existing crossings in the lower reaches while becoming more open and accessible in the upstream reaches.

### *A, B, B1, D Pastures*

The A, B, and B1 pastures are entirely upland pastures with no live water or streams. The D pasture is also an upland pasture however it does have a water gap containing approximately 20 feet of Mosquito Creek.

### *G and H Pastures*

The G and H pastures do not contain steelhead critical habitat. Armstrong Creek does originate in the H pasture and during spring run-off flows through the G pasture, however it does not support steelhead or bull trout critical habitat.

## 2.2 YORK ON/OFF ALLOTMENT

Grazing activities on the York On/Off Allotment are conducted within the Middle Fork John Day River (HUC 4# 17070203) subbasin. This allotment lies within the Big Creek-Middle Fork John Day River (HUC 5# 1707020303) watershed.

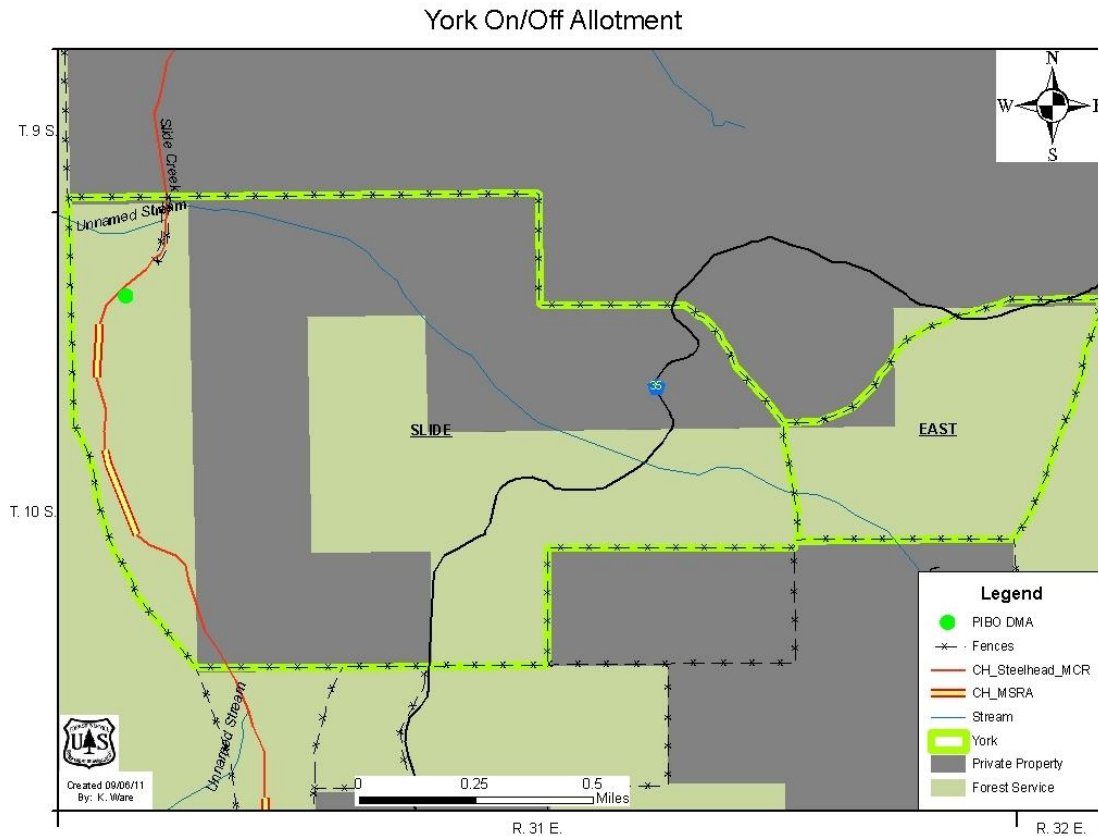
Elevations within the allotment range from approximately 3800 feet to 4600 feet. Vegetative types range from ponderosa pine to mixed conifer. Dominant grass species are bluebunch wheatgrass/Idaho fescue on the open hill slopes and elk sedge/pine grass in the forested areas.

There are no bull trout or designated critical habitat for bull trout in the allotment.

TABLE 5. MILES OF STEELHEAD CRITICAL HABITAT – YORK ON/OFF ALLOTMENT

Stream Name	Miles of Steelhead Critical Habitat	Miles of MSRA
Slide Creek	0.88	0.30
<b>TOTAL</b>	<b>0.88</b>	<b>0.30</b>





**FIGURE 3. YORK ON/OFF ALLOTMENT AND PASTURE MAP WITH STEELHEAD DCH AND MSRA**

This allotment consists of 2 pastures (Figure 3) and contains 0.88 miles of steelhead critical habitat (Table 5). Within streams containing steelhead critical habitat, the Forest Service has identified areas that are the most accessible and sensitive to livestock impacts. These areas are typically production flats and generally contain valuable spawning grounds. They are located in low gradient, open meadow reaches of a stream. Because of the life-cycle stages of MCR steelhead relevant to the streams within these allotments, the Forest Service decided to focus the identification of the “Most Sensitive Riparian Areas” (MSRA) on the known and likely spawning areas for MCR steelhead. Riparian areas adjacent to spawning areas can be considered relatively more sensitive for the species because they occur on low gradient sections of a stream and often prove to be particularly attractive to grazing livestock as a water and shade source for much the same reasons. The process for determining MSRA can be found in Appendix G. The application of the criteria in respect to the York On/Off Allotment is displayed in Table 5 and Figure 3.

#### *East Pasture*

The East pasture is approximately 150 acres and is almost entirely Forest Service land with predominately ponderosa pine and Douglas fir uplands, aspen stands and highly productive meadows composed of native bunchgrasses and forbs. Past timber harvesting activities have maintained enough openings between the forested areas to support a healthy herbaceous understory.

### *Slide Pasture*

The Slide pasture contains approximately 780 acres, of which 400 acres are private land and 380 acres of National Forest System Lands. Slide Creek runs through the Forest portion of this pasture for 0.88 miles.

## 3 FOREST DIRECTION AND POLICIES GUIDING DEVELOPMENT OF PROPOSED ACTION

Forest direction and policies provide a management framework that direct or guide development of grazing actions on the MNF. Components of the management framework include the MNF Land and Resource Management Plan (LRMP), pertinent LRMP amendments and Forest policies. The most pertinent amendments to the MNF LRMP are PACFISH and Amendment 29. The MNF Riparian Monitoring Strategy is a forest policy, MNF (2006).

### 3.1 MALHEUR NATIONAL FOREST LRMP

The MNF LRMP (MNF 1990) contains goals and objectives for the Range Program that provide direction with respect to range management and other resources. Goals 20 - 22 of the MNF LRMP for the Range program state:

*“Provide a sustained production of palatable forage for grazing by livestock and dependent wildlife species.”*

*“Manage rangelands to meet the needs of other resources and uses at a level which is responsive to site-specific objectives.”*

*“Permit livestock use on suitable range when the permittee manage livestock using prescribed practices.”*

A Range program MNF LRMP Objective also provides context:

*“Analyze allotments to determine proper stocking levels. Use specific management area goals and standards to resolve conflicts between wild horses, cattle, and big game.”*

Rangeland will be managed to meet the needs of ESA-listed species (MCR Steelhead, CR bull trout) and big game as “other resources.” The MNF Range program LRMP Objective directs that when there are conflicts between wild horse, cattle and big game uses in determining stocking levels, management area goals and standards will be used to resolve the conflicts. The LRMP direction described above provides conservation benefits to ESA-listed species (MCR Steelhead, CR bull trout) and its designated CH by directing that the needs of other resources will be met. In practice, the BMRD doesn’t conduct pre-season monitoring in these allotments, but monitoring livestock grazing use during the season accounts for effects from all resources.

Other components of the Forest management framework (MNF LRMP) that guide the development of the proposed action are discussed under the Forest amendments sections of the BA. The most pertinent amendments to the MNF LRMP are PACFISH and Amendment 29.

The MNF Land and Resource Management Plan (MNF 1990) was amended in 1994 (Amendment 29) in response to the Columbia River Basin Anadromous Fish Habitat Management Policy and Implementation Guide (USDA FS 1991). The Forest modified Standard 5 of the Fish and Wildlife resource elements. The amended Standard 5 included specific

numerical desired future conditions (DFCs) for Management Area 3A (non-anadromous riparian areas) and Management Area 3B (anadromous riparian areas). The DFCs addressed: 1) sediment/substrate, 2) water quality, 3) stream channel morphology, and 4) riparian vegetation. See Amendment 29 for the specific numeric values (Appendix B). The numerical DFCs were selected to protect water quality, features of riparian vegetation, and components of fish habitat.

Amendment 29 did not set specific quantifiable standards for livestock grazing activities. However, grazing activities can directly affect the attainment of Amendment 29 DFCs for: 1) sediment/substrate (cobble embeddedness), 2) water quality (water temperature – Forest wide or by fish species), 3) channel morphology (large woody debris, bank stability, lower bank angle, width to depth ratios, 4) riparian vegetation (ground cover, percentage of stream bank vegetated), and 5) shade/canopy closure (hardwood/meadow complex). DFCs were developed to provide the criteria against which attainment or progress toward attainment of the riparian goals are measured. The MNF manages according to the more stringent standards applicable to habitat components of anadromous riparian areas as between Amendment 29 DFCs and the Riparian Management Objectives (RMOs) of the PACFISH amendment, although it should be added that the two are not always directly correlative or equally applicable; for example, with respect to bank stability, the Amendment 29 DFC applies to forested areas only. Nevertheless, this overarching directive provides conservation benefits to ESA-listed species (MCR Steelhead, CR bull trout) and its designated CH. Table 6 presents Amendment 29 DFCs and PACFISH RMOs by habitat indicator/criterion and displays which of the two is more stringent to the extent that both may be applicable in a given management situation.

The numeric values were developed for the Resource/Habitat Elements (features) of the MNF LRMP management areas 3A and 3B in amendment 29 (Appendix B). Amendment 29 states, *“These values are based upon the best information currently available and are considered to be consistent with management area desired future condition. If new information becomes available in the future which indicates changes in the numeric values to achieve the stated desired condition, these values may be inserted as a clarification/correction to the individual standard.”* Since the Forest Service adopted the Inland Native Fish Strategy (INFS) in 1995, it has been considered to contain better numeric values for bull trout water temperatures to achieve the stated desired conditions of amendment 29 (USDA FS 1995).

In general, the MNF applies the INFS RMO for water temperature to bull trout rather than standards from PACFISH or Amendment 29. INFS established a water temperature RMO that used the best available published and non-published scientific literature to define favorable water temperatures for inland native fish. The PACFISH RMO for water temperature was developed to meet the habitat needs of anadromous fish such as steelhead and Chinook salmon rather than bull trout. The INFS RMO for water temperature identified maximum water temperatures below 59°F within adult holding habitat and below 48°F within spawning and rearing habitats. The INFS RMO is more conservative for bull trout than the water temperature standards of either Amendment 29 or PACFISH (Table 6). The MNF considers the INFS water temperature standard to be the best available, favorable water temperatures for inland native fish such as bull trout.

TABLE 6. IDENTIFICATION OF MORE STRINGENT HABITAT INDICATOR NUMERIC VALUES OR CRITERIA BETWEEN AMENDMENT 29 DESIRED FUTURE CONDITIONS OR PACFISH RIPARIAN MANAGEMENT OBJECTIVES.

Habitat Indicator	Desired Future Condition or Riparian Management Objective		More Stringent Condition or Objective
	Amendment 29	PACFISH RMO	
Cobble embeddedness	<20%	NA	Amendment 29
Water temperature	Forestwide: No increase if < 68°F, reduce to 68°F if >68°F ≤ 55°F Bull Trout spawning and rearing habitat	No measurable increase. Max below 64°F for migration/rearing, max below 60°F for spawning	MCR steelhead: PACFISH RMO  CR bull trout: Amendment 29 but MNF uses INFS RMO. <sup>4</sup>
Large Woody Debris Stream Densities	Varies by ponderosa (20-70/mi),  Mixed conifer (80- 120/mi), lodgepole (100-350/mi). Sizes vary.	>20/mi >12" dia >35' length	Amendment 29 is more specific
Pool frequency	Range expected for Rosgen B&C streams, upper limits adjusted for streams >75 ft. to be consistent w/PACFISH. Provides table w/ranges by bankfull width	Table provided shows pools/mile by wetted width. All values fall within ranges by BFW of Amendment 29	Same
Bank stability (forested)	90% and no decrease if above 90%	>80%	Amendment 29
Lower bank angle (non-forested)	50-75% of banks w/90 degree angle or greater	>75% w/90 degree angle	PACFISH RMO
W/D ratio	<10	<10	Same
Potential LWD forest	To provide a rate of input to maintain LWD	NA	Amendment 29

<sup>4</sup> Bull trout have the coldest water temperature requirements of any native salmonid in the Pacific Northwest. The MNF considers INFS to contain better numeric values for bull trout water temperature to achieve the stated desire conditions of amendment 29.

Habitat Indicator	Desired Future Condition or Riparian Management Objective		More Stringent Condition or Objective
	Amendment 29	PACFISH RMO	
	standard		
Ground cover	90% of site potential	NA	Amendment 29
% streambank vegetated	90% of site potential	NA	Amendment 29
Shade/canopy closure	Varies by conifer species forest. Hardwood/meadow complex 80% shaded	NA	Amendment 29

### 3.1.1 PACFISH LRMP AMENDMENT

PACFISH applies specifically to the MNF lands within the range of anadromy including the Bear and York On/Off Allotments. PACFISH amended Forest Service Land and Resource Management Plans (LRMPs). PACFISH contains the following components that provide the necessary direction and objectives, and regulatory certainty that FS management actions will be designed to maintain and restore ecological processes that support high quality habitat for salmon and steelhead over the long term:

- Riparian Goals;
- Riparian Management Objectives (RMOs);
- Delineation of streamside areas (Riparian Habitat Conservation Areas) that are important to maintenance of high quality aquatic habitat and where special management considerations are applied;
- Standards and/or guidelines to ensure projects do not prevent or retard attainment of riparian goals and management objectives;
- Designation of Key watersheds where additional management emphasis and/or watershed analysis is required to ensure that salmon and steelhead habitat is maintained or provided priority for restoration;
- Watershed analyses to provide sufficient context for designing actions that support maintenance or restoration of aquatic habitats needed for recovery of ESA-listed salmon and steelhead;
- Targeted watershed restoration identified through watershed analysis;
- Monitoring program to evaluate the implementation (compliance) and effectiveness of PACFISH in improving aquatic habitat on federal lands.

Riparian Goals provide management context for proposed activities. The goals of PACFISH establish an expectation of the characteristics of healthy, functioning watershed, riparian areas, and associated fish habitats. They are stated in relatively broad, generic terms such that they can be said to apply to most riparian areas regardless of stream type and other more site-specific conditions, but need to be evaluated in the context of the particular stream at issue. Since the

quality of water and fish habitat in aquatic systems is inseparably related to the integrity of upland and riparian areas within watersheds, PACFISH articulates the following goals to maintain or restore:

- Water quality, to a degree that provides for a stable and productive riparian and aquatic ecosystem;
- Stream channel integrity, channel processes and sediment regime (including the elements of timing, volume, and character of sediment input and transport) under which riparian and aquatic ecosystems developed;
- Instream flows to support healthy riparian and aquatic habitats, stable and functioning channels, and the ability to route flood flows;
- Natural timing and variability of water tables in meadows and wetlands;
- Diversity and productivity of native and desirable non-native plant communities in riparian zones;
- Riparian vegetation to provide for 1) an amount and distribution of large woody debris characteristic of natural aquatic and riparian ecosystems, 2) adequate summer and winter thermal regulation within the riparian and aquatic zone, and 3) rates of surface erosion, bank erosion, and channel migration characteristics of those under which the communities developed;
- Riparian and aquatic habitats necessary to foster unique genetic fish stock that evolved within the specific geo-climatic region; and,
- Habitat to support populations of well-distributed native and non-native plant, vertebrate and invertebrate populations that contribute to the viability of riparian-dependent communities.

#### 3.1.1.1 PACFISH RIPARIAN MANAGEMENT OBJECTIVES

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Interim quantitative Riparian Management Objectives (RMOs) for stream channel, riparian and watershed conditions were also developed as a part of PACFISH to provide criteria against which attainment or progress of the strategy's riparian goals may be measured. The objectives need to be evaluated and assessed temporally to reflect the ecological capabilities of specific ecosystems and the fact that attainment of or progress toward many of the objectives is only able to occur over extended periods of time. In general, and to the extent applicable and feasible, the MNF manages livestock grazing so as not to prevent or retard attainment of these RMOs unless Forest Plan amendment # 29 is more stringent, which will benefit habitat for MCR Steelhead.

**Bank Stability:** at least 80%

**Water Temperature:** No measureable increase in maximum temperature; Meet state water quality standards. The standard is defined as: All streams identified as having anadromous fish passage and salmonid rearing use for Designated Beneficial Use purposes. 7 Day Mean Max 64°F (17.8°C) (migration and rearing habitat); 7 Day Mean Max 60°F (15.6°C) (spawning habitat).

**Width-to-Depth Ratio (W:D):** W:D <10, mean wetted width divided by mean depth (NMFS PACFISH BO 1998); or **Bankfull Width-to-Depth Ratio** within 75<sup>th</sup> percentile of the range for minimally managed or reference watershed conditions (i.e. healthy streams) by stream type (analysis pending from PACFISH/INFISH biological opinions (PIBO) Effectiveness Monitoring Team).



### 3.1.1.2 PACFISH RIPARIAN HABITAT CONSERVATION AREAS AND STANDARDS

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Project- and site-specific standards apply to all Riparian Habitat Conservation Areas (RHCAs) and to projects and activities in areas outside RHCAs that would degrade them. Standards and/or guidelines were developed to ensure to the extent practicable given site conditions that projects do not prevent or retard attainment of or reasonable progress toward riparian goals and management objectives. PACFISH (USDA FS and USDI BLM 1995) standards for livestock management are presented below.

- GM-1 - Modify grazing practices (e.g., accessibility of riparian area to livestock, length of grazing season, stocking levels, timing of grazing, etc.) that retard or prevent attainment of Riparian Management Objectives or are likely to adversely affect listed anadromous fish. Suspend grazing if adjusting practices is not effective in meeting Riparian Management Objectives and avoiding adverse effects on listed anadromous fish (PACFISH).
- GM-2 – Locate new livestock handling and/or management facilities outside of Riparian Habitat Conservation Areas. For existing livestock handling facilities inside the Riparian Habitat Conservation Areas, assure that facilities do not prevent attainment of Riparian Management Objectives or adversely affect listed anadromous fish. Relocate or close facilities where these objectives cannot be met.
- GM-3 – Limit livestock trailing, bedding, watering, salting, loading, and other handling efforts to those areas and times that will not retard or prevent attainment of Riparian Management Objectives or adversely affect listed anadromous fish.

Implementing these standards clearly provide a conservation benefit to MCR Steelhead and designated CH.

### 3.1.1.3 PACFISH MONITORING

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The PACFISH Monitoring Strategy was designed to feed information back to management for decision making. The implementation strategy uses “endpoints” and “triggers” to assess whether the authorized grazing is having the anticipated effects to resources and as an additional precautionary measure that helps the Forest Service to ensure that PACFISH direction is met as the season progresses. The endpoints measure annual conditions after grazing has been completed and the triggers measure conditions during grazing to determine if adjustments are necessary or appropriate to be made to the rotation and schedule the Forest Service has developed based on its best professional judgment and projections for a particular grazing season. The results of implementation monitoring are to be fed into effectiveness monitoring, which uses “trend” of long-term indicators of habitat condition to assess the need for revising management. This monitoring has a direct application to habitat features essential to long-term conservation of salmon and steelhead.

The PIBO Effectiveness Monitoring (EM) Program (under the Interagency Deputy Team) was initiated to evaluate the effects of land management activities on aquatic and riparian communities at multiple scales and to determine whether PACFISH management practices are effective in maintaining or improving the structure and function of riparian and aquatic conditions. A pilot study was begun in 1998 on Forest Service lands within the Salmon River basin of central Idaho. In 2000, the pilot study was expanded to include additional Federal

Lands in the interior Columbia River basin. The study area includes 20 National Forests and nine BLM field units within the interior Columbia River basin. Results from sample size analysis suggested that the monitoring program will be able to detect changes in resource condition at the scale of individual Forests and BLM field offices (35 to 90 sites) for many of the attributes measured. The PACFISH Effectiveness Monitoring Program sampling design anticipates collecting information at least through 2015.

The PIBO EM and Implementation Monitoring (IM) programs are coordinated such that data collected to assess trend is linked to management actions taken under the PACFISH strategy. Thus, monitoring sites selected for evaluating the effectiveness of PACFISH are also monitored for compliance with the standards and guidelines. Preliminary results from broad-scale aquatic habitat status and trend monitoring of FS and BLM lands within the interior Columbia River basin since 2001 indicates conditions have improved over the past 5 years, continuing the habitat recovery the agencies intended to commence upon their adoption in 1995 of the protections in PACFISH (NMFS 2009).

#### 3.1.1.4 PACFISH ENCLOSURE B: LIVESTOCK GRAZING GUIDELINES

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A revision of PACFISH Enclosure B, the “Recommended Livestock Grazing Guidelines,” was sent to the PACFISH Forest Supervisors on August 14, 1995 (Appendix B). The guidelines were recommended for use in modifying applicable allotment management plans, annual operating plans, project decision documents and instructions to permittees to provide a high degree of assurance that objectives for conservation and restoration of anadromous fish habitat would be met.

The revision identified a set of key assumptions. One of the assumptions is that the goals or desired outcomes of management efforts provide the foundation for the recommended programmatic livestock grazing guidelines. The PACFISH EA was described as providing suitable riparian goals. All management activities should be structured so as not to prevent or meaningfully hinder accomplishment of the goals.

A summary of key priorities identified in the Enclosure B revision are:

- Maintain or allow for improvement of conditions where criteria for late-seral ecological status are met or exceeded.
- Adjust management practices where the criteria for mid-seral ecological status are met but the trend is static or downward. This is especially important where vegetative factors are primarily responsible for the mid-seral rating.
- Adjust management practices where the criteria for early seral ecological status are met, with the understanding that deteriorated stream bank and channel conditions may not be recovered in the near term.

The Enclosure B revision stated that Al Winward, in Clary and Webster (1989) defined ecological status as a measure of the degree of similarity between current vegetation and potential vegetation for a given riparian area. Refined definitions for the three ecological classes were presented:

- Early seral. Percent similarity of riparian vegetation to the potential natural community/composition less than or equal to 25%; or, stream bank/channel condition rating “poor”.



- Mid-seral. Percent similarity of riparian vegetation to the potential natural community/composition 26-50% or better; and, stream bank/channel condition of at least “fair”.
- Late seral. Percent similarity of riparian vegetation to the potential natural community/composition greater than or equal to 50% or better; and, stream bank/channel condition rating “good” or better.

The MNF is utilizing Winward (2000) to evaluate ecological status of riparian vegetation, in place of the process described in Enclosure B. If similarity of riparian information is lacking, the Enclosure B revision suggested using PFC condition classes as a substitute.

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### 3.2 MALHEUR NATIONAL FOREST RIPARIAN MONITORING STRATEGY

Many accepted methodologies and analytical tools are available to monitor short-term and long-term rangeland and forest health. The methods and tools chosen are dependent on the specific monitoring objectives as well as constraints such as timing, available funding and personnel, other priorities, and the geographical area to be monitored. Described below are the overall monitoring strategy, methods and analytical tools that the Malheur National Forest is currently using for determining condition and trend of riparian ecosystems as they relate to grazing activities. The assessments and monitoring methods used are intended to be an important part of the adaptive management process and are subject to changes or modifications based on new scientific findings and improvements in methodologies as well as changes in definitions and policy. In particular, see Appendix D for a discussion of the monitoring protocol the Forest Service intends to use to evaluate compliance with bank alteration thresholds. Moreover, risk analyses and prioritization generally should be completed in all areas prior to initiating monitoring in order to determine the level and intensity of quantitative data collection. PFC assessments can serve as the risk analyses/prioritization step.

Below are the key components of the MNF Riparian Monitoring Strategy:

1. Information Gathering and Interpretation
  - Proper Functioning Condition (PFC) Assessment –qualitative condition assessment over a stream reach (geomorphic or unit-specific)
  - Multiple Indicator Monitoring (MIM) – quantitative monitoring protocol at designated DMAs
  - Analysis – interpretation and evaluation of assessment and monitoring information to determine current riparian condition and, to the extent feasible, trend
  - Channel cross-section, streambed particle size distribution, and reach description measurements (i.e. Rosgen Channel Type)
  - Forest Service Region 6 Level II Stream Inventory Surveys – extensive quantitative assessment of stream channel, riparian vegetation, aquatic habitat condition, and biota to determine condition of selected stream systems
  - Spawning Surveys – quantitative assessment of redd vulnerability to disturbance
2. Support determinations of plan compliance – Provide information on which MNF can assess compliance with Forest Plan, including PACFISH & INFISH amendments. See Appendix C. for further discussion of Forest Plan standards and objectives related to riparian areas, water quality and fish habitat.

- Standards are GM 1-4 in PACFISH & INFISH; standards 15-21 in Forest Plan (see Chapter IV).
- Management Objectives for stream and riparian areas are described in PACFISH & INFISH amendments (RMO's) and in Amendment 29 of Forest Plan for MA3A/B (DFC's).

### 3. Recommendations

- Shows linkage between condition, trend, and past/current management activities
  - A process that provides support for grazing management decisions or any necessary or appropriate adaptive management adjustments
    - Allows annual adjustment of management strategies, as needed, to achieve compliance with plan direction

### *Proper Functioning Condition Assessments*

Proper functioning condition (PFC) assessments are a qualitative method for determining the condition of riparian areas. The term PFC is used to describe both the assessment process, and a defined, on-the-ground condition of a riparian area. PFC assessments can be an appropriate starting point for determining and prioritizing the type and location of quantitative inventory or monitoring necessities, and has been proven to be an excellent communication tool for bringing a wide diversity of publics to agreement. All PFC assessments are to be conducted with a journey level interdisciplinary team. One purpose of these assessments is to help correlate the findings with the trend towards attainment of the Malheur Forest Plan Riparian Management Objectives (RMOs), more specifically, to determine whether grazing practices are retarding attainment of Near Natural Rates of Recovery of RMOs.

### *Multiple Indicator Monitoring*

The July 1, 2003 PACFISH/INFISH Implementation Monitoring Program Manual provides the background and direction for monitoring. The Multiple Indicator Monitoring (MIM) supplement by Cowley/Burton, dated May 2005 with addendums, provides the procedures in use by the MNF to monitor stream banks and riparian vegetation. The Interagency Implementation Team created the above documents; see Appendix E for these documents. The authors of MIM recently issued a 2011 technical guide as well (Burton et.al. 2011). MIM for grazing activities is designed to determine whether or not livestock grazing management is resulting in "Near Natural Rates of Recovery" as defined by PACFISH/INFISH. Below are the three components, which comprise MIM. Monitoring is to be conducted by an interdisciplinary professional team trained in riparian plant identification and channel classification. Multiple indicator monitoring consists of implementation (endpoint indicator) monitoring and effectiveness (riparian objective) monitoring at designated locations (i.e. designated monitoring areas).

### *Designated Monitoring Areas*

Designated Monitoring Areas (DMA's) are the locations in riparian areas and along streambanks where quantitative monitoring takes place. They are monitored to provide information concerning the management of critical areas. Essentially DMA selection relies on the theory that if proper management occurs in that location, proper management will be occurring throughout the rest of the management unit. See Appendix E for the procedures used to collaboratively establish DMA's. The goal is to establish more DMA's each grazing season in order to establish a 5-year re-monitoring schedule and have coverage across the Forest's allotments.

### *Implementation Monitoring - Endpoint Indicators*

Implementation (endpoint indicator) monitoring measures indicators to determine if the authorized livestock grazing strategy for a particular season has had the projected effects to resources that the MNF has anticipated in developing the strategy and to determine if any adaptive management adjustments need to be made for the following season(s). It provides information to assist with making decisions under adaptive management. Presently, implementation monitoring includes: modified extensive browse utilization (Interagency Technical References, 1996), modified stubble height (Interagency Technical Reference, 1996 and Challis Resource Area, 1999), and streambank alteration (Cowley, 2004). These procedures provide information to refine and make annual adjustments to livestock grazing management practices necessary to meet long-term management objectives (adaptive management). They can be used as early warning indicators that current grazing impacts may prevent the achievement of management objectives and can also be used to help explain changes in riparian vegetation and channel conditions over time. See web-site (<http://www.rmsmim.com/>) for sampling procedures used.

### *Effectiveness Monitoring - Riparian Objectives*

Effectiveness (riparian objective) monitoring is designed to address the question of whether or not management practices currently applied to the area are achieving the desired results. These procedures are designed to assess the current condition and measure changes in streambanks, channels, and streamside vegetation over time, i.e., trend. They help determine if local livestock grazing management strategies and other land management actions are making progress toward achieving the long-term goals and objectives for streamside riparian vegetation and aquatic resources. The goal is to conduct effectiveness monitoring every three to five years on riparian areas and streambanks. This period of time is considered to be the minimum necessary to detect changes, although unusually wet years and/or flood events may result in short-term changes that validate the need to monitor more frequently, or at least at the time of the event. Budget and personnel constraints may limit the extent in which monitoring of this type will be conducted. Presently, effectiveness monitoring includes: modified greenline composition (adapted from Winward 2000 and BLM 1996), woody species height class (Kershner et al. 2004), streambank stability and cover (adapted from Kershner et al. 2004), woody species age class (adapted from Winward 2000), greenline-to-greenline width (Burton et al. 2008), substrate (Bunte and Abt 2001), and residual pool depth and pool frequency (Lisle 1987). These provide data and information concerning the present conditions and trend of riparian vegetation, channels, and streambanks, and to help determine if aquatic systems are being degraded, maintained, or restored across the Malheur National Forest. See web-site (<http://www.fs.fed.us/biology/fishecology/emp/>) for PIBO data and sampling procedures used.

### *Forest Service Region 6 Level II Stream Inventory Surveys*

Forest Service Region 6 Level II Stream Inventory Surveys generate comparable baseline information on conditions of fish-bearing streams to support a variety of management activities. As inventories are completed and repeated over time, the information generated by them can be useful in measuring changes in stream channel conditions and determining attainment of habitat management objectives. The Level II inventory generates quantitative measurements and estimates of channel conditions and habitat attributes, including core attributes of streamflow, temperature, substrate composition, width/depth ratio, channel length and sinuosity, gradient,

pool frequency, large wood, bank stability, and special habitats. Numerous non-core optional attributes may also be evaluated based on Forest needs, such as stream shading and overstory/understory vegetation. The Forest goal is to inventory 10 percent of fish-bearing streams per year, inferring a 10-year re-inventory recurrence interval. The 2010 Region 6 Stream Inventory Handbook can be found at:

(<http://www.fs.fed.us/r6/water/fhr/sida/handbook/Stream-Inv-2010.pdf>)

### *Spawning Surveys*

Spawning surveys are a quantitative method to assess steelhead redd presence and vulnerability to livestock disturbance and may also be used to assess compliance with the level of “take” authorized within a Biological Opinion. The Forest has developed a strategy to avoid redd trampling “take” of steelhead and bull trout (see Appendix F).

### *Uplands Monitoring*

Beginning in the 1930s, permanent camera points were established on the Malheur National Forest. Their purpose was to monitor the effects of management on the resources of the Forest (Fifty Years of Change on the Range, R6-Mal-035-89). Many camera points have been re-photographed a number of times. This monitoring will continue.

In the 1950s and early 1960s Parker Three-Step C&T (Condition & Trend) Transects were installed throughout the Forest. The majority of these were established in the uplands. Over the last five years some of these transects have been re-examined. The procedure has been to read the transect using the original Three-Step method and then reread the transect using a modified Daubenmire cover/frequency method (see Technical Reference 1734-4). This allows comparisons between old and new information to determine ecological condition and trend and establishes a baseline using the more accurate cover/frequency method for gathering future data. The re-examining of these established transects will continue. If new trend transects are established the modified Daubenmire cover/frequency will be used.

There are a variety of additional or other monitoring methods available for use. The method or methods to be used will depend on the questions needing to be answered and considering other priorities. In some cases ocular observation (qualitative) will be sufficient to measure utilization, but when specific concerns are identified the forest may need quantitative methods such as Paired Clipped plots or development and use of height/weight curves may be necessary. Some of the more commonly used methods can be found in “*Utilization Studies and Residual Measurements*” (BLM 1996). (Please refer to ***Malheur National Forest Range Monitoring Guidelines***, October 16, 2006 for additional accepted methodology)

All of the monitoring methods used by the Forest are also intended to facilitate communications between forest range and resource personnel, grazing permittees and consulting agency personnel. This will largely be accomplished through participation and one-on-one interaction during the interdisciplinary, on the ground implementation.

## 4 PROPOSED ACTION AND ESA ACTION AREA

The MNF used the LRMP direction and policies presented in Section 3 to design the 2012-2016 proposed grazing action for Bear and York On/Off Allotments. Public laws such as Clean Water Act were considered. The development of the project design criteria, grazing end-points and grazing strategies for the allotments considered the PACFISH RMOs and Grazing Management

Standards, desired conditions and standards from amendment 29 to the MNF LRMP, PACFISH livestock grazing guidelines (Enclosure B) and MSRAs (Appendix G). Forest policies on Riparian Monitoring and MSRAs were also informed by the LRMP direction and Clean Water Act. Examples of resource objectives and their sources include:

- Greenline successional status value of at least 61, indicating late seral or the current value, whichever is greatest (Winward 2000, Burton et al. 2008) was developed in response to the PACFISH Enclosure B revision;
- Woody species regeneration sufficient to develop and maintain healthy woody plant communities (diversity of age and structure classes) was developed in response to MNF LRMP Amendment 29;
- Bank stability criteria (80% or current value, whichever is greatest for non-priority watersheds; at least 90% or current value, whichever is greatest, for priority watersheds) were from PACFISH;
- Water temperature criteria are from the MNF LRMP; and
- Width-depth ratio criteria are from PACFISH

The above resource objectives from the Forest direction and policies (See Section 3) are long term objectives that are achieved by developing a proposed action with project design criteria and annual monitoring indicators.

#### 4.1 PROPOSED ACTION

##### *Bear Allotment*

The Bear Allotment is located northeast of the town of John Day on National Forest System Lands, within T. 9 and 10 S, R. 32 E. The Allotment encompasses approximately 1,532 acres and is divided into 9 pastures: A, B, B1, C1, C2, D, E/F, G, and H (Figure 1 and 2).

##### *York On/Off Allotment*

The York On/Off allotment is located northeast of the town of John Day, within T 10 S, R 31 E. The Allotment encompasses approximately 780 acres and is divided into 2 pastures: Slide and East (Figure 3).

#### 4.1.1 PERMIT INFORMATION AND GRAZING SYSTEMS

##### *Bear Allotment*

The Bear Allotment is permitted for 84 cow/calf pairs (499 AUM's) from 6/1 to 10/15. Permittees, permitted livestock numbers, and permit issuance and expiration dates are identified in Table 7.

TABLE 7. PERMIT INFORMATION FOR THE BEAR ALLOTMENT

Permit Number	Permitted Livestock (Cow/Calf Pairs) / AUMs	Permit Issuance Date	Permit Expiration Date
01807	84 / 499	04/20/2005	12/31/2014

*Grazing System:*

- The Bear Allotment's grazing rotation system will utilize a deferred rotation with a staggered season of use entry system.
- Grazing will not be authorized on the C2 unit through 2015 as a result of the restoration work on the Middle Fork John Day River. An IDT evaluation will be conducted prior to the 2016 grazing season to evaluate the reauthorization of grazing in this pasture.
- Cattle will be excluded from accessing the Middle Fork John Day River in the C1 pasture using electric fencing. A water gap would be located on the river at the upstream end of the pasture.
- Range Readiness and utilization levels may vary on/off dates and pasture rotations within the parameters of authorize use.

***York On/Off Allotment***

The York On/Off Allotment is currently permitted for 12 cow/calf pairs from June 1<sup>st</sup> through October 31<sup>st</sup> (79 AUMs). Permittees, permitted livestock numbers, and permit issuance and expiration dates are identified in Table 8.

TABLE 8. PERMIT INFORMATION FOR THE YORK ON/OFF ALLOTMENT

Permit Number	Permitted Livestock (Cow/Calf Pairs) / AUMs	Permit Issuance Date	Permit Expiration Date
01760	12 / 79	05/09/2003	12/31/2012

*Grazing System:*

- Range Readiness and utilization levels may vary on/off dates and pasture rotations within the parameters of authorized use.

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**4.1.2 PROJECT DESIGN CRITERIA**

The following project design criteria (PDC) will be used to minimize or eliminate adverse effects of the PEs on MCR Steelhead, bull trout, and designated CH. The MNF regards these PDC as integral components of the proposed action and expects that all proposed project activities will be completed consistent with those measures.

1. Management will be framed in a manner that will allow managers to manipulate grazing strategies (dates, stocking levels, rotational patterns) depending on annual environmental factors and permittee success at meeting standards during the previous year.
2. Permittees must maintain perimeter and interior fences prior to turn-out.
3. Standards that are required of the permittee (e.g., turn on dates, move triggers, end point standards) will be outlined in an addendum to Part III of the grazing permit.
4. MSRA will be located and used to identify stream sections that are most vulnerable to livestock impacts. Identifying MSRA locations will guide application of bank alteration values.
5. Spawning surveys will occur within all pastures containing MSRA's where turn out is expected to occur prior to June 30. Of the remaining CH reaches 20% will be randomly



surveyed for redds where turn out is prior to June 30 (See Appendix F. Strategy to minimize Redd Trampling “Take” of Steelhead and Bull trout).

6. Where there is significant risk for redd trampling, the Forest and permittees will utilize a number of tools to protect redds, which include but are not limited to: alternative rotation, rest, exclusion with water gaps, temporary electric fences, additional riding.
7. Complete all required monitoring at PIBO Effectiveness Monitoring DMAs. DMA's to be monitored are provided to the Forest yearly by the EM Team via the Regional Office. This will effectively satisfy Interagency Implementation Team (IIT) monitoring requirements.
8. Annual use indicators will dictate when livestock are moved between units or off the allotment, within the terms of the term grazing permit, including moves in response to fish spawning. This will help us meet our long term riparian resource objectives.
9. The Forest Service will provide the Services with an End of Year Grazing Report by March 1 of each year.
10. Use of roads and off-road travel by permittees and staff will follow these PDC:
  - a. Vehicles are not authorized to travel through seeps, springs or streams except for use of existing fords on road crossings;
  - b. All refueling activities and fuel storage will occur at least 150 feet away from live streams;
  - c. OHV routes within 100 feet of streams will be camouflaged so that access routes do not become new trails and minimize disturbance to riparian vegetation;
  - d. OHV travel off established roads within 100 feet of streams would occur only during periods when soil is dry.

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#### 4.1.3 GRAZING USE INDICATORS AND SUPPORTING RATIONALE

The Forest Service's Regions 1, 4, 6 and Bureau of Land Management's Idaho, Montana, Nevada, Oregon and Washington have made commitments through the PACFISH and INFISH Management Strategies to protect and improve aquatic resources found in the interior Columbia River basin. Since the Forest Service Pacific Northwest Region (Oregon and Washington) which includes the Malheur National Forest began implementing these strategies, there has been marked improvement in management of aquatic resources. To strengthen the implementation of the aquatic strategies, an interagency group consisting of deputies from the various action and regulatory agencies (Deputy Team) was formed to provide oversight of the strategies and subsequent biological opinions commonly known as PIBO. Under the Deputy Team oversight, an implementation monitoring module was developed for livestock grazing, and its application is required where listed fish species occur in the interior Columbia River basin. Compliance with these requirements is monitored and presented to the Deputy Team during their annual reviews.

The PNW region requires application of the PIBO implementation and effectiveness monitoring program for National Forest LRMPs amended by PACFISH and INFISH, and the regional office annually coordinates the PIBO monitoring programs with the National Forests with listed fish species. FS line officers continue to work with their staffs and grazing permittees to ensure that implementation monitoring requirements are met. As described in the PIBO monitoring strategy and the annual regional coordination letter, the Forest established Designated Monitoring Areas (DMAs) and annually monitors the grazing use indicators at these PIBO DMAs as well as DMAs

established by the MNF.

Data collected at the PIBO effectiveness monitoring DMAs reflect the grazing use indicators applicable to stream banks and stream channels. Accordingly, the regional monitoring coordination letter identifies the following requirements will apply to these DMAs:

- A. Measurements will be on the greenline (first perennial vegetation above the channel). Measurements must include, at minimum: 1) bank alteration and 2) stubble height if any herbaceous vegetation is present.
- B. Where woody riparian vegetation dominates the DMA with little or no herbaceous vegetation along the greenline, woody use (browse) should be measured and may be sampled in lieu of stubble height.
- C. These measures will be made using the current MIM protocol.

Therefore, the grazing use indicators required by the PIBO and used by the Forest riparian monitoring programs at all DMAs are: browse of woody vegetation, stubble height of greenline vegetation and streambank alteration. Woody vegetation browse is used to regulate impacts on woody recruitment to streams, greenline stubble height is used to regulate grazing impacts on greenline ecological status and streambank alteration is used to regulate grazing impacts on streambank stability and channel width. For consistency with the PIBO monitoring program and regional direction regarding coordination with it, the Forest elected to use the current MIM for their monitoring protocol.

The MNF utilizes move trigger and endpoint (annual) indicators to manage livestock. The underlying concept behind the use of end-point indicators for livestock grazing management is that the selected end-points, if not exceeded, will allow for the attainment of or reasonable progress to be made toward desired conditions for riparian areas and fish habitat as described in Section 3 – Malheur NF LRMP and Section 4 – Proposed Action.

The MNF developed values for livestock move trigger and annual endpoint indicators (Table 9, 11, and 12). The ranges of values are starting points based on research and the MNF's best collective professional judgment for establishing desired riparian conditions. The end-point indicator values (allowable use in riparian areas) are, to the extent feasible and appropriate data are available, to be site-specifically designed to prevent any meaningful carry-over effects. They also provide for the evaluation of management practices to determine if they are effective in maintaining the desired and/or proper functioning condition, or improving the structure and function of riparian and aquatic conditions. These values could be adjusted as more site-specific information is gathered. End-point indicators (allowable use in riparian areas) should be adjusted for timing, intensity, frequency, and duration. The rationale for the development of the move trigger and end-point grazing use indicators is discussed below.

Livestock grazing along the greenline of stream channels will be limited to attain the numeric move trigger and end-point indicator values in Table 9, 11 and 12.

Annual Indicators will be adjusted if resource objectives are not being met.

TABLE 9 – BEAR ALLOTMENT ANNUAL USE INDICATORS

Pasture Name/ DMA or Key area Name / Creek Name	Monitoring Attribute	Key Species	Move Trigger	Endpoint Indicator
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C1 DMA site yet to be determined/Middle Fork John Day River MSRA-Present	Browse use		40% 30%	50% (early season use) 40% (late season use)
	Greenline stubble <sup>1</sup>	Deep-rooted hydric spp.	4 in.	4 in. (early season use) 6 in. (late season use)
	Streambank Alteration		10%	15%
C2 DMA site yet to be determined/Middle Fork John Day River MSRA-Present	Browse use		40% 30%	50% (early season use) 40% (late season use)
	Greenline stubble <sup>1</sup>	Deep-rooted hydric spp.	4 in.	4 in. (early season use) 6 in. (late season use)
	Streambank Alteration		10%	15%
E/F DMA site yet to be determined/Middle Fork John Day River MSRA-None	Browse use		40% 30%	50% (early season use) 40% (late season use)
	Greenline stubble <sup>1</sup>	Deep-rooted hydric spp.	4 in.	4 in. (early season use) 6 in. (late season use)
	Streambank Alteration		15%	20%
Upland Sites (All Pastures)	% Utilization	Upland grass species	35%	45%
Riparian Areas (All Pastures)	% Utilization	Riparian grass species	35%	45%

<sup>1</sup>Browse use and greenline stubble will be used until next trend reading is completed to determine which attribute will be best suited to attain long term objectives.

TABLE 10 – YORK ON/OFF ALLOTMENT ANNUAL USE INDICATORS

Pasture Name/ DMA or Key area Name / Creek Name	Monitoring Attribute	Key Species	Move Trigger	Endpoint Indicator
Slide/DMA site yet to be	Browse use		40%	50% (early season use)

determined/Slide Creek			30%	40% (late season use)
MSRA-Present	Greenline stubble <sup>1</sup>	Deep-rooted hydric spp.	4 in.	4 in. (early season use) 6 in. (late season use)
	Streambank Alteration		10%	15%
Upland Sites (All Pastures)	% Utilization	Upland grass species	35%	45%
Riparian Areas (All Pastures)	% Utilization	Riparian grass species	35%	45%

### Scientific Rationale to Support the Range of Initial Values for Selected End-Point Indicators/Condition Thresholds/Allowable Use Criteria

- Stubble Height: 4-6 inches;** Stubble height has been identified as being related to the physiological health and vigor of individual plants/communities as well as the ability of vegetation to protect streambanks and filter sediment during overbank flows, although by itself it is generally not sufficient to establish a relationship between grazing and riparian vegetative conditions. Research is limited, but the literature generally suggests 4-6 inches of residual stubble height allows for improved riparian grazing management and provides for adequate riparian protection. Clary and Leininger (2000) conducted studies on stubble height and its ability to improve riparian habitats and to capture and stabilize sediment. They concluded that stubble heights of 4-6 inches appear to stabilize the greatest amount of sediment. Clary (1999) states that by maintaining stubble heights of 4-5.5 inches allowed for streambank recovery. End-point indicator values are intended to vary by site depending on similarity to desired conditions and the resiliency of the site being monitored (UI Stubble Height Review Team Report 2004, Clary and Leininger 2000, Clary et al. 1996, Hall and Bryant 1995, PACFISH Enclosure B – Recommended Livestock Grazing Guidelines 1995, Clary and Webster 1989). Stubble height is an annual use indicator that should be used in combination with long-term monitoring of vegetation and stream channel attributes.
- Bank Alteration: 10-30%;** In general, the most widespread impact livestock have on riparian areas is trampling stream banks (Bengetyfield, 2006). Like stubble height, streambank alteration is another annual or short-term indicator used to evaluate the potential effects of livestock grazing in riparian areas, primarily evaluating potential effects to long-term streambank stability and channel shape. It is used as a tool to assess the intensity of grazing along streambanks and to determine when such intensity may be appropriate or deemed excessive. It can also prove useful in determining the cause-and-effect relationships between livestock grazing and stream channel conditions and whether management changes are needed for the following year. Streams are naturally dynamic

and have the ability to repair a certain amount of annual disturbance each year (the amount is variable based upon stream gradient, substrate composition, streambank materials, vegetation type and abundance, channel geometry, flow regime, etc.). Again, although the literature is not extensive, it generally suggests that 10-30% of annual streambank alteration is consistent with providing adequate riparian protection, and is intended to vary by site depending on similarity to desired conditions and the resiliency of the site being monitored (Burton et al. 2011, Heitke et al. 2008, Bengeyfield 2006, Cowley 2002, Bengeyfield and Svoboda 1998). Bengeyfield (2006) found that when streambank alteration measured 15-20%, width to depth ratios showed an improving trend. He also noted that the vegetation improvements kept pace with the physical changes. Streambank alteration is an annual use indicator that should be used in combination with long-term monitoring of streambank stability and channel geometry. In addition, it is worth noting that research is continuing to be conducted on the various ways that can be used to monitor for and measure actual streambank alteration (including MIM, which the District is presently using) to account for accuracy of results, reduction of variability among observers, and the resources necessary to carry out such measures.

- **Mean incidence of use on woody species: 30-50%;** Woody vegetation is an important component of many stream/riparian ecosystems as it can provide a strong root system, filter sediment, and provide stream shade and habitat diversity. Woody species browse is a short-term indicator of grazing utilization of woody species. Overall, there is generally a reduction in seed production of woody plants that receive more than 55 percent utilization, and when heavy and severe utilization levels are sustained over time overall plant health, including size and root strength, is reduced. Although the literature is not extensive, it generally suggests light to moderate allowable use on woody species (~30-50%) can be sustained and not meaningfully impede the potential for improved conditions of affected woody plant communities; and is intended to vary by site depending on similarity to desired conditions and the resiliency of the site being monitored (Winward 2000, Utilization Studies and Residual Measurements – USDI BLM 1996b, PACFISH Enclosure B – Recommended Livestock Grazing Guidelines 1995). Woody species browse is an annual use indicator that should be used in combination with the long-term monitoring indicators of woody species age class and greenline composition to help determine the health of woody plant communities.

Livestock grazing along the greenline of stream channels will be limited to attain the numeric move trigger and end-point indicator values in Tables 10, 11, and 12. The numeric values in Tables 10, 11, and 12 are considered starting points for allowable use since values could be adjusted as more site-specific information is gathered.

### **Initial Values for Grazing Use Indicators**

Based on the best available science, applied science publications, and professional judgment, the Forest interdisciplinary team selected initial values for each indicator. The season of use determined the initial values of endpoint indicators for woody shrub use and stubble height of greenline vegetation. The early season initial values for shrub use and stubble height are 50% and 4 inches, and the late season initial values for shrub use and stubble height are 40% and 6 inches, respectively. Grazing use in the early season allows time for vegetation growth after livestock use. The exact dates and times of early and late season can vary across the Forest and

between given years, and therefore are not specified. However, to provide some typical guidelines, early season is usually defined as the beginning of the growing season to mid-July and late season from mid-August to the end of the growing season.

To determine an initial value for the streambank alteration grazing use indicator, the Forest also looked at a Regional Technical Team (RTT) report prepared under the Streamlining Consultation Procedures resolution process. The NMFS had stated that “The best available science indicates that the 10% and 20% bank alteration levels are appropriate in preventing bank destabilization and protecting habitats critical to listed fish.” Their position paper cited numerous references to support these values. The RTT reviewed documents cited by NMFS, and they concluded that NMFS had reasonably established a causal link between streambank alteration-related habitat effects caused by livestock grazing activity and the taking of the species (i.e., grazing will affect stream channel conditions that will affect fish habitat conditions such as water quality, food, cover, etc.). The literature generally supports the concept that increased streambank alteration will, at some point, adversely affect stream channel conditions, and therefore fish habitat conditions. However, there was uncertainty relative to the percentage of streambank alteration at which habitat conditions were significantly altered and take of the species is likely to occur.

There is little field research supporting any specific percent streambank alteration standard using a defined and repeatable measurement protocol. The above RTT report and citations provide recommendations and professional judgments that range from 10% to 30% streambank alteration, but do not present empirical evidence from grazing monitoring data to support the percentages. The Forest interdisciplinary team recognizes a connection between the streambank alteration grazing use indicator and long term fish habitat conservation objectives in the LRMP, but couldn't determine a consensus value. Therefore, the Forest interdisciplinary team selected an initial value of 20% streambank alteration for endpoint grazing use indicator which is the statistical median of the range.

### **Adjustments to Values of Grazing Use Indicators and/or Grazing Strategy**

The interdisciplinary team considers available information on riparian condition (eg. succession status of greenline vegetation and woody species regeneration) and presence of MSRAs to adjust values of grazing use indicators and/or the grazing strategy. Information wasn't available for the successional status of the greenline and woody species regeneration for the pastures in this allotment. Therefore, it didn't affect the values of the grazing use indicators in Tables 10, 11, and 12. However, the presence of MSRAs resulted in reductions of the streambank alteration values for several pastures.

### **Criteria used to evaluate the riparian condition (eg. succession status of greenline vegetation and woody species regeneration):**

**When** these conditions apply:

- Greenline plant communities show moderate to high similarity to desired condition class/seral stage -- greenline successional status value is 41% or greater (mid-late seral) as defined by Winward (2000); *and/or*
- Stream/riparian systems have been assessed as being in Properly Functioning Condition or Functioning-at-Risk (high to moderate) category (Prichard et al. 1998); *and/or*
- Riparian/channel attributes are near desired conditions in a Unit,

**Then** allowable use within riparian areas will be:

1. <40-50% browse on both clumped, multi-stemmed species (i.e. most willows (*Salix spp.*) and single-stemmed species (i.e. coyote willow (*Salix exigua*), birch (*Betula spp.*), alder (*Alnus spp.*), cottonwood or quaking aspen (*Populus spp.*)) i.e. <50% browse if early season use, and <40% browse for mid and late season use<sup>5</sup>; and
2. >4-6 inch residual stubble height (will vary based on greenline successional status/seral stage, and season of use) i.e. >4 inches if early season use, and >6 inches for mid and late season use; and
3. Allowable bank alteration will be limited to 20% streambank alteration<sup>6</sup> by large herbivores (% of linear length of greenline altered) (Multiple Indicator Monitoring (MIM) of Stream Channels and Streamside Vegetation (Interagency Technical Reference 1737-23 2010)<sup>7</sup>. The estimated combined variability, observer error and sampling error or sample size, results in a 95% confidence interval of 6% for this bank alteration monitoring method. Thus, by setting a trigger for moving livestock at ~14%, we can be reasonably confident that livestock would be off the pasture before an additional 12% alteration was reached. The upper level for reasonable confidence would be 26% -- which represent an upper limit for the associated conservation measure.

**However**, when these conditions apply:

- When greenline plant communities show low similarity to desired condition class/seral stage -- greenline successional status value is less than 41 (early seral) as defined by Winward (2000); and/or
- Stream/riparian systems have been assessed as being in a *Functional-at-Risk (low) to Non-Functional category* (Prichard et al. 1998) {a Non-Functional system is one that clearly does not provide adequate vegetation, landform, or large woody debris to dissipate stream energy associated with high flows and thus are not reducing erosion, improving water quality, etc.; there is an absence of certain physical attributes such as a floodplain where one should be} in a Unit,

**Then,**

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<sup>5</sup> The exact dates and times of “early”, “mid” and “late” can vary across the Forest and between given years, and therefore are not specified. However, to provide some typical guidelines, “early” is usually defined as the beginning of the growing season to mid-July, “mid” season from mid-July to mid-August, and “late” season from mid-August to the end of the growing season.

<sup>6</sup> The allowable level of bank alteration for a specific site should allow for no more than 5% of the lineal bank distance (includes both sides of the stream) displaying evidence of new bank instability that has become perceptible after livestock grazing is initiated in a pasture. Note: hoof prints by themselves are not a sign of instability unless they move the bank by > 10 cm (direct shearing or sloughing of the bank).

<sup>7</sup> Research is presently ongoing, which may result in a new and or modified method of measuring Bank Alteration. If in fact this occurs the PIBO EM Team and or other researchers would present findings and provide a cross walk and rational to the existing monitoring method and endpoint indicators.

1. Consider resting the area/s for one or more years until condition reaches moderate similarity for those riparian areas with moderate and low gradient channels, such as Rosgen "B" and "C" channel types, with substrates composed of medium to fine easily eroded materials; or
2. If grazing is allowed, use should be for only short duration (i.e. to facilitate moves, etc.) and during a period of least environmental impacts

### **Most Sensitive Riparian Areas:**

The MNF has identified stream reaches of high quality steelhead spawning and rearing critical habitat called Most Sensitive Riparian Areas (MSRAs). The process and criteria for identifying MSRAs is described in detail in Appendix G- Methods for determining Most Sensitive Riparian Areas in relation to Mid Columbia River Steelhead. The miles of MSRA by stream and location in the allotments are displayed in Tables 3 and 5 and Figures 1 and 4, respectively.

MSRAs are typically steelhead critical habitat that is most accessible and sensitive to livestock use. MSRA and the grazing strategies described below are part of the proposed action. Certain grazing strategies can be used to minimize livestock and stream interactions and promote maintenance of, or recovery towards, desired conditions. Pastures containing MSRA that include one or more of the following grazing strategies would result in allowable use levels of 20% bank alteration versus the more restrictive standard of 15%:

- Rest Rotation – 1 year of complete rest during a grazing cycle (grazing cycle is typically 3-4 pastures)
- Double Rest Rotation – pasture is rested for two consecutive years, then grazed either early or late to following year depending on recovery needs (i.e. herbaceous or shrubs)
- Corridor Fencing – complete rest from grazing for a specified period of time or until specified objectives are met.

If at any point during this consultation a permittee adopts one or more of the strategies listed above the endpoint indicator would be adjusted to reflect such management changes in the annual instructions.

Other useful tools to minimize riparian use include – using a full-time rider (7 days/week), using electric fence, using low-stress stockmanship, placing low-moisture nutrient supplement blocks (as well as using other supplementations) in uplands, less than 21 days grazing duration in any pasture during the hot season (typically mid and late seasons, or mid-July to end of growing season). The use of these tools will be evaluated by the IDT on an annual basis to determine if the level of allowable use would be raised to 20% bank alteration. If none of these tools are in place, allowable use will remain at the 15% bank alteration.

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#### **4.1.4 MONITORING**

The Malheur National Forest monitoring strategy for determining condition and trend of riparian ecosystems as they relate to grazing activities was described in detail in Section 3.2. The goal is to determine site-specific desired riparian/stream channel conditions and the levels of allowable use (annual indicators also known as end-points) that will improve conditions that are not at the desired and/or proper functioning condition. The assessments and monitoring protocols used, as well as the values for desired conditions and allowable use, are intended to be an important part

of the adaptive management process and are subject to changes or modifications based on new scientific findings and improvements in methodologies as well as changes in definitions and policy.

The annual indicators are used in implementation monitoring to ensure that grazing does not prevent the attainment of the desired conditions. Riparian annual use indicators used on the Malheur National Forest include greenline stubble height, bank alteration, and woody browse. Greenline stubble height is used to regulate grazing impacts on greenline ecological status, bank alteration is used to regulate grazing impacts on bank stability, and woody browse is used to regulate impacts on woody recruitment. The specific indicators selected for a specific unit should be those that correspond with the riparian resources that are most sensitive to the impacts of livestock grazing. For example, if bank stability was the riparian feature most likely to be impacted by livestock grazing in a unit, then bank alteration would be selected as the annual use indicator for that unit.

Annual use indicators will be measured at key areas by key species (on uplands) and at DMA greenlines annually. Key areas are monitoring sites chosen to reflect the effects of grazing over a larger area (Burton et al 2008). Key species are preferred by livestock and an important component of a plant community, serving as an indicator of change (Coulloudon et al 1999). The Interagency Technical Reference or other best available science would be used to monitor grazing use. The MIM Interagency Technical Bulletin (Burton et al 2008) or other best available science would be used to monitor grazing use at DMAs. The Forest Service will monitor annual use indicators. Triggers will be used by permittees as a tool to help ensure annual use indicators are met. Endpoint indicators will be monitored by MNF personnel at designated monitoring areas (DMAs), following the MIM protocol (Burton et al. 2010). Move trigger evaluations will be conducted by the permittee.

Effectiveness (riparian objective) monitoring is designed to address the question of whether or not management practices currently applied to the area are achieving the desired results. These procedures are designed to assess the current condition and measure changes in streambanks, channels, and streamside vegetation over time, i.e., trend. They help determine if local livestock grazing management strategies and other land management actions are making progress toward achieving the long-term goals and objectives for streamside riparian vegetation and aquatic resources. The goal is to conduct effectiveness monitoring every three to five years on riparian areas and streambanks. This period of time is considered to be the minimum necessary to detect changes, although unusually wet years and/or flood events may result in short-term changes that validate the need to monitor more frequently, or at least at the time of the event. Budget and personnel constraints may limit the extent in which monitoring of this type will be conducted.

Presently, effectiveness monitoring includes: modified greenline composition (adapted from Winward 2000 and USDI BLM 1996a), woody species height class (Kershner et al. 2004), streambank stability and cover (adapted from Kershner et al. 2004), woody species age class (adapted from Winward 2000), greenline-to-greenline width (Burton et al. 2008), substrate (Bunte and Apt 2001), and residual pool depth and pool frequency (Lisle 1987). These provide data and information concerning the present conditions and trend of riparian vegetation, channels, and streambanks, and to help determine if aquatic systems are being degraded, maintained, or restored across the Malheur National Forest.

#### 4.1.5 ADAPTIVE MANAGEMENT

The adaptive management strategy described below and depicted in Appendix H diagrams 1.0 (Long-term) and 2.0 (Annual) is intended for allotments requiring consultation. It is designed to provide the MNF the ability to make management decisions based on new information, changing conditions, or the results of implementation/effectiveness monitoring. Adaptive management will be used to ensure: 1) sites at desired condition remain in desired condition; 2) sites not in desired condition have an upward trend; and 3) direction from consultation with the Services is met.

The overall strategy consists of a long-term adaptive management strategy and an annual adaptive management strategy. The long-term strategy describes how adaptive management will be used to ensure the three objectives previously stated are achieved and to maintain consistency with Forest Plan level direction. The annual adaptive management strategy describes how adjustments will be made within the grazing season to ensure annual use indicators and other direction from consultation is met, it also describes when and how regulatory agencies will be contacted in the event direction from consultation is not going to be met.

Ideally, the value associated with the annual use indicator is customized to the specific circumstances in each unit. However, customizing this value generally requires a significant amount of data and/or experience with a particular unit. As data is gathered and analyzed the annual use indicators may be adjusted to reflect the new information.

The annual use indicators within the Multiple Indicator Monitoring (MIM) method will be used to detect the annual use of wild horses, wildlife, and livestock at the end of a grazing period or growing season, whichever occurs first. Although the Proposed Action includes a suite of measures designed to avoid such an outcome, the MNF acknowledges that it is nevertheless possible that annual use indicators could be exceeded in a particular year. If this occurs, the MNF proposes the adaptive management process to be initiated immediately and will make any necessary adjustments to the current or future grazing strategy to ensure that the exceedances do not recur.

When the annual utilization data is collected at the end of the growing season, the MNF will consider adjustments of livestock numbers, timing of grazing, and duration of grazing. Or, the MNF may choose to rest the pasture or allotment. If big game populations exceed ODFW Management Objectives, appropriate coordination will occur among the agencies.

If there are recurring exceedances of annual indicators, or if there is a failure to comply with the terms and conditions of the grazing permit, the issuance of a Notice of Non-Compliance may be warranted. This notice, issued to the permittee(s), is likely to be in addition to the outcomes that result from following the adaptive management process described above. The issuance of a Notice of Non-Compliance and resulting action taken by the MNF will be consistent with FSH 2209.13 Section 16 and 36 CFR 222.4. All exceedances of annual indicators and subsequent grazing strategy adjustment recommendations will be documented by the MNF in the annual EOY Report and presented to the Level I consultation team. A specific strategy for when the endpoint indicator for streambank alteration is exceeded is discussed in Section 4.1.5.1 below.



#### 4.1.5.1 COMPLIANCE STRATEGY FOR THE STREAMBANK ALTERATION ENDPOINT INDICATOR

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The MNF acknowledges that there is a  $\pm 6\%$  margin of error associated with the MIM protocol (see section 4.1.3). Action would only be taken for permit violations and not as a result of wild ungulate or unauthorized use. The MNF will follow the strategy outlined below for exceedance of the bank alteration endpoint indicator. For each level of exceedance, the BMRD will incorporate adaptive management strategies into the following season's grazing strategy which may include: adjustments of livestock numbers, timing of grazing, or duration of grazing.

- **Measured bank alteration  $\leq 6\%$  over the endpoint indicator:** will be evaluated by the District IDT. The IDT will examine the level of measured use on stubble height and woody browse to determine if an exceedance of the endpoint indicator occurred. If the IDT concludes that the endpoint indicator has been exceeded the permittee will be contacted via phone or in person to notify them of the IDT findings. The permittee would be given 1 year to remedy. A follow-up letter will be sent to the permittee to document the verbal discussion and include what action is expected of the permittee to remedy the situation, to what standard, and by when (FSH 2209.13, 10, 16.2e).
- **Measured bank alteration 7-13% over the endpoint indicator:** the BMRD will, at a minimum, issue a notice of Non-Compliance for violation of terms and conditions of the term grazing permit and be given 1 year to remedy the non-compliance. Adjustments to the grazing strategy may be made following the adaptive management process. Failure to remedy the non-compliance during the following grazing season may result in a reduction of 25% of permitted AUMs for the following grazing season, **or** rest the pasture the following grazing season (FSH 2209.13, 10, 16.2e).
- **Measured bank alteration 14-20% over the endpoint indicator:** the BMRD will, at a minimum, issue a notice of Non-Compliance for violation of terms and conditions of the term grazing permit and will give the permittee 1 year to remedy the non-compliance. Adjustments to the grazing strategy may be made following the adaptive management process. When documented inspection indicates that the initial non-compliance has not been remedied as specified, or if a second situation of non-compliance has occurred, the permittee will be contacted by phone or in person describing the specific non-compliance. The BMRD will either reduce the authorized use by 25% of permitted AUMs for the following grazing season, **or** rest the pasture the following grazing season. A follow-up letter of a notice of permit action for non-compliance will be sent to the permittee indicating that a specified part of the permitted numbers or seasons is being suspended for a period of at least two years (FSH 2209.13, 10, 16.2e).
- **Measured bank alteration  $>21\%$  over the endpoint indicator:** the BMRD will, at a minimum, issue a notice of Non-Compliance for violation of terms and conditions of the term grazing permit and will give the permittee 1 year to remedy the non-compliance. Adjustments to the grazing strategy may be made following the adaptive management process. When documented inspection indicates that the initial non-compliance has not been remedied as specified, or if a second situation of non-compliance has occurred, the permittee will be contacted by phone or in person describing the specific non-compliance. The BMRD will reduce the authorized use by 25% of permitted AUMs for the following grazing season, **and** rest the pasture the following grazing season using the adaptive management process. A follow-up letter of a notice of permit action for non-

compliance will be sent to the permittee indicating that a specified part of the permitted numbers or seasons is being suspended for a period of at least two years (FSH 2209.13, 10, 16.2e).

Recurring non-compliance may lead to suspension of AUMs and/or the cancellation in part or whole of the Term Grazing Permit. Permit action involving the suspension or cancelation of grazing permits as per direction outlined in FSH 2209.13, 10, 16.2 and 36 CFR 222.4.

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#### 4.1.6 COORDINATION AND REPORTING

##### *Reporting*

Annual end-of-year (EOY) grazing reports are prepared by BMRD staff for all livestock grazing allotments. The reports include monitoring results, descriptions of any exceedance of grazing end-points and recommendations for management changes for the next grazing season. See monitoring section for a description of the grazing use and stream channel condition indicators for which information is collected, evaluated and reported. The report is sent to the NMFS and/or FWS by March 1 of each year.

##### *Coordination*

EOY report. Both internal and external coordination takes place regarding information and recommendations for changes in management found within the EOY report. The recommendations for changes in management in the EOY report are developed in an interdisciplinary manner. Typically, range conservationists, fish biologists, hydrologists, and the line officer will be involved. On occasion, wildlife biologists and botanists will participate.

Level 1 Team Meeting: A Level 1 team meeting is scheduled after a draft EOY report is sent to NMFS and/or FWS. The Level 1 Team discusses the draft EOY monitoring results, proposed remedies, and application of the compliance strategy (Section 4.1.5.1).

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#### 4.2 INTERRELATED ACTIONS

The regulations require the MNF to impose penalties for violation of prohibited acts on public lands. Unauthorized use is a prohibited act, and therefore is not a federal action. If unauthorized use occurs, the MNF's response could constitute a separate, interrelated federal action.

Unauthorized livestock grazing can occur in the allotment, and is reasonably certain to occur in the future. It is not a large problem in the allotments analyzed in this BA. Typically, it involves a few cow/calf pairs every few years.

Forest Service grazing regulations define unauthorized use, also known as "trespass," as occurring when livestock not under permit enter National Forest System (NFS) lands. It is a violation of 36 CFR 261.7. When unauthorized use occurs, the MNF attempts to identify and contact the owner of the livestock with instructions to remove the unauthorized livestock from NFS. The MNF can then bill the owner for the unauthorized use at the appropriate rate as identified in 36 CFR 222.50(h). If the ownership of the livestock is unknown, or the owner fails to comply with instructions to remove the livestock, the impoundment of said livestock by the MNF can occur as per 36 CFR 262.10.

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#### 4.3 PROJECT ELEMENTS

Project elements are the component parts of the action. Project elements will be assessed in the effects analysis section of the BA. Several of the project elements involve the use of vehicles on and off roads to access sites, such as four wheel drive trucks and OHVs.

1. Livestock use of allotment/pastures. Livestock will utilize the allotment/pastures consistent with the permitted numbers, season of use and grazing system described above and in the term grazing permit.
2. Permittee management of livestock and infrastructure maintenance. This includes move-in and move-out of cattle, herding, placement of nutrient (salt blocks) in the uplands, and maintenance of troughs, springs, ponds, fences and gates. Use of highway and off-road vehicles is included in this PE.
3. Range improvements. This includes the construction of fences for riparian pastures, and the construction/development of off-stream water sources.
4. Exclusionary fencing. Fences are constructed or placed to exclude areas from grazing. This is done to prevent livestock damage of riparian areas and in the case of electric fencing, to eliminate the potential for cattle stepping on redds.
5. Monitoring. A variety of implementation and effectiveness monitoring techniques are employed to determine if desired conditions are being met. The MNF Riparian Monitoring Strategy is discussed in detail in Section 4.1.4 below. Workers use manual and electronic equipment to measure vegetation, water quality and stream channel/streambed characteristics.
6. Adaptive management. An adaptive management strategy is designed to provide the MNF the ability to make management decisions based on new information, changing conditions, or the results of implementation/effectiveness monitoring. It will be used to ensure: (1) Sites at desired condition remain in desired condition; (2) sites not in desired condition have an upward trend; and (3) direction from ESA consultation with NMFS is met. The adaptive management strategy describes how adjustments will be made to ensure annual endpoint indicators as well as other direction from this consultation are met, and describes when and how regulatory agencies will be contacted in the event direction from this ESA consultation is not going to be met. The MNF Adaptive Management Strategy is described in Section 4.1.5.

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#### 4.4 ESA ACTION AREA

The ESA action area for this consultation is defined by the Bear and York On/Off allotments that flow into the John Day River basin (Figure 4). The ESA action area is displayed by the cross-hatched area in Figure 4.





The MCR Steelhead DPS includes all naturally-spawned populations of steelhead in streams within the Columbia River basin from above the Wind River in Washington and the Hood River in Oregon (exclusive), upstream to, and including, the Yakima River in Washington, excluding steelhead from the Snake River basin (64 FR 14517; March 25, 1999). The major tributaries occupied by this DPS are the Deschutes, John Day, Klickitat, Umatilla, Walla Walla, and Yakima River systems. The John Day River (JDR) probably represents the largest naturally spawning and most viable native stock of steelhead in the region (NMFS 2009). The MCR Steelhead DPS does not include co-occurring resident forms of *O. mykiss* (rainbow trout).

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### 5.1.2 LIFE HISTORY AND HABITAT REQUIREMENTS

Steelhead trout are the anadromous form of *O. mykiss*. Adult summer steelhead typically return to freshwater from June through September. Adults overwinter in large rivers while sexually maturing. Adults resume migration to spawning streams in early spring.

The JDR adult summer steelhead enter the lower river as early as September and as late as March, depending on water temperatures. Adult migration in the JDR generally peaks in October. The JDR below the North Fork JDR is used only for migration due to high summer water temperatures. Spawning takes place from March through May. Eggs incubate during the spring and emergence occurs from April through July depending on water temperatures. Juveniles typically rear for 2 to 3 years in freshwater before smolting and migrating to the ocean.

Juvenile steelhead generally utilize habitats with higher water velocities than juvenile Chinook salmon. In winter, juveniles utilize deep pools with abundant cover. Juveniles may reside in their natal stream for their entire freshwater rearing phase or may migrate to other streams within a watershed. Smoltification occurs during late winter and emigration to the ocean occurs during spring. Smolts outmigrate rapidly, taking 45 days or less to reach the ocean from upstream rearing areas. In the JDR below the North Fork, smolts generally stay within the thalweg, taking advantage of cover provided by depth and turbidity. Approximately 80% of the steelhead rear in the ocean for 2 years before returning to the JDR system as adults to spawn (PD BLM 2006).

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### 5.1.3 MCR STEELHEAD POPULATIONS

The Interior Columbia Basin Technical Recovery Team (ICTRT) (2003) identified 15 populations in four major population groups (MPG) (Cascades Eastern Slopes Tributaries, John Day River (JDR), the Walla Walla and Umatilla Rivers, and the Yakima River) and one unaffiliated independent population (Rock Creek) in this steelhead DPS. There are two extinct populations in the Cascades Eastern Slopes Tributaries MPG, the White Salmon River and Deschutes River above Pelton Dam.

The JDR Subbasin contains the MCR Steelhead JDR MPG that consists of the Lower Mainstem John Day (LMJD), North Fork John Day (NFJD), Middle Fork John Day (MFJD), South Fork John Day (SFJD), and Upper Mainstem John Day (UMJD) populations (ICTRT 2003). The action area is associated with the MFJD population.

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### 5.1.4 MCR STEELHEAD DPS VIABILITY STATUS

The status of a salmon or steelhead species is expressed in terms of likelihood of persistence over 100 years, or in terms of risk of extinction within 100 years. The ICTRT defined viability at two

levels: less than 5 percent risk of extinction within 100 years (viable) and less than 1 percent risk of extinction within 100 years (highly viable). A third category, “maintained,” represents a less than 25 percent risk. The risk level of the steelhead DPS as a whole is built up from the aggregate risk levels of the populations and MPGs. The viable salmonid population (VSP) parameters (abundance, productivity, spatial structure, and diversity of the component populations) must be taken into account to determine the risk level.

The MCR Steelhead DPS does not currently meet viability criteria because its four component MPGs are not at low risk. However, for this DPS the outlook is relatively optimistic. One population, North Fork John Day, is currently at very low risk or “highly viable.” Two populations are currently viable (Deschutes Eastside, Fifteenmile); eleven are at moderate risk, with good prospects for improving. However, three large populations at high risk (Deschutes Westside, Naches, and Upper Yakima) are important to DPS viability; these present significant challenges.

Significant programs are underway for natural recolonization (White Salmon) or reintroduction (Deschutes Crooked River above Pelton Dam) of two of the extirpated populations to historically accessible habitat. Success of these programs should help improve overall DPS viability.

The MCR Steelhead Recovery Plan (NMFS 2009) presented viability ratings for the MCR Steelhead MPG. The risk of extinction is displayed as a combination of ratings for Abundance/Productivity Risk and Spatial Structure/Diversity Risk (Figure 5). The North Fork John Day MPG rates low/very low by the two criteria. The Middle Fork and South Fork MPGs rate low/moderate and the Lower Mainstem and Upper Mainstem MPGs have the highest extinction risk at moderate/moderate.

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#### 5.1.5 JOHN DAY RIVER MPG POPULATION STATUS

The current status of the MCR Steelhead John Day River MPG populations, showing 10-year geometric mean abundance by population, estimated productivity, and the minimum abundance threshold needed for long-term viability is summarized in Table 11. The table also includes the 10-year geometric mean proportion of hatchery spawners for the populations where data are available, and the risk ratings of high, moderate, low, and very low, for abundance and productivity combined, and spatial structure and diversity combined.

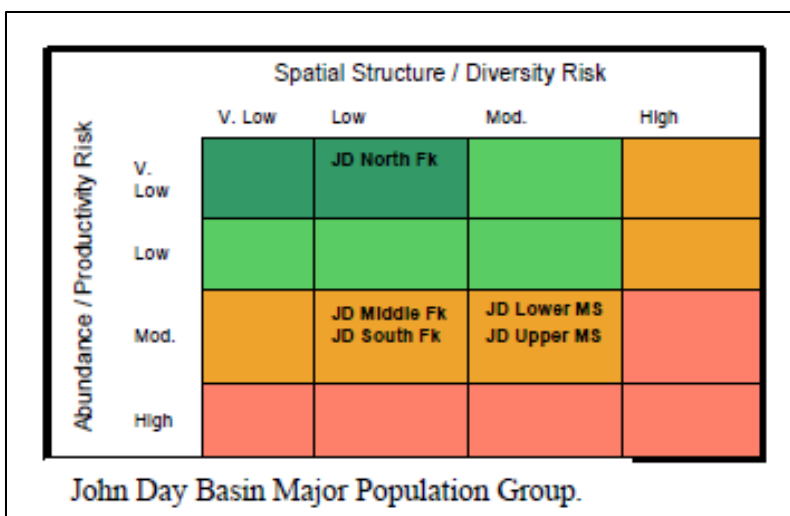


FIGURE 5. VIABILITY RATINGS FOR THE MCR STEELHEAD MAJOR POPULATION GROUP (NMFS 2009). SHADES OF GREEN INDICATE LOWER RISK OF EXTINCTION AND SHADES OF RED INDICATE HIGHER RISK.

TABLE 11. MCR STEELHEAD JOHN DAY RIVER MPG SUMMARY OF ABUNDANCE, PRODUCTIVITY, RISK RATINGS, AND MINIMUM ABUNDANCE THRESHOLDS (SOURCE: MIDDLE COLUMBIA RIVER STEELHEAD DPS RECOVERY PLAN SUMMARY 2009).

Population	Abundance Threshold <sup>1</sup>	Size Category	Run Timing	10-year Geomean abundance	Abundance Range	10-yr Hatchery Fraction <sup>2</sup>	Productivity <sup>3</sup>	Productivity Standard Error	A&P Risk Rating <sup>4</sup>	SSD Risk Rating
Lower Mainstem John Day	2250	Very Large	Summer	1800	563-6257	0.1	2.99	0.24	M	M
North Fork John Day	1500	Large	Summer	1740	369-10,235	0.08	2.41	0.22	VL	L
Upper Mainstem John Day	1000	Intermed.	Summer	524	185-5169	0.08	2.14	0.33	M	M
Middle Fork John Day	1000	Intermed.	Summer	756	195-3538	0.08	2.45	0.16	M	M
South Fork John Day	500	Basic	Summer	259	76-2729	0.08	2.06	0.27	M	M

<sup>1</sup> Abundance threshold for viability based on habitat intrinsic potential

<sup>2</sup> Average proportion of hatchery spawners over most recent 10 years in the data series.

<sup>3</sup> Geomean return per spawner calculated over most recent 20 years in data series.

<sup>4</sup> Abundance & Productivity Risk Ratings: H = high risk, M= moderate risk, L = low risk, VL = very low risk

### 5.1.6 POPULATION LIMITING FACTORS

The Middle Columbia River Steelhead ESA Recovery Plan (NMFS 2009) identified population limiting factors. For the NFJD population the primary tributary habitat limiting factors identified by the recovery planning team are degraded floodplain connectivity and function, degraded channel structure and complexity (habitat quantity, habitat diversity, channel stability), altered sediment routing, water quality (temperature), and altered hydrology. For the MFJD population they are degraded floodplain and channel structure (habitat quantity/diversity), altered sediment routing, altered hydrology and water temperature. The primary tributary limiting factors for the SFJD population include altered sediment routing, degraded floodplain and channel structure (habitat quantity and habitat diversity), altered hydrology, water quality (temperature) and blocked or impaired fish passage.

Tributary limiting factors for the UMJD population include degraded channel structure and complexity (habitat quantity and diversity), degraded riparian areas and LWD recruitment, altered sediment routing, water temperatures and altered hydrology. Impaired fish passage is also a priority limiting factor in Beech and Laycock creeks. Habitat limiting factors identified in ODFW (2010) for the Beech Creek Major Spawning Area within the ESA action area are displayed in Table 12.

TABLE 12. HABITAT LIMITING FACTORS WITHIN THE BEECH CREEK MAJOR SPAWNING AREA IDENTIFIED IN NMFS (2009).

Limiting Factor	Beech Creek Major Spawning Area
Degraded floodplain connectivity and function	X
Degraded channel structure and complexity	Lower reaches next to Highway 395
Altered hydrology	Lower reach of McClellan Creek
Altered sediment routing	X
Water Temperature	X
Degraded riparian communities	X
Man-made block to migration	
Impaired fish passage	X

## 5.2 CRITICAL HABITAT FOR MIDDLE COLUMBIA RIVER STEELHEAD DPS

### 5.2.1 DESIGNATION HISTORY

Critical habitat (CH) was designated for MCR Steelhead on February 16, 2000 (65 FR 7764) that encompassed the major Columbia River tributaries known to support the DPS, including the



Deschutes, John Day, Klickitat, Umatilla, Walla Walla, and Yakima Rivers, as well as the Columbia River and estuary.

In late 2000, a lawsuit was filed challenging the NMFS February 2000 final designation of CH for ESUs/DPSs of Pacific salmon and steelhead listed under the ESA. A federal court ruled that the agency did not adequately consider the economic impacts of the CH designations. In April 2002, NMFS withdrew its 2000 CH designations.

Critical habitat for MCR Steelhead was designated again on September 2, 2005 (70 FR 52630). Designated CH includes the stream channels within the designated stream reaches, and includes a lateral extent as defined by the ordinary high-water line (33 CFR 319.11). In areas where ordinary high-water line has not been defined, the lateral extent is defined by the bankfull elevation. Bankfull elevation is the level at which water begins to leave the channel and move into the floodplain and is reached at a discharge which generally has a flood recurrence interval of 1 to 2 years on the annual flood series.

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### 5.2.2 PRIMARY CONSTITUENT ELEMENTS

The physical or biological features of CH essential to the conservation of the species are known as primary constituent elements (PCEs). The PCEs of MCR Steelhead CH are those sites and habitat components that support one or more life stages, including:

- (1) Freshwater spawning sites with water quantity and quality conditions and substrate supporting spawning, incubation and larval development;
- (2) Freshwater rearing sites with:
  - (i) Water quantity and floodplain connectivity to form and maintain physical habitat conditions and support juvenile growth and mobility;
  - (ii) Water quality and forage supporting juvenile development; and
  - (iii) Natural cover such as shade, submerged and overhanging large wood, log jams and beaver dams, aquatic vegetation, large rocks and boulders, side channels, and undercut banks.
- (3) Freshwater migration corridors free of obstruction and excessive predation with water quantity and quality conditions and natural cover such as submerged and overhanging large wood, aquatic vegetation, large rocks and boulders, side channels, and undercut banks supporting juvenile and adult mobility and survival.

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### 5.2.3 STATUS OF MIDDLE COLUMBIA RIVER STEELHEAD CRITICAL HABITAT

Migratory habitat quality for MCR steelhead has been severely degraded by the development of the Federal Columbia River Power System. Depending on the their natal watershed, adults and out-migrating juvenile steelhead encounter between one and three mainstem Columbia River dams migrating to and from the ocean. Hydroelectric development has modified natural flow regimes resulting in higher water temperatures, changes in fish community structure, and increased travel time for migrating adults and juvenile salmonids. Physical features of dams such as turbines also kill migrating fish. The only substantial habitat blockages at present for this species are Pelton Dam on the Deschutes River and Condit Dam on the White Salmon River. However, minor blockages from smaller dams, impassable culverts, and irrigation dams occur

throughout the region. Several dams in the John Day River basin previously blocked habitat, but they have since been modified with ladders; however, there is a possibility that local native stocks were extirpated before these ladders were built (NMFS 2004).

Water quality impairment that affects spawning, migration, and rearing is a problem in many areas of designated CH for the MCR Steelhead. Summer stream temperature is the primary water quality problem for this species, and many of the stream reaches proposed as CH are listed on the Clean Water Act (CWA) 303(d) list for water temperature. Many areas that were historically suitable rearing and spawning habitat are now unsuitable due to high summer stream temperatures. Elevated stream temperatures may form thermal barriers to juvenile migration within tributaries. Removal of riparian vegetation, alteration of natural stream morphology, and withdrawal of water for agricultural or municipal use all contribute to elevated stream temperatures. Contaminants such as insecticides and herbicides from agricultural run-off and heavy metals from mine waste are common in some areas of designated critical habitat for this species.

Low summer stream flow is also a common characteristic affecting spawning, rearing, and migration PCEs for this DPS. There is little or no late summer flow in sections of the lower Umatilla and Walla Walla rivers. Withdrawal and storage of natural stream flow in spawning and rearing areas have altered hydrological cycles, causing a variety of adverse impacts to MCR Steelhead habitat. Increased summer stream temperatures, migration blockages, stranding of fish, and alteration of sediment transport processes can result from water withdrawal for irrigation or municipal use (NMFS 1996; Spence *et al.* 1996). In many river basins, the amount and quality of available rearing habitat has been reduced by water withdrawals. Many stream reaches are over-appropriated under state water law, with more allocated water rights than existing stream flow conditions can support.

Spawning and rearing salmonids, such as steelhead, require physically complex lotic habitats with pools, large woody debris, undercut banks, and substrates with low levels of fine sediments (Spence *et al.* 1996; Bjornn and Reiser 1991). Although these habitat conditions are still present in many wilderness, roadless, and undeveloped areas, recent subbasin assessments and plans (NPCC 2005) indicate that habitat complexity has been greatly reduced in many areas of designated critical habitat. Channel and riparian alterations for agricultural purposes, transportation, mining, forestry and other development activities have affected spawning, rearing and migration PCEs by reducing overall habitat complexity, cover, food availability, and spawning and rearing quality and quantity.

Under section 303(d) of the Clean Water Act, the Oregon Department of Environmental Quality (ODEQ) identified many streams within the LJD, UJD, MFJD, and NFJD watersheds that are water quality limited for high temperatures, dissolved oxygen, or biological criteria. Additionally, the ODEQ identified total phosphates and fecal coliform as water quality limitations for many streams within the Lower Mainstem John Day River, and sediment for many NFJD streams (NMFS 2004).

Critical Habitat Analytical Review Teams (CHARTs) were convened by NMFS for each recovery domain (NMFS 2005). CHARTs were charged with analyzing the best available data for each listed species, to make findings regarding the presence of essential habitat features in each watershed, identify potential management actions that may affect those features, and determine the conservation value of each watershed within each species' range. The action area

occurs within two 5<sup>th</sup>-field HUCs in the range of MCR steelhead: Big Creek-Middle Fork John Day and Camp Creek-Middle Fork John Day. These HUCs have high conservation value. Mid-Columbia CHART members noted that PCEs in these HUCs support unique genetic resources since there is minimal hatchery influence on these populations.

The John Day Subbasin Plan (NPCC 2005) included an Ecosystem and Diagnostic Treatment (EDT) analysis of habitat conditions for the 5<sup>th</sup> field HUCs located in the action area. The approach was to display the top quartile protection and/or restoration 5<sup>th</sup> field HUCs and their important restoration attributes.

Fourteen 5<sup>th</sup> field HUCs identified as important to Middle John Day summer steelhead were evaluated. The Big Creek and Camp Creek 5<sup>th</sup> field HUCs creeks were in the top quartile for restoration benefit. They were also listed as high priority for both protection and restoration, signifying that all three should be protected from further degradation and that restoration on any of the limiting factors listed would have the potential to increase productivity and abundance for the population (NPCC 2005). The attributes for restoration in common for these watersheds were sediment load and key habitat quantity.

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### 5.3 COLUMBIA RIVER BULL TROUT DISTINCT POPULATION SEGMENT

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#### 5.3.1 LISTING HISTORY, LOCATION AND STATUS REVIEW

The U.S. Fish and Wildlife Service (FWS) listed the Columbia River and Klamath River populations of bull trout (*Salvelinus confluentus*) in a final rule as “threatened” on June 10, 1998 (63 FR 31647) (FWS 1998) under the authority of the Endangered Species Act of 1973. The Columbia River DPS occurs throughout the entire Columbia River basin within the United States and its tributaries, excluding bull trout found in the Jarbidge River, Nevada. Five different DPS were identified in the 1998 final rule. In a subsequent 1999 final rule for listing other DPS (64 FR 58909), the five DPSs were combined into one DPS for the coterminous United States (FWS 1999). In the 1999 final rule, the FWS stated: “ In recognition of the scientific basis for the identification of these bull trout population segments as DPS’s, and for the purposes of consultation and recovery planning, we will continue to refer to these populations as DPS’s. These DPS’s will serve as interim recovery units (RUs) in the absence of an approved recovery plan.”

The FWS conducted a 5-year review on the status of bull trout in 2008 (FWS 2008). The review concluded that a change in classification from “Threatened” was not warranted. The review also evaluated the DPS policy. The FWS concluded that since the conterminous listing in 1999, new information suggests potential changes in the number of distinct population segments and their boundaries (e.g., Spruell et al., 2003). In addition to strong scientific evidence continuing to support identification of multiple population segments of bull trout, the Service and some State and Tribal partners have identified policy reasons for revisiting the Service’s application of its DPS policy to bull trout.

Consequently, the FWS recommended evaluating designation of multiple bull trout DPS’s and stated that it would initiate a new, separate status assessment effort to identify DPSs and evaluate their status. Recommendation 3 in the recommendations for Future Action section of the five-year review is to develop a number of RUs for bull trout (perhaps 5 to 10 for management purposes) that contain assemblages of core areas that retain genetic and ecological integrity, and

allow potential future options to pursue regulatory relief/delisting on a RU basis.

The final rule designating bull trout CH (75 CFR 63898) (FWS 2010) discussed the 5-year review recommendation and stated that six draft RUs were subsequently identified. Each of the six RUs was evaluated. It was confirmed that each RU was needed to ensure a resilient, redundant, and representative distribution of bull trout populations throughout the range of the listed entity. To accomplish these goals, protection of large areas of high-quality habitat, multiple populations, and diverse genetic and life-history aspects will be required. The six draft RUs identified for bull trout in the coterminous United States include: Mid-Columbia (containing the JDR basin); Saint Mary; Columbia Headwaters; Coastal; Klamath; and Upper Snake. These six new biologically based RUs will be proposed to replace the 27 RUs previously identified in the bull trout Draft Recovery Plan (DRP).

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### 5.3.2 LIFE HISTORY AND BIOLOGICAL REQUIREMENTS

The taxonomy, physical description, distribution, life history, habitat characteristics and diet of bull trout are described in detail in the final rule for listing the Columbia River DPS (63 FR 31647) (FWS 1998), which is herein incorporated by reference.

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#### 5.3.3 BULL TROUT POPULATIONS

A bull trout DRP was prepared in 2002 (FWS 2002). The DRP identifies the JDR Basin as one of 22 bull trout RUs for the Columbia River population. RUs place the scope of bull trout recovery on smaller spatial scales than distinct population segments.

Chapter 9 of the DRP focuses on the JDR basin. As of 2002, bull trout distribution was limited primarily to the headwaters of the North Fork JDR, Middle Fork JDR, and upper mainstem JDR and tributaries, with seasonal use of the mainstem river downstream to the vicinity of the town of John Day (Ratliff and Howell 1992, Buchanan *et al.* 1997).

The John Day River Recovery Unit Team (JDRRUT) identified one core area (the John Day River Core Area). For the purposes of recovery, a core area represents the closest approximation of a biologically functioning unit. Core areas consist of both habitat that could supply all the necessary elements for every life stage of bull trout (*e.g.*, spawning, rearing, migratory, and adult) and have one or more groups of bull trout. Core areas are the basic units on which to gauge recovery within a RU. The JDRRUT also identified 12 extant local populations. Three are located in the Middle Fork JDR subbasin: Clear Creek; Granite Boulder Creek; and, Big Creek, with interconnective migratory habitat between populations (Buchanan *et al.* 1997). While most bull trout in these streams appear to be resident forms, larger, potentially migratory bull trout have recently been observed in the Middle Fork JDR (T. Unterwegner, ODFW, pers. comm. 2000).

Streams with documented bull trout spawning in the Middle Fork JDR include Clear Creek, Big Creek, Deadwood Creek and Granite Boulder Creek. Current distribution in the Middle Fork JDR is based on isolated sightings with the primary distribution restricted to tributaries and limited to 22 percent of stream miles previously known to support bull trout (Claire and Gray 1993, Buchanan *et al.* 1997). Ratliff and Howell (1992) considered bull trout to be extirpated from the Middle Fork JDR and at high risk of extinction in the tributaries. The status evaluation by Buchanan *et al.* (1997), maintains this assessment with the addition of Clear Creek as a high

risk tributary.

Biological assessments for the Middle Fork JDR subbasin (MNF 1998a and 1999) provide detailed descriptions of baseline habitat conditions. Resident (summer distribution) bull trout currently occupy only approximately 23 miles of streams in the MFJD metapopulation area (9.5 miles in Big Creek; 4.5 miles in Granite-Boulder Creek; and nine miles in Clear Creek) (USDA 1998a). The 1990 and 1992 Oregon Department of Fish and Wildlife Aquatic Inventory Project also identified 2.5 miles of summer habitat in Deadwood Creek, a tributary to Big Creek. Bull trout migration from these tributary streams during the summer is highly unlikely due to high water temperatures and habitat modifications in the mainstem.

Aquatic inventory surveys conducted by the Oregon Department of Fish and Wildlife in 1990 and 1991 detected 60 bull trout in the Middle Fork JDR basin; two fish were measured at 260 millimeters (10 inches) and 360 millimeters (14 inches), all others were less than 210 millimeters (8 inches) in length (Buchanan *et al.* 1997). In the 1999 and 2000 surveys of Clear Creek, eight redds were observed each year (MNF 2001).

#### 5.3.4 BULL TROUT VIABILITY STATUS

The FWS conducted a conservation status assessment for 121 core areas (FWS 2005).

The threat rank characterization combines an analysis of severity, scope and immediacy of threats to bull trout. Severity captures the degree to which a threat impacts the population and the degree to which the threat is reversible. Scope refers to the proportion of the core area affected by a threat. The ranking for immediacy of threat was a straightforward analysis of the operational timeframe of the threat. A rank of high immediacy meant the threat was operational immediately or within a year. For a rank of moderate immediacy, the threats would be operational in 2 to 5 years, and for a rank of low immediacy, the threats were estimated to be operational in 5 to 20 years. The Natural Heritage Program criteria attach greatest significance to the severity of the threat, followed by scope and then immediacy in synthesizing the three threat categories into eight rank classifications. The result for all 121 core areas is shown in Table 13.

TABLE 13. THREAT CATEGORY RANKING FOR 121 BULL TROUT CORE AREAS (AFTER FWS 2005)

Threat category	Number of core areas
Substantial, imminent threat	44
Moderate, imminent threat	31
Substantial, non-imminent threat	3
Moderate, non-imminent threat	7
Localized, substantial threat	2
Widespread, low-severity threat	19
Slightly threatened	18
Unthreatened	5

Unknown	2
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Three of the core areas are in the JDR basin: MFJD; NFJD; and UMJD. A summary of the status assessment results is presented in Table 14. A final rank of “at-risk” indicates that a core area has very limited and/or declining numbers, range, and/or habitat, making the bull trout in this core area vulnerable to extirpation.

TABLE 14. RESULTS OF 2005 CONSERVATION STATUS ASSESSMENT FOR JDR BASIN CORE AREAS FOR BULL TROUT (ADAPTED FROM FWS 2008).

Core Area	Population Abundance Category (individuals)	Distribution Range Rank (stream length miles)	Short-term Trend Rank	Threat Rank	Final Rank
Middle Fork John Day River	Unknown	125-620	Increasing	Substantial, imminent	At-risk
North Fork John Day River	Unknown	125-620	Increasing	Substantial, imminent	At-risk
Upper Mainstem John Day River	1-50	125-620	Increasing	Moderate, non-imminent	At-risk

### 5.3.5 JOHN DAY RIVER BULL TROUT POPULATION STATUS

A key concern for the JDR core area is isolation and habitat fragmentation. The major isolating mechanism affecting bull trout local populations in the JDR basin is seasonally inadequate water quality and quantity in the mainstem river and tributaries, due to degraded riparian and stream habitat conditions. Other barriers include low head dams, diversions, and natural waterfalls (Claire and Gray 1993, ODFW *in litt.* 2000).

Populations within the Middle Fork subbasin are at greatest risk from isolation due to habitat fragmentation, seasonally high water temperatures, and reduced flows in the connecting mainstems (ODFW *in litt.* 2000). Bull trout are found in only three Middle Fork tributaries that are geographically distant. Population estimates for two of the tributaries are below 800 total fish of all ages, and existing data show no evidence of interchange between the local populations (ODFW *in litt.* 2000).

Bull trout in the MFJD River persist at low abundance levels. In 1999, population surveys were conducted in Clear Creek, Big Creek, Deadwood Creek, and Granite Boulder Creek to estimate abundance. Total numbers of bull trout consisting of primarily juvenile and sub-adult fish, were estimated to be 1,950 individuals in Big Creek, 640 individuals in Clear Creek, and 368 individuals in Granite Boulder Creek (Hemmingsen 1999).

Recovery goals and objectives are described in Chapter 9 of the DRP. The goal of the bull trout recovery plan is to ensure the long-term persistence of self-sustaining, complex interacting groups of bull trout distributed throughout the species’ native range, so that the species can be



delisted. Four objectives were identified to achieve the goal for bull trout in the JDR RU:

- Maintain current distribution of bull trout and restore distribution in previously occupied areas within the JDR RU.
- Maintain stable or increasing trends in abundance of bull trout.
- Restore and maintain suitable habitat conditions for all bull trout life history stages and strategies.
- Conserve genetic diversity and provide opportunity for genetic exchange.

Rieman and McIntyre (1993) and Rieman and Allendorf (2001) evaluated the bull trout population numbers and habitat thresholds necessary for long-term viability of the species. They identified four elements, and the characteristics of those elements, to consider when evaluating the viability of bull trout populations. These four elements are: 1) number of local populations; 2) adult abundance (defined as the number of spawning fish present in a core area in a given year); 3) productivity, or the reproductive rate of the population (as measured by population trend and variability); and 4) connectivity (as represented by the migratory life history form and functional habitat). For each element, the JDRRUT classified bull trout into relative risk categories.

For the population element, the team assigned the risk category of “diminished risk,” which is the lowest risk category. For the adult abundance element, the team assumed that abundance levels for migratory bull trout in individual local populations was below 100 spawners per year, and therefore are at risk of inbreeding depression. Similarly, the team concluded that the core area currently supported less than 1,000 migratory adults per year and consequently was at risk from genetic drift. For the productivity element, bull trout were classified at an increased risk, due to either the short duration of population census information, or the incomplete record of the redd count surveys within each core area. Finally, for the connectivity element, the team concluded that while the migratory form persists within the JDR core area, local populations are fragmented by habitat degradation, and the core area is considered at increased risk.

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### 5.3.6 LIMITING FACTORS

The 2002 DRP states that the main threats to bull trout persistence are habitat fragmentation and degradation, passage barriers that isolate populations, competition and predation from nonnative fishes, angling mortality, and effects resulting from isolation and small population sizes. The recovery strategy is to restore habitats and connectivity, reduce effects of nonnative fishes, and reduce angling mortality.

Livestock grazing is identified as a specific threat for listing factor A (the present or threatened destruction, modification or curtailment of bull trout habitat or range) in Chapter 1, Table 2 of the DRP. Categories of recovery actions identified to address livestock grazing in Table 2 that are germane to the proposed action include:

- Maintain or improve water quality in bull trout core areas or potential core habitat.
- Identify impaired stream channel and riparian areas and implement tasks to restore their functions.
- Design and implement a standardized monitoring program to assess the effectiveness of recovery efforts affecting bull trout and their habitats.

- Conduct evaluations of the adequacy and effectiveness of current and past best management practices in maintaining or achieving habitat conditions conducive to bull trout recovery.
- Use partnerships and collaborative processes to protect, maintain, and restore functioning core areas for bull trout.

Reasons for the decline of bull trout in the John Day Basin are presented in a draft summary of the John Day Subbasin Plan (NPPC 2001). Two overriding factors influencing fisheries and fish habitat are land use practices that have altered “the storage, movement and character of water resources over entire areas of the JDR subbasin and its tributary system,” and these prevalent land uses in combination with altered hydrologic responses are translated into stream channel instability in many area streams. The Oregon Department of Environmental Quality (1998) identified all streams inhabited by bull trout in the MFJD system (MFJDR, Big Creek, Granite Boulder Creek, and Clear Creek) as water quality limited, primarily for high summer temperatures, but also flow modification of the river.

Degraded fish habitat conditions are identified as occurring in approximately 966 kilometers (600 miles) of stream, due to erosion and sedimentation which reduce pool habitat, alter hydrographs, and result in loss of instream habitat elements. Although a variety of factors are at play in each of the major stream segments (south, north, middle, and upper mainstem), the results have been similar across the basin.

Buchanan *et al.* (1997) indicates other limiting factors including: (1) chemical mine waste, (2) reduction in anadromous fish populations, (3) past opportunities for over-harvest and poaching, and (4) hybridization and competition with brook trout.

Key habitat and population elements identified in Middle Fork JDR subbasin biological assessments, of concern to bull trout, include isolated populations, high summer water temperatures, substrate embeddedness, and high road densities (MNF 1998a, 1998b and 1999).

Chapter 9 of the DRP discusses reasons for decline (limiting factors). They include mainstem Columbia River dams, forest management practices, livestock grazing, agricultural practices, the transportation network, mining, residential development, recreation and isolation and habitat fragmentation due to seasonally inadequate water quality and quantity in the mainstem river and tributaries, due to degraded riparian and stream habitat conditions. Other barriers include low head dams, diversions, and natural waterfalls (Claire and Gray 1993, Oregon Department of Fish and Wildlife *in litt.* 2000).

Livestock grazing is discussed in some detail and is summarized here. Historic, and to a lesser degree, current grazing practices have removed extensive amounts of riparian vegetation that help stabilize stream channels and provide essential shade and cover to streams throughout the entire basin (ODFW, *in litt.* 2000). In many areas native bunch grasses have been replaced with invasive nonnative vegetation such as cheat grass and medusa head wild rye (ODFW, *in litt.* 2000). These nonnative grasses are more susceptible to frequent fires and retard the growth of native species by reducing the amount of water available during critical growing periods (ODFW, *in litt.* 2000).

High intensity livestock grazing and agricultural development has contributed to aquatic habitat degradation. Local areas may experience concentrations of livestock sufficient to damage streambanks and degrade habitat quality (ODFW, *in litt.* 2000). Grazing on private land varies



widely, but oftentimes shows ongoing stream bank damage from livestock (ODFW, *in litt.* 2000).

Livestock grazing occurs over much of the area along the Middle Fork JDR and its tributaries, and during the hot season, livestock tend to congregate near water sources, often leading to damaged riparian zones. This is particularly true on private lands adjacent to the river and tributaries where livestock are allowed to graze during summer and fall months. Several areas of the Middle Fork JDR lack adequate riparian vegetation and shrubs necessary to prevent bank erosion and heating of water (ODFW, *in litt.* 2000). The absence of shrubs and deciduous trees in meadows along the upper reaches of the Middle Fork has been attributed to summer long grazing (ODFW, *in litt.* 2000).

A Qualitative Habitat Assessment (QHA) modeling tool was used to assess bull trout habitat in the John Day Subbasin for the revised draft Subbasin Plan (NPCC 2005). For QHA modeling, a reach system consisting of 61 reaches was developed by the John Day fisheries technical team. The reach system encompassed all streams that bull trout presently inhabit, or are believed to formerly inhabit. The QHA model determined which attributes are most important in each geographic area in terms of limiting bull trout production. For the Middle Fork, of 11 attributes evaluated, the top five habitat limiting factors in order of importance are: riparian condition; high water temperature; fine sediment; habitat diversity and channel stability.

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## 5.4 CRITICAL HABITAT FOR BULL TROUT DPS

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### 5.4.1 DESIGNATION HISTORY

A final rule to designate CH for Columbia River bull trout was published in the Federal Register on October 18, 2010 (75 CFR 63898) (FWS 2010). The effective date of the final rule was November 17, 2010. Figure 2 displays streams designated as CH for bull trout within the Middle Fork JDR subbasin.

When the FWS identifies CH in occupied areas, it also assesses whether the PCEs may require any special management considerations or protection. In the final rule the FWS identified livestock grazing as one of the primary land and water management activities impacting the physical or biological features essential to the conservation of bull trout that may require special management considerations. The FWS stated that improper livestock grazing can promote streambank erosion and sedimentation and limit the growth of riparian vegetation important for temperature control, streambank stability, fish cover, and detrital input (Platts 1991, pp. 397–399). In addition, grazing often results in increased organic nutrient input in streams (Platts 1991, p. 423). These activities can directly and immediately threaten the integrity of the essential physical or biological features described in PCEs 1 through 8. Special management could include best management practices specifically designed to reduce these types of impacts in streams with bull trout, such as fencing livestock from stream sides, moving animal feeding operations away from surface waters, using riparian buffer strips near crop fields, minimizing water withdrawal from streams, avoiding stream channel and spring head alteration, and avoiding stream dewatering.

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### 5.4.2 PRIMARY CONSTITUENT ELEMENTS

The FWS identified nine PCEs of bull trout CH in the 2010 final rule. Table 15 lists the nine

PCEs.

TABLE 15. THE NINE PRIMARY CONSTITUENT ELEMENTS OF BULL TROUT CH THAT POTENTIALLY ARE AFFECTED BY LIVESTOCK GRAZING.

PCE	Description
1	Springs, seeps, groundwater sources, and subsurface water connectivity (hyporheic flows) to contribute to water quality and quantity and provide thermal refugia.
2	Migration habitats with minimal physical, biological, or water quality impediments between spawning, rearing, overwintering, and freshwater and marine foraging habitats, including but not limited to permanent, partial, intermittent, or seasonal barriers.
3	An abundant food base, including terrestrial organisms of riparian origin, aquatic macroinvertebrates, and forage fish.
4	Complex river, stream, lake, reservoir, and marine shoreline aquatic environments and processes that establish and maintain these aquatic environments, with features such as large wood, side channels, pools, undercut banks and unembedded substrates, to provide a variety of depths, gradients, velocities, and structure.
5	Water temperatures ranging from 2 to 15° C (36 to 59 ° F), with adequate thermal refugia available for temperatures that exceed the upper end of this range. Specific temperatures within this range will vary depending on bull trout life-history stage and form; geography; elevation; diurnal and seasonal variation; shading, such as that provided by riparian habitat; and local groundwater influence.
6	In spawning and rearing areas, substrate of sufficient amount, size, and composition to ensure success of egg and embryo overwinter survival, fry emergence, and young-of-the-year and juvenile survival. A minimal amount of fine sediment, generally ranging in size from silt to coarse sand, embedded in larger substrates, is characteristic of these conditions. The size and amounts of fine sediment suitable to bull trout will likely vary from system to system.
7	A natural hydrograph, including peak, high, low, and base flows within historic and seasonal ranges or, if flows are controlled, minimal flow departure from a natural hydrograph.
8	Sufficient water quality and quantity such that normal reproduction, growth, and survival are not inhibited.
9	Sufficiently low levels of occurrence of nonnative predatory (e.g., lake trout, walleye, northern pike, smallmouth bass); interbreeding (e.g., brook trout); or competing (e.g., brown trout) species that, if present, are adequately temporally and spatially isolated from bull trout.

### 5.4.3 STATUS OF BULL TROUT CRITICAL HABITAT

The limiting factors discussion presented earlier in Section 5.3.6 describes the condition of CH throughout the JDR Basin. A brief summary is that stream flows have been reduced by withdrawals and their timing has changed, water quality has been reduced (sedimentation and water temperature increases), stream substrates have become embedded, stream channels have been altered and riparian vegetation has been impacted with a resultant loss of shade canopy and large woody debris potential.

The QHA for bull trout provided a ranking of 61 JDR Basin stream reaches for both habitat protection and habitat restoration (NPPC 2005). Of those with extant populations in the Middle Fork, Granite Boulder, Big Creek reach 2 and Clear Creek were in the top quartile for protection benefit, while Big Creek reach 1 was in the second quartile. Big Creek reach 1 and Clear Creek were in the second quartile for restoration benefit, while Big Creek reach 2 was in the third quartile and Granite Boulder was in the fourth quartile.

## 6 ENVIRONMENTAL BASELINE

As mentioned in earlier sections, the predominant land use activity in the action area is livestock grazing for which there have been MNF formal and informal ESA consultations. The past, present and anticipated impacts of future Federal livestock grazing which have undergone formal consultation have been taken into account in the following description of the environmental baseline.

A NMFS process paper titled “Making Endangered Species Act Determinations of Effect for Individual or Grouped Actions at the Watershed Scale” (NMFS 1996) and A Framework to Assist In Making Endangered Species Act Determinations of Effect for Individual or Grouped Actions at the Bull Trout Subpopulation Watershed Scale (USFWS 1998) is used to describe the environmental baseline. It is commonly known as the NMFS and or FWS Matrix of Pathways and Indicators, hereafter referenced as the “NMFS& FWS MPI.”

The NMFS MPI identifies indicators to analyze for the following pathways: 1) Water quality; 2) Habitat access; 3) Habitat elements; 4) Channel condition and dynamics; 5) Flow/hydrology; and, 6) Watershed condition. The condition of each indicator is described as either “Properly Functioning” (PF), “At Risk (AR),” or “Not Properly Functioning (NPF)” based upon specific numeric or qualitative criteria.

The FWS MPI identifies indicators to analyze for the following pathways: 1) Subpopulation Characteristics within subpopulation watersheds; 2) Water quality; 3) Habitat access; 4) Habitat elements; 5) Channel condition and dynamics; 6) Flow/hydrology; 7) Watershed condition; and 8) Integration of Species and Habitat Conditions.

Table 16 presents the current status of the environmental baseline for the Middle Fork John Day River sub-basin, which includes the action area, utilizing the NMFS and FWS MPI. Table cells in bold print indicate the current status of each indicator. The habitat indicators in the NMFS and FWS matrix also correspond to the PCEs of designated CH. The relationship between NMFS and FWS MPI habitat indicators and PCEs of CH is discussed in Section 7.2 (Analysis of Effects to Designated Critical Habitat).

TABLE 16. MCR STEELHEAD AND COLUMBIA RIVER BULL TROUT STATUS OF ENVIRONMENTAL BASELINE FOR THE MIDDLE FORK JOHN DAY SUB-BASIN.1

Pathway	Indicators	Properly Functioning	At Risk	Not Properly Functioning
<b>Species: Specific for Bull Trout</b>				
Subpopulation Characteristics within subpopulation watersheds	Subpopulation Size	Mean total subpopulation size or local habitat capacity more than several thousand individuals. All life stages evenly represented in the subpopulation.	Adults in subpopulation are less than 500 but >50.	<b>Adults in subpopulation has less than 50.</b>
	Growth and Survival	Subpopulation has the resilience to recover from short term disturbances (e.g. catastrophic events, etc) or subpopulation declines within one to two generations (5 to 10 years). The subpopulation is characterized as increasing or stable. At least 10+ years of data support this estimate.	When disturbed, the subpopulation will not recover to predisturbance conditions within one generation (5 years). Survival or growth rates have been reduced from those in the best habitats. The subpopulation is reduced in size, but the reduction does not represent a long-term trend. least 10+ years of data support this characterization.2 If less data is available and a trend cannot be confirmed, a subpopulation will be considered at risk until enough data is available to accurately determine its trend.	<b>The subpopulation is characterized as in rapid decline or is maintaining at alarmingly low numbers. Under current management, the subpopulation condition will not improve within two generations (5 to 10 years). This is supported by a minimum of 5+ years of data.</b>

Pathway	Indicators	Properly Functioning	At Risk	Not Properly Functioning
	Life History Diversity and Isolation	The migratory form is present and the subpopulation exists in close proximity to other spawning and rearing groups. Migratory corridors and rearing habitat (lake or larger river) are in good to excellent condition for the species. Neighboring subpopulations are large with high likelihood of producing surplus individuals or straying adults that will mix with other subpopulation groups.	The migratory form is present but the subpopulation is not close to other subpopulations or habitat disruption has produced a strong correlation among subpopulations that do exist in proximity to each other.	<b>The migratory form is absent and the subpopulation is isolated to the local stream or a small watershed not likely to support more than 2,000 fish.<sup>1</sup></b>
	Persistence and Genetic Integrity	Connectivity is high among multiple (5 or more) subpopulations with at least several thousand fish each. Each of the relevant subpopulations has a low risk of extinction. The probability of hybridization or displacement by competitive species is low to nonexistent.	<b>Connectivity among multiple subpopulations does occur, but habitats are more fragmented. Only one or two of the subpopulations represent most of the fish production. The probability of hybridization or displacement by competitive species is imminent, although few documented cases have occurred.</b>	Little or no connectivity remains for refounding subpopulations in low numbers, in decline, or nearing extinction. Only a single subpopulation or several local populations that are very small or that otherwise are at high risk remain. Competitive species readily displace bull trout. The probability of hybridization is high and documented cases have occurred.
<b>Habitat: MCR Steelhead and Bull Trout</b>				
Water Quality	Temperature	MCR Steelhead 50 – 57° F (max 7-day average)	MCR Steelhead 57 – 61° F (spawning, max 7-day average)	<b>MCR Steelhead &gt; 61° F (spawning, max 7-day average)</b>

Pathway	Indicators	Properly Functioning	At Risk	Not Properly Functioning
		<p>Bull Trout</p> <p>7 day average maximum temperature in a reach during the following life history stages:</p> <p>incubation 2 - 5 °C</p> <p>rearing 4 - 12 °C</p> <p>spawning 4 - 9 °C</p> <p>also temperatures do not exceed 15°C in areas used by adults during migration (no thermal barriers)</p>	<p>57 – 64° F (migration and rearing, max 7-day average)</p> <p>Bull Trout</p> <p>7 day average maximum temperature in a reach during the following life history stages:</p> <p>incubation &lt;2°C or 6 °C</p> <p>rearing &lt;4 °C or 13 - 15 °C</p> <p>spawning &lt;4 or 10 °C</p> <p>also temperatures in areas used by adults during migration sometimes exceeds 15°C</p>	<p><b>&gt; 64° F (migration and rearing, max 7-day average)</b></p> <p><b>Bull Trout</b></p> <p><b>7 day average maximum temperature in a reach during the following life history stages:</b></p> <p><b>incubation &lt;1°C or &gt;6 °C</b></p> <p><b>rearing &gt;15 °C</b></p> <p><b>spawning &lt;4 or &gt;10 °C</b></p> <p><b>also temperatures in areas used by adults during migration regularly exceed 15°C (thermal barriers present)</b></p>
	Sediment	< 12% fines (<0.85mm) in gravel	<b>12 – 20% fines</b>	> 20% fines
	Chemical Contaminants or Nutrients	Low levels of chemical contamination from agricultural, industrial, and other sources; no excess nutrients; no CWA 303d designated reaches	<b>Moderate levels of chemical contamination from agricultural, industrial, and other sources; some excess nutrients; one CWA 303d designated reach</b>	High levels of chemical contamination from agricultural, industrial, and other sources; high levels of excess nutrients; more than one CWA 303d designated reach

Pathway	Indicators	Properly Functioning	At Risk	Not Properly Functioning
Habitat Access	Physical Barriers	Any man-made barriers present in watershed allow upstream and downstream fish passage at all flows	Any man-made barriers present in watershed do not allow upstream and/or downstream fish passage at base/low flows	<b>Any man-made barriers present in watershed do not allow upstream and/or downstream fish passage at a range of flows</b>
Habitat Elements	Substrate	Dominant substrate is gravel or cobble (interstitial spaces clear), or embeddedness <20%	Gravel and cobble is subdominant, or if dominant, embeddedness 20 – 30%	<b>Bedrock, sand, silt, or small gravel dominant, or if gravel and cobble dominant, embeddedness &gt;30%</b>
	Large Woody Debris	<p>MCR Steelhead</p> <p>&gt; 20 pieces/mile (&gt; 12 inch diameter and &gt; 35 ft length), and adequate sources of woody debris recruitment in riparian areas</p> <p>Bull Trout</p> <p>current values are being maintained at greater than 80 pieces/mile that are &gt;24" diameter and &gt;50 ft length on the Coast 9, or &gt;20 pieces/ mile &gt;12" diameter &gt;35 ft length on the Eastside also adequate sources of woody debris are available for both long and short-term recruitment</p>	<b>Currently meets standards for Properly Functioning, but lacks potential sources from riparian areas of woody debris recruitment to maintain that standard</b>	Does not meet standards for Properly Functioning and lacks potential large woody debris recruitment

Pathway	Indicators	Properly Functioning	At Risk	Not Properly Functioning																				
	Pool Frequency	<p>MCR Steelhead</p> <p>Meets pool frequency standards and meets large woody debris recruitment standards for Properly Functioning habitat</p> <p>Bull Trout</p> <p>Pool frequency in a reach closely approximates:</p> <table><tr><td>Wetted width (ft)</td><td>#pools/mile</td></tr><tr><td>0-5</td><td>39</td></tr><tr><td>5-10</td><td>60</td></tr><tr><td>10-15</td><td>48</td></tr><tr><td>15-20</td><td>39</td></tr><tr><td>20-30</td><td>23</td></tr><tr><td>30-35</td><td>18</td></tr><tr><td>35-40</td><td>10</td></tr><tr><td>40-65</td><td>9</td></tr><tr><td>65-100</td><td>4</td></tr></table>	Wetted width (ft)	#pools/mile	0-5	39	5-10	60	10-15	48	15-20	39	20-30	23	30-35	18	35-40	10	40-65	9	65-100	4	<p>MCR Steelhead</p> <p>Meets pool frequency standards but large woody debris recruitment inadequate to maintain pools over time</p> <p>Bull Trout</p> <p>Pool frequency is similar to values in “functioning appropriately”, but pools have inadequate cover/temperature, and/or there has been a moderate reduction of pool volume by fine sediment.</p>	<p><b>MCR Steelhead</b></p> <p><b>Does not meet pool frequency standards</b></p> <p><b>Bull Trout</b></p> <p><b>Pool frequency is considerably lower than values desired for “functioning appropriately”; also over/temperature is inadequate and there has been a major reduction of pool volume by fine sediment.</b></p>
Wetted width (ft)	#pools/mile																							
0-5	39																							
5-10	60																							
10-15	48																							
15-20	39																							
20-30	23																							
30-35	18																							
35-40	10																							
40-65	9																							
65-100	4																							
	Pool Quality	<p>Pools &gt; 1 meter deep (holding pools) with good cover and cool water; minor reduction of pool volume by fine sediment</p>	<p>Few deeper pools (&gt; 1 meter) present or inadequate cover/ temperature; moderate reduction of pool volume by fine sediment</p>	<p><b>No deep pools (&gt; 1 meter) and inadequate cover/temperature; major reduction of pool volume by fine sediment</b></p>																				



Pathway	Indicators	Properly Functioning	At Risk	Not Properly Functioning
	Large Pools	Bull Trout each reach has many large pools >1 meter deep	Bull Trout reaches have few large pools (>1 meter) present	<b>Bull Trout reaches have no deep pools (&gt;1 meter)</b>
	Off Channel Habitat	Backwaters with cover, and low energy off-channel areas (ponds, oxbows, etc.)	<b>Some backwaters and high energy side channels</b>	Few or no backwaters; no off-channel ponds
	Refugia	Habitat refugia exist and are adequately buffered (e.g., by intact riparian reserves); existing refugia are sufficient in size, number, and connectivity to maintain viable populations or subpopulations (all life stages and forms)	<b>Habitat refugia exist but are not adequately buffered (e.g., by intact riparian reserves); existing refugia are insufficient in size, number, and connectivity to maintain viable populations or subpopulations (all life stages and forms)</b>	Adequate habitat refugia do not exist
	Average Wetted Width/Maximum Depth Ratio in scour pools in a reach	Bull Trout <10	Bull Trout 11 - 20	<b>Bull Trout &gt;20</b>
Channel Condition & Dynamics	Width/Depth Ratio	MCR Steelhead < 10	MCR Steelhead 10 – 12	<b>MCR Steelhead &gt; 12</b>
	Stream Bank Condition	> 80% of any stream reach has > 90% stability	<b>50 – 80% of any stream reach has &gt; 90% stability</b>	< 50% of any stream reach has > 90% stability

Pathway	Indicators	Properly Functioning	At Risk	Not Properly Functioning
	Floodplain Connectivity	Off-channel areas are frequently hydrologically linked to main channel; overbank flows occur and maintain wetland functions, riparian vegetation, and succession	<b>Reduced linkage of wetland, floodplains, and river areas to main channel; overbank flows are reduced relative to historic frequency, as evidenced by moderate degradation of wetland function and riparian vegetation/succession</b>	Severe reduction in hydrologic connectivity between off-channel, wetland, floodplain, and riparian areas; wetland extent drastically reduced, and riparian vegetation/success altered significantly
Flow/Hydrology	Change in Peak/Base Flows	Watershed hydrograph indicates peak flow, base flow, and flow timing characteristics comparable to an undisturbed watershed of similar size, geology, and geography	<b>Some evidence of altered peak flow, baseflow, and/or flow timing relative to an undisturbed watershed of similar size, geology, and geography</b>	Pronounced changes in peak flow, baseflow, and/or timing relative to an undisturbed watershed of similar size, geology, and geography
	Increase in Drainage Network	<p>MCR Steelhead</p> <p>Zero or minimum increases in drainage network density due to roads</p> <p>Bull Trout</p> <p>zero or minimum increases in active channel length correlated with human caused disturbance</p>	<p>MCR Steelhead</p> <p>Moderate increases in drainage network density due to roads (e.g., 5%)</p> <p>Bull Trout</p> <p>low to moderate increase in active channel length correlated with human caused disturbance</p>	<p><b>MCR Steelhead</b></p> <p><b>Significant increases in drainage network density due to roads (e.g., 20 – 25%)</b></p> <p><b>Bull Trout</b></p> <p><b>greater than moderate increase in active channel length correlated with human caused disturbance</b></p>

Pathway	Indicators	Properly Functioning	At Risk	Not Properly Functioning
Watershed Condition	Road Density & Location	<p>MCR Steelhead &lt; 2 mi/mi<sup>2</sup>; no Valley bottom roads</p> <p>Bull Trout &lt;1mi/mi<sup>2</sup> ; no valley bottom roads</p>	<p>MCR Steelhead 2 – 3 mi/mi<sup>2</sup>; some valley bottom roads</p> <p>Bull Trout 1 - 2.4 mi/mi<sup>2</sup>; some valley bottom roads</p>	<p><b>MCR Steelhead</b> <b>&gt; 3 mi/mi<sup>2</sup>; many valley bottom roads</b></p> <p><b>Bull Trout</b> <b>&gt;2.4 mi/mi<sup>2</sup> ; many valley bottom roads</b></p>
	Disturbance History	< 15% ECA (entire watershed) with no concentration of disturbance in unstable or potentially unstable areas, and/or refugia, and/or riparian areas	<b>&lt; 15% ECA (entire watershed) but disturbance concentrated in unstable or potentially unstable areas, and/or refugia, and/or riparian areas</b>	> 15% ECA (entire watershed) and disturbance concentrated in unstable or potentially unstable areas, and/or refugia, and/or riparian areas
	Riparian Management Areas	The riparian reserve system provides adequate shade, large woody debris recruitment, and habitat protection and connectivity in all subwatershed, and buffers or includes known refugia for sensitive aquatic species (>80% intact), and/or for grazing impacts; percent similarity of riparian vegetation to the potential natural community/ composition > 50%	<b>Moderate loss of connectivity or function (shade, LWD recruitment, etc.) of riparian reserve system, or incomplete protection of habitats and refugia for sensitive aquatic species (~ 70 – 80% intact), and/or for grazing impacts; percent similarity of riparian vegetation to the potential natural community/ composition 25 – 50% or better</b>	Riparian reserve system is fragmented, poorly connected, or provides inadequate protection of habitats and refugia for sensitive aquatic species (< 70% intact), and/or for grazing impacts; percent similarity of riparian vegetation to the potential natural community/ composition < 25%

Pathway	Indicators	Properly Functioning	At Risk	Not Properly Functioning
	Disturbance Regime	Environmental disturbance is short lived; predictable hydrograph, high quality habitat and watershed complexity providing refuge and rearing space for all life stages or multiple life-history forms. Natural processes are stable.	<b>Scour events, debris torrents, or catastrophic fire are localized events that occur in several minor parts of the watershed. Resiliency of habitat to recover from environmental disturbances is moderate.</b>	Frequent flood or drought producing highly variable and unpredictable flows, scour events, debris torrents, or high probability of catastrophic fire exists throughout a major part of the watershed. The channel is simplified, providing little hydraulic complexity in the form of pools or side channels. Natural processes are unstable.
<b>Species and Habitat: Bull Trout</b>				
Integration of Species and Habitat Conditions		Habitat quality and connectivity among subpopulations is high. The migratory form is present. Disturbance has not altered channel equilibrium. Fine sediments and other habitat characteristics influencing survival or growth are consistent with pristine habitat. The subpopulation has the resilience to recover from short-term disturbance within one to two generations (5 to 10 years). The subpopulation is fluctuating around an equilibrium or is growing. <sup>1</sup>	Fine sediments, stream temperatures, or the availability of suitable habitats have been altered and will not recover to predisturbance conditions within one generation (5 years). Survival or growth rates have been reduced from those in the best habitats. The subpopulation is reduced in size, but the reduction does not represent a long-term trend. The subpopulation is stable or fluctuating in a downward trend. Connectivity among subpopulations occurs but habitats are more fragmented. <sup>1</sup>	<b>Cumulative disruption of habitat has resulted in a clear declining trend in the subpopulation size.</b> <b>Under current management, habitat conditions will not improve within two generations (5 to 10 years). Little or no connectivity remains among subpopulations. The subpopulation survival and recruitment responds sharply to normal environmental events. 1</b>

The environmental baseline using the NMFS MPI ratings (Table 16) is based on scientific literature review, management documents and the professional judgment of MNF Forest and District fishery biologists, hydrologists, soil scientists and range conservationists. The MCR Steelhead Recovery Plan, the Malheur National Forest Roads Analysis Report, and the Forest water temperature monitoring program support the environmental baseline ratings. The rationale from the supporting documents and programs for these ratings are summarized in the following three sections: 6.1.1; 6.1.2; and 6.1.3.

Historic and current cattle grazing in the ESA action area likely play varying roles in the current environmental baseline ratings for these affected subbasins. In some situations the actual small streams and corresponding 6th field subwatersheds draining the ESA action area may be properly functioning or functioning at risk while the larger subbasins are not properly functioning. Grazing is one of multiple natural and human-caused watershed disturbances influencing environmental baseline ratings. In some circumstances in these subbasins the proposed action has causal mechanisms that affect fish habitat indicators analyzed in the environmental baseline ratings.

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#### 6.1.1 MIDDLE COLUMBIA RIVER STEELHEAD RECOVERY PLAN

The Middle Columbia River Steelhead Recovery Plan (NMFS 2009) describes habitat conditions for the UMJD and SFJD River populations. The recovery plan states that there has been a loss of off-channel and side-channel habitats. Consequently there is reduced spawning and rearing habitat and a loss of refugia from high flows. Floodplain connectivity has been degraded. Removal of beavers has reduced habitat complexity, floodplain function and the amount of stored water. Livestock grazing has increased channelization and negatively impacted large wood debris (LWD), cover and bank stability.

More than 50 percent of measured streams had greater than 35 percent substrate embeddedness (NMFS 2009). Embeddedness is also a problem in 5 of 8 Canyon Creek reaches and 10 of 32 mainstem John Day River reaches (MNF 2004). Livestock grazing, logging, road construction, beaver removal and fire suppression have contributed to altered flow regimes. Water withdrawals dewater the lower reach of McClellan Creek. Elevated stream temperatures are common to almost all streams in the Upper John Day. Historical mining has added to water temperature problems by removing riparian vegetation, simplifying stream channels and changing substrate composition (NMFS 2009). High water temperatures in tributaries and portions of the mainstem alter or block juvenile steelhead movements in the summer months.

Riparian conditions are degraded. Roads have altered riparian functions. The MCR steelhead Recovery Plan states that the MNF identified about 124 miles of roads within Riparian Habitat Conservation Areas (MNF 2004). Forest practices and grazing on private and public land have altered riparian vegetation and LWD potential has also been reduced.

There are additional sources of information to describe the environmental baseline at finer scales than the UMJD River population. They include a MNF roads analysis report, water temperature monitoring information, PIBO EM results, Forest Service level II stream inventories, annual end-of-year grazing reports for the years 2008-2010 (MNF 2009a, MNF 2009b, MNF 2011), Proper Functioning Condition (PFC) assessments (Prichard et al. 1994) and Multiple Indicator

Monitoring (MIM) (Burton et al. 2011) monitoring results at Designated Monitoring Areas (DMA) for specific pastures in the allotments for years ranging from 2008 to 2010. The information provided by each of these sources is presented and interpreted below.

### 6.1.2 MALHEUR NATIONAL FOREST ROADS ANALYSIS REPORT

The MNF prepared an analysis of its road system in a document titled “Malheur National Forest Roads Analysis Report” (MNF 2004). Among the issues analyzed was the risk of the existing road network to general watershed health at the scale of 6<sup>th</sup> level hydrologic unit codes (HUC), commonly known as sub-watersheds. Many scientific studies have documented the impacts of roads on to fish, fish habitat, and watershed function. Effects include habitat fragmentation from stream crossing structures that block migration, increases in peak flows from high road density, increased sedimentation and isolating streams from their floodplains (USDA FS 2001). The MNF used a Geographic Information System assessment to determine watershed risk. The following watershed risk rating elements were used:

- Total road density (roads in management levels 1-5)
- Road density (roads in management levels 1 and 2)
- Total road density within 200 feet of perennial and intermittent streams
- Density within 200 feet of perennial and intermittent streams (roads in management levels 1 and 2)
- Total road-stream crossing density (crossings/square mile)
- Geologic Sensitivity
- Soil Sensitivity

Ranges of values for each element were assigned a risk rating of low, moderate, high or extreme (Table 17). For example, for total road density, an “extreme” risk rating was for densities greater than five miles per square mile and a “low” rating was for densities less than one mile per square mile. To determine the overall subwatershed risk rating, the risk rating for each element was assigned a numeric value. They ranged from 1 for a rating of “low” to 4 for a rating of “extreme.” The individual element numeric scores were then added for a total score. Total scores exceeding 23 were given an overall watershed risk rating of “extreme,” scores in the 17-23 range were given a “high” rating, scores from 11-17 were given a “moderate” rating and scores less than 11 were given a rating of “low.”

TABLE 17. RANGES OF VALUES BY RISK CATEGORY FOR ELEMENTS USED IN THE WATERSHED RISK ANALYSIS.

Risk Element	Risk Category			
	Low	Moderate	High	Extreme
Total Road				

Risk Element	Risk Category			
	Low	Moderate	High	Extreme
Density (miles/mile <sup>2</sup> )	0 - 1	1 - 3	3 - 5	>5
Level 1-2 Road Density (miles/mile <sup>2</sup> )	0 - 1	1 – 2.5	2.5 – 4	>4
Road Density w/in 200 feet of streams (miles/mile <sup>2</sup> )	0 – 0.2	0.2 – 0.6	0.6 – 0.9	>0.9
Level 1-2 Road Density w/in 200 feet of streams (miles/mile <sup>2</sup> )	0 – 0.2	0.2 – 0.5	0.5 – 0.8	>0.8
Road Stream Crossing Density (#crossings/mile)	0 – 1.5	1.5 - 3	3 – 4.5	>4.5
Percent of Subwatershed with Sensitive Geology	1 - 20	20 – 50	50 – 100	Not applicable
Percent of Subwatershed with Sensitive Soils	1 - 20	20 – 50	50 – 100	Not applicable

The results for the two 6<sup>th</sup> field HUCs represented in the action area for this consultation are shown in Table 18. Bear Creek and Slide Creek were given a “high” watershed risk rating. The individual risk ratings for the road density, road proximity within 200 feet, and road crossings elements were “high” or “extreme” for these two 6<sup>th</sup> fields. These ratings suggest that the legacy road system has adversely impacted riparian and aquatic environmental baseline conditions in these two subwatersheds.

The NMFS MPI values for the Road Density and Location (RDL) indicator are  $<2$ , 2-3 miles/mile<sup>2</sup> and  $>3$  miles/mile<sup>2</sup> for the PF, AR and NPF categories, respectively. The “high” and “extreme” risk ratings for both road density risk elements are equivalent to the NMFS MPI NPF category. Therefore, these two 6<sup>th</sup> fields are NPF for RDL.

While there are no other roads risk analysis elements that are directly comparable to NMFS MPI indicators, it is logical that the Road Stream Crossing Density (RSCD) watershed risk element would inform an analysis of the NMFS MPI Increase in Drainage Network (IDN) indicator. The “high” or “extreme” risk score for the RSCD risk element for these two 6<sup>th</sup> field HUCs support a NMFS MPI classification of NPF for the IDN indicator.

Road crossings at streams are the primary mechanism for rainfall runoff intercepted by roads to enter stream channels. Roads tend to concentrate runoff, resulting in higher peak flows than would occur without roads. Fine sediments from road surfaces also enter stream channels at road crossings, increasing turbidity, substrate embeddedness and substrate composition. The “extreme” or “high” risk ratings for these two 6<sup>th</sup> field HUCs would logically support classification of NPF for the Change in Peak/Base Flows, Sediment and Substrate NMFS MPI indicators.

The vast majority of road crossings at streams are culverts. Poorly designed culverts can be barriers to juvenile or adult fish passage. The RSCD risk scores do not incorporate fish passage barrier information, but high or extreme risk ratings imply a large number of culverts with potential fish passage problems. The MNF has identified road crossing structures that may have fish passage concerns from Level II stream inventory reports. Three culverts with potential fish passage problems were identified in Mosquito Creek within the ESA action area. The culverts that may have fish passage concerns are located in the Bear Allotment, pasture H. No other streams had identified fish passage barriers within the action area.

The two risk elements for road density within 200 feet of streams do not have a comparable NMFS MPI indicator. However, roads within floodplains have the potential to negatively affect the Off-channel Habitat and Floodplain Connectivity NMFS MPI indicators. Many FS roads are in the valley bottoms, in or adjacent to riparian areas, and affect the ability of a stream to meander laterally through its floodplain. There is no information to determine to what degree a distance of 200 feet includes the floodplains for the various streams associated with the road system in the two 6<sup>th</sup> field HUCs represented in the York and Bear allotments. However, the “extreme” risk ratings for the two risk elements for the two 6<sup>th</sup> field HUCs support a NMFS MPI classification of NPF for the two NMFS MPI indicators.



TABLE 18. SUB-WATERSHED RISK RATINGS FOR SIXTH FIELD HYDROLOGIC UNITS IN THE ESA ACTION AREA FOR THE YORK AND BEAR ALLOTMENTS CONSULTATION (FROM MNF 2004B).

Watershed Risk Element	Hydrologic Unit Code Name and Number	
	Bear Creek-Middle Fork John Day River <sup>2</sup> 170702030301	Slide Creek 170702030304
Road Density Risk (ML <sup>1</sup> 1-5)	High	High
Road Density Risk (ML 1-2)	Extreme	Extreme
Road 200' Proximity Risk (ML 1-5)	Extreme	Extreme
Road 200' Proximity Risk (ML 1-2)	Extreme	Extreme
Road Crossings Risk	High	Extreme
Geologic Sensitivity	Moderate	Low
Soil Erosion Sensitivity	Low	Low
Overall Watershed Risk	High	High

<sup>1</sup>ML = Road Maintenance Level (see footnote)

<sup>2</sup>Previous HUC6 name used during roads analysis was Mosquito Creek/Bear Creek

Road maintenance level (ML) designations are defined as:

Level 1. These are intermittent service roads during the time they are closed to motorized traffic. The closure period must exceed one year. Basic custodial maintenance is performed to keep damage to adjacent resources to an acceptable level. Emphasis is normally given to maintaining drainage facilities and runoff patterns. Planned road deterioration may occur at this level. Roads receiving level 1 maintenance may be of any type, class, or construction standard.

Level 2. Roads open for use by high clearance vehicles. Passenger car traffic is not a consideration. Traffic is normally minor.

Level 3. Roads open and maintained for travel by a prudent driver in a standard passenger car. User comfort and convenience are not considered priorities. Roads in this maintenance level are typically low speed, single lane with turnouts and spot surfacing. Some roads may be fully surfaced with either native or processed material.

Level 4. Roads that provide a moderate degree of user comfort and convenience at moderate travel speeds. Most roads are double lane and aggregate surfaced. However, some roads may be single lane. Some roads may be paved and/or dust abated.

Level 5. Roads that provide a high degree of user comfort and convenience. These roads are normally double lane, paved facilities. Some may be aggregate surfaced and dust abated.

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### 6.1.3 MALHEUR NATIONAL FOREST WATER TEMPERATURE MONITORING

Appendix I presents water temperature monitoring information for Slide, Rice, and Mosquito creeks and the mainstem Middle Fork John Day River in the vicinity of the action area. Sites had 1-3 years of data and most were located in or directly adjacent to the Bear allotment. Four of the monitoring sites were located in MCR Steelhead CH, while the two sites on the Middle Fork John Day River were also in CR bull trout CH. The Middle Fork John Day River is on the Oregon Department of Environmental Quality 303(d) list for water temperature (ODEQ 2010).

For the Bear Allotment, the mean yearly maximum of seven day rolling means of the daily maximum in degrees Fahrenheit (7 day mean max) was Mosquito Creek 75.5 °F, upper Middle Fork John Day River 79.6 °F, lower Middle Fork John Day River 75.1 °F. The mean number of days per year over 64 degrees F was Mosquito Creek 70, upper Middle Fork John Day River 79, and lower Middle Fork John Day River 69.

For streams adjacent to but outside the York Allotment the mean yearly maximum of seven day rolling means of the daily maximum in degrees Fahrenheit (7 day mean max) was Slide Creek 67.4 °F, Rice Creek 64.4 °F. The number of days per year over 64 degrees F was Slide Creek 33, Rice Creek 7.

In the tables, data are evaluated using the following criteria: 1) State water quality standards; 2) Amendment 29 DFC; 3) PACFISH RMO; and, 4) NMFS/FWS MPI.

The state water quality standard of the seven-day mean maximum temperature of 64 °F for streams with anadromous fish passage and salmonid rearing use was *not met* for any of the sites. Likewise, the Amendment 29 DFC for Chinook salmon and/or westslope cutthroat trout spawning and rearing habitat of seven day mean maximum of 55 °F was *not met* for any site.

Cutthroat trout are not present in streams within the action area, but spring Chinook salmon use the mainstem Middle Fork John Day River for spawning, migration, and rearing (MNF corporate GIS files).

The PACFISH RMO has three criteria. Criterion 1 is no measurable increase in summer maximum temperature. Sites lacked adequate sample size to determine if a trend existed; therefore this criterion could not be evaluated. Criterion 2, seven-day mean maximum below 64 °F for migration and rearing habitat, was *not met* for Mosquito Creek and the Middle Fork John Day River in the allotments and Slide and Rice creeks outside the allotments. Criterion 3, seven-day mean maximum below 60 °F for spawning habitat, was *not met* for any stream. The data supported a NMFS MPI rating of Functioning at Unacceptable Risk (FAUR, equal to NPF) for streams in this area.

## 6.2 PIBO MONITORING

The PACFISH-INFISH Biological Opinion (PIBO) monitoring strategy is described in section 3.1.2.3. Monitoring consists of two components: effectiveness and implementation.

### 6.2.1 EFFECTIVENESS MONITORING

The following provides a summary of data collected by the PIBO Effectiveness Monitoring Program (EMP) for monitoring locations within the York On/Off Allotment. Sites included within the summary are either integrator<sup>8</sup> reaches within randomly selected watersheds and/or Designated Monitoring Area (DMA)<sup>9</sup> reaches that lie within pastures contained by the randomly selected watersheds. There has been only one reach evaluation (sites evaluated two or more times) within the allotment between 2005 and 2010. Within this monitored reach the data suggests that most stream attributes are improving, some are remaining relatively static, and some are declining (see Table 19 below).

TABLE 19. PIBO MONITORING RESULTS – YORK ON/OFF ALLOTMENT

Stream	Site Type	Year	Rosgen Stream Type	Bf W/D	Mean Part. Size (D50) (mm)	Pool % /Res. Pool depth (m)	%Fine s <2mm /<6mm	Bank Stab. (%)	Bank Angle (°)	Under-cut %	GL Wet Rat	GL Woody CV
Slide Cr.	I	2005	B4	11.9	61	53.1/	31.9/	98	124	12.8	57	77

<sup>8</sup> PIBO monitoring sites chosen within randomly selected subwatersheds to show integrated effects of upstream management—most are located in the most downstream response reach (stream gradient less than 3%), while the remaining at the downstream most transport reach (stream gradient between 3 and 5%)

<sup>9</sup> PIBO monitoring sites located within each randomly selected subwatershed where livestock grazing occurs within the riparian area; locations are selected by Field Units and used for annual implementation monitoring; objective is to develop link between implementation and effectiveness monitoring as part of adaptive management feedback process

						0.16	33.0					
Slide Cr.	I	2010	B4	8.2	50	44.6/ 0.16	6.9/ 10.3	98	114	33.3	61	55

In reviewing the data, it appears that the monitoring site on Slide Creek has shown improvement for most attributes. Overall, bankfull width-to-depth ratios as well as bank angle, undercut banks, and percent fines <6mm have shown considerable improvement while the remaining evaluated attributes, those considered to be potentially affected by livestock grazing – D50, pool %, residual pool depth, bank stability, greenline wetland rating, and greenline woody cover, remained relatively unchanged. It should be noted that besides riparian vegetation, the stream attributes most directly affected by grazing activities are bank stability, bank angle, width to depth ratio, and percent undercut banks. There are too few reference sites within the John Day Basin to determine whether similar changes are occurring in unmanaged watersheds. Although only bankfull width-to-depth ratios, bank angle, % undercut banks, and percent fines <6mm showed improvement, this is actually a promising indicator given that both Bengeyfield (2006) and Rosgen (1996) have indicated that the relationship between a stream's width and depth is perhaps the most revealing of all stream channel indicators as to whether the stream is in a condition to perform the various tasks that lead to a healthy riparian area.

While some attributes, such as channel shape and the frequent floodplain, are generally formed in 1.5-2 year events, others, such as habitat complexity, are formed during moderately high events of 10-25 year return intervals. If the stream cannot maintain its dimension, pattern, and profile during these moderately high events, then habitat or other desired values will probably not be created or sustained over time. Thus, given the short 5-year time frame between site visits for monitoring and developing trends, and the analysis of data, it is our reasoned opinion that the evaluated attributes are being maintained or showing a slight overall improvement. Although a few of the stream habitat attributes have shown little change, there has been improvement in bankfull width-to-depth ratios, bank angle, undercut banks, and percent fines <6mm, on federal lands of the Malheur National Forest (MNF) within the York On/Off Allotment.

These results align with those found for integrator and DMA reaches as summarized in Dr. Roper's 2009 declaration (ONDA v. Tidwell, Civ. No. 07-1871-HA, Docket #235). Based on his analysis he also stated that "it is my reasoned conclusion that the trend in the conditions of integrator reaches evaluated on the Malheur National Forest is similar or slightly better than those of other integrator reaches within the John Day Basin ... and it is my reasoned conclusion the trend in the DMA reaches within the Malheur National Forest are similar (or slightly better) than those of other DMA reaches within the John Day River Basin."

### **Evaluation of Existing Conditions to PIBO Managed and Reference Means**

Identifying the existing condition of streams within a particular watershed or management area is an important step in evaluating how land management may be affecting the quality of stream habitats. To help assess these conditions, we are utilizing recent information from the PIBO EMP to represent mean habitat conditions for both managed and reference conditions (see Table 20). As part of a recent publication, they developed an index of physical habitat conditions – 8 commonly collected metrics used in stream habitat monitoring – by evaluating the status of the

condition of 217 reference and 934 managed streams in the Interior Columbia River and Upper Missouri River Basins (Al-Chokhachy et al. 2010). Comparing our existing condition information to these values helps provide for the evaluation of management practices to determine if they are effective in maintaining the desired and/or proper functioning condition, or improving the structure and function of riparian and aquatic conditions. The values shown in Table 20 are not intended to replace current RMOs but to illustrate how specific streams on the MNF compare to other managed and reference stream sites within the Interior Columbia Basin.

We believe this evaluation is merited as new information/articles are calling for the re-examination of certain RMOs/RMO values. Kershner and Roper (2010) found that many of the PIBO EMP reference reaches did not meet RMOs, such as wetted width-to-depth, percent undercut banks, number of pieces of large wood, and numbers of days exceeding 15°C. These authors also stated that the current RMOs were originally designed as an early warning of potential negative effects of land management on stream/riparian conditions, and values that did not meet RMOs were thought to potentially represent unsuitable habitat conditions for important salmonids. Their analysis of data from federally-managed sites in the interior Columbia River basin indicates that the usefulness of RMOs may be questionable. In summary, they found that none of the 726 reference and managed reaches surveyed met all RMOs, and in a previous analysis (Henderson et al. 2005) found that only 2% of the reference reaches met the RMO for wetted width-to-depth ratio and that 16% met the reference criteria for percent undercut banks. These may be examples of RMOs warranting exclusion or reconsideration. Modifications in the selection of RMOs and their application are clearly needed if the results in this study are to represent conditions across the broader landscape.

The authors also acknowledged that one of the drawbacks of the use of RMOs has been to disregard the role of disturbance in shaping stream habitats. Natural disturbances play an imperative role in shaping the setting of streams and the conditions that are found within them (Benda et al. 1998). They went on to say that it is apparent that all streams will most likely not meet all habitat objectives during some point in their history as the series of natural disturbances both influences and resets them. In fact, some of the PIBO reference sites come from wilderness areas that have experienced severe disturbance from wildfires and associated debris flows. These sites provide valuable information when describing the distribution of conditions that may be possible in a reference setting and provide important information on recovery trajectories in the absence of land management. Based on this information and findings, MNF staff feels it is important to evaluate conditions against selected managed and reference riparian/channel attribute values, shown in Table 20. Again, these values only represent interim aspirations and are intended as a provisional step in the measurement of movement towards existing LRMP RMOs and ESA related MPI habitat values.

TABLE 20. PIBO MANAGED AND REFERENCE MEAN VALUES FOR SELECTED STREAM, CHANNEL MORPHOLOGY, AND HABITAT ATTRIBUTES.

Stream/Channel/Habitat Attribute	PIBO Managed Mean	PIBO Reference Mean	RMSE*	Source of Information
Bankfull W/D Ratio	23.9	22.6	4.0	Henderson et al. 2005

<b>Percent fines sediment (&lt;6mm)</b>	26.7	18.0	4.9 (4.8)	Al-Chokhachy et al. 2010
<b>Bank angle (°)</b>	108.0	99.3	(6.5)	Al-Chokhachy et al. 2010
<b>Percent undercut banks (%)</b>	26.4	32.7		Al-Chokhachy et al. 2010
<b>Residual pool depth (m):</b>	0.26	0.31	.027 (.016)	Al-Chokhachy et al. 2010
<b>Percent pool habitat (%)</b>	40.9	43.3	12.9 (5.8)	Al-Chokhachy et al. 2010
<b>Bank stability (%)</b>	74.6	79.9		Henderson et al. 2005
<b>D50 (mm)</b>	43.0	58.0	13.8 (8.0)	Al-Chokhachy et al. 2010

\* RMSE is an estimate of the potential sampling error for selected stream attributes (a measurement of the temporal variability – based on repeat measures analyses at sites over a 9-year period). Values are from Roper et al. 2010 and (Al-Chokhachy et al. 2011).

A cursory examination for the York On/Off Allotment reveals that six of the eight stream attributes considered to be potentially affected by livestock grazing – bankfull width-depth ratios, bank stability, % undercut banks, percent fines <6mm, pool %, and D50 – are found to be within PIBO managed mean values, whereas the other two, residual pool depth and bank angle, are outside the range of managed mean values. It should be noted that besides riparian vegetation, the stream attributes most directly affected by grazing activities are bank stability, bank angle, width to depth ratio, and percent undercut banks. As discussed earlier, a channel's bankfull width-depth ratio is an important indicator of whether a stream is able to perform the various tasks that lead to a healthy riparian area. This indicator, along with appropriate riparian vegetation, is critically important for a stream to maintain its dimension, pattern, and profile even during moderate to high (10-25+ year return intervals) flow events, like those that occurred in 2011. If continued monitoring shows that overall channel shape was maintained, the expected outcome will be improvement in the other stream attributes, thereby enhancing habitat complexity. Based on monitoring results, the current livestock management strategies utilized over the last several years should be continued. Continuing to meet the allowable use standards of the BiOp and LRMP, both in the form of “move triggers” and end of season minimum requirements, should avoid any negative effects to riparian or aquatic habitats that would carry over in any meaningful way to the following grazing season. In the absence of site-specific information to the contrary, it is fair to say that livestock grazing that complies with these applicable allowable use standards has a high likelihood of not meaningfully impeding the capacity for the structure and function of riparian and aquatic conditions to achieve recovery, consistent with PACFISH/INFISH S&Gs.

The results of this analysis can further be used to assist in the evaluation of other important parameters, such as water temperature in low-gradient alluvial systems, as they are directly tied to stream morphology (e.g., width/depth ratio) and vegetative cover. For example, as the streams narrow and deepen over time in an improving trajectory, water temperatures will be less susceptible to diurnal fluctuations, and as inflow from stored water in riparian banks increases,

cooler water temperatures in summer and warmer temperatures in winter would be expected. Thus, the evaluation of trend in stream morphology attributes such as width/depth ratio can help inform the Line Officer about other important fish habitat elements.

## 7 EFFECTS OF THE PROPOSED ACTION

The direct and indirect effects of implementing the action, including interrelated and interdependent actions, on the listed species and designated CH are evaluated in this section. In addition, the probability of directly affecting juveniles, spawning adults, and incubating embryos in redds will be assessed. The environmental impacts of implementing the project elements (PE) will be evaluated by use of NMFS/FWS MPI indicators to determine effects to ESA-listed MCR Steelhead and CR bull trout and their designated CH.

As described in this document, the proposed action is expected to allow previously degraded riparian areas/habitat indicators to continue recovery. However, it is anticipated that the proposed grazing activities in all cases will maintain the current environmental baseline condition for each indicator. In some cases indicators are rated as Not Properly Functioning, which suggests that the proposed grazing activities will be maintaining this risk rating. However, because the environmental baseline rating is determined at the subbasin scale, the proposed grazing activities tend to influence only portions of subbasins, and watershed restoration activities needed to improve the baseline indicators at the subbasin scale will not likely occur over the life of this consultation, it is anticipated that the proposed grazing activities will maintain the current environmental baseline condition. Historic and current cattle grazing in the ESA action area likely plays varying roles in the current environmental baseline ratings for these affected subbasins.

### 7.1 PROJECT ELEMENT AND INTERRELATED ACTION EVALUATION

The component parts of the action are listed in Section 4.1 as six project elements and are also shown below.

1. Livestock use of allotment/pastures
2. Permittee management of livestock and infrastructure maintenance
3. Range improvements
4. Exclusionary fences
5. Monitoring
6. Adaptive management

We determined that unauthorized use (trespass) is not an action. However, the implementation of FS enforcement actions regarding unauthorized use is an interrelated action.

#### 7.1.1 PROJECT ELEMENTS DROPPED FROM FURTHER ANALYSIS

An initial step in the analysis process is to determine if any of the project elements are already provided ESA coverage in a concluded programmatic consultation. The consultation history section (Section 1.1) described the Blue Mountain Expedited Section 7 Consultation (BMESSC) programmatic consultation. Range improvements are covered under that consultation. Range improvements in the BMESSC concurrence letter are described as: “e.g. fencing, off-site water

developments.” The consultation history section also described the Aquatic and Riparian Restoration Programmatic Consultation (ARRPC). Riparian exclusion fencing with water gaps and stream crossings is a category covered under the ARRPC biological opinion. Consequently, PEs 3 and 4 below already have existing ESA coverage and will not be further evaluated in this BA.

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#### 7.1.2 PROJECT ELEMENTS AND INTERRELATED ACTIONS WITH ENTIRELY BENEFICIAL EFFECTS

PE 6, adaptive management, provides a mechanism to adjust management if end-point indicators and desired conditions are not being met. Examples of adaptive management measures include reducing livestock numbers, changing the timing and duration of grazing, adjusting the numeric end-point indicators and constructing more exclusion fences. Making adjustments to ensure that end-point indicators and desired conditions are met will result in positive effects to habitat indicators and therefore to CH. The results would also have beneficial effects to the species, as many adaptive management adjustments will reduce the time that livestock are in or adjacent to streams.

Law enforcement actions to remove cattle not under permit will result in entirely beneficial effects to the species and designated CH.

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#### 7.1.3 PROJECT ELEMENTS REMAINING FOR ANALYSIS

Of the six PEs initially developed for this livestock grazing consultation, PEs 3 and 4 have been addressed as already covered by existing programmatic consultations still in effect, and the effect of implementing PE 6 has been determined to be entirely beneficial to CH and to the species. The PEs remaining for analysis are:

1. Livestock use of allotment/pastures
2. Permittee management of livestock and infrastructure maintenance
5. Monitoring

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##### 7.1.3.1 PE1: LIVESTOCK USE OF ALLOTMENT/PASTURES

Livestock will graze the allotment and individual pastures in the numbers, time frames and locations described in the proposed action section and in the term grazing permit.

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##### 7.1.3.2 PE2: PERMITTEE MANAGEMENT OF LIVESTOCK AND INFRASTRUCTURE MAINTENANCE

This PE includes the move-in and move-out of livestock using highway and off-road vehicles, and herding by range riders. While vehicles are also used to access sites for monitoring purposes (PE 5), the effects of vehicle use to CH and to the species will only be assessed for this PE to reduce redundancy in the analysis. Project Design Criteria for vehicle use are provided by the PDCs described earlier in the proposed action section.

Several hundred troughs, springs and ponds are maintained by grazing permittees to provide off-stream water for livestock across the MNF. In addition, there are miles of fence and dozens of gates that are maintained each year. Typical maintenance activities involve the use of hand tools



or machines on a small footprint of land. Some work such as repairing troughs or replacing wire will not involve any soil or vegetation disturbance. Other maintenance activities may disturb small amounts of soil and vegetation, but rarely within riparian areas adjacent to MCR Steelhead CH. Workers performing maintenance activities rarely walk in riparian areas or in stream channels where listed fish are present or in designated CH.

### 7.1.3.3 PE5: MONITORING

A variety of implementation and effectiveness monitoring techniques are employed to determine if desired conditions are being met. The MNF Riparian Monitoring Strategy is discussed in detail in the Monitoring section (Section 4.1.4). Workers use manual and electronic equipment to measure vegetation, water quality and stream channel/streambed characteristics. Some monitoring actions include wading in stream channels.

## 7.2 ANALYSIS OF EFFECTS TO DESIGNATED CRITICAL HABITAT

The three PEs will be analyzed first for their effects to designated CH, then for effects to the species. The freshwater primary constituent elements (PCE) of MCR Steelhead CH applicable to the action area are presented in Table 21.

TABLE 21. PRIMARY CONSTITUENT ELEMENTS OF MCR STEELHEAD CRITICAL HABITAT APPLICABLE TO THE ACTION AREA.

PCE	Description
1	Freshwater spawning sites with water quantity and quality conditions and substrate supporting spawning, incubation and larval development.
2	Freshwater rearing sites with: (i) Water quantity and floodplain connectivity to form and maintain physical habitat conditions and support juvenile growth and mobility; (ii) Water quality and forage supporting juvenile development; and (iii) Natural cover such as shade, submerged and overhanging large wood, log jams and beaver dams, aquatic vegetation, large rocks and boulders, side channels, and undercut banks.
3	Freshwater migration corridors free of obstruction and excessive predation with water quantity and quality conditions and natural cover such as submerged and overhanging large wood, aquatic vegetation, large rocks and boulders, side channels, and undercut banks supporting juvenile and adult mobility and survival.

The effects to each PCE, and ultimately to designated CH as a whole, can be determined by evaluating the effects to indicators of the NMFS MPI that correspond to each PCE. The MNF uses a crosswalk table format for this purpose. Table 22 presents the analysis for effects of the action to the PCEs of MCR Steelhead designated CH. Table 23 presents a summary of effects to the indicators associated with each PCE of MCR Steelhead CH. Table 24 presents the analysis for effects of the action to the PCEs of CR bull trout designated CH. Measurable effects to several habitat indicators that correspond to specific PCEs were concluded for each analysis.

The effects by indicator and PCE are summarized in Table 23 for the MCR Steelhead PCEs and in Table 25 for the CR bull trout PCEs.

Measurable effects to several habitat indicators of PCEs were concluded.

TABLE 22. ANALYSIS OF EFFECTS TO MPI INDICATORS CORRESPONDING TO PCEs OF DESIGNATED CRITICAL HABITAT FOR MCR STEELHEAD WITHIN THE BEAR AND YORK ON/OFF ALLOTMENTS.

PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
(1) Freshwater spawning sites with water quantity and quality conditions and substrate supporting spawning, incubation and larval development;	Water Quantity	Flow/ Hydrology	Changes in Peak/Base Flows	<p>Riparian vegetation has been linked to the water-holding capacity of streamside aquifers (Platts 1991). As riparian vegetation is removed by grazing and streamside soils are compacted by hooves, the ability of areas to retain water is decreased. Decreased evapotranspiration and infiltration increase and hasten surface runoff, resulting in a more rapid hydrologic response of streams to rainfall. When this occurs, high flows in the spring tend to increase in volume, leading to bank damage and erosion, and channel downcutting. Summer and fall base flows are decreased, often resulting in flows that are insufficient to provide suitable rearing habitat for juvenile salmonids. If aquifers lose their capacity to hold and slowly deliver water to the stream, differences between peak and base discharge rates increase dramatically (USEPA 1993). Some streams that typically flowed perennially may experience periods of no flow in the summer or fall. Li <i>et al.</i> (1994) found that flow in a heavily grazed eastern Oregon stream became intermittent during the summer, while a nearby, well-vegetated reference stream in a similar-sized watershed had permanent flows. They suggested that the difference in flow regimes was due to diminished interaction between the stream and floodplain with resultant lowering of the water table.</p> <p>Indirect effects of historic livestock grazing in the ESA action area (including trailing and watering), on channel and bank features such as bank stability, undercut banks and width to depth ratio, as well as impacts to shrub recruitment and green line plant vigor, have likely affected peak and base flows on some streams. It is anticipated that PE 1 (livestock use) will have negative effects to this indicator, but they will be <b>too small</b> to be <b>meaningfully measured</b>, particularly to flows at the time of year when spawning, incubation and larval development occur. The use of BMPs, end point indicators, and adaptive management should minimize effects. If hydrophytic vegetation, bank stability, width-depth ratio, and undercut banks show a static and/or downward trend and the Forest is not meeting RMOs, grazing practices will be modified (See Adaptive Management Section VI). PE 2 (permittee mgt. and mtce.) includes off-road vehicle use. This has the potential to increase soil compaction, but it will be minimized by use of PDCs. Little to no riparian vegetation is affected by vehicle use, range riding or maintenance activities. PE 2 overall will have slight negative effects to the indicator that are too small to be meaningfully measured. PE5 (monitoring) will not increase compaction or remove vegetation, and therefore does not have</p>

PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
				a mechanism to affect peak/base flow. The effect to the indicator is neutral.
			Increase in Drainage Network	None of the PEs has road construction, so no change to the drainage network will occur. The proposed action would have a neutral effect on the indicator.
	Water Quality	Flow/ Hydrology	Temperature	<p>The temperature monitoring data for the Mosquito (Bear allotment) and Slide (York allotment) creeks and Middle Fork John Day River (Bear allotment) monitoring sites did not meet state water quality standards and rated NPF using the NMFS MPI criteria (see Section 6.1.3).</p> <p>Many grass/grass-like species found on the MNF have an ungrazed potential height of 2 to 3 feet (MNF 2007). In meadow streams with narrow channels, they often are the plants that provide stream shade. PE 1 (livestock use) will potentially reduce vegetation heights to 4 or 6 inches. This will considerably reduce stream shade in those circumstances compared to the ungrazed potential vegetation heights (see discussion that follows in Effects to Listed Species section).</p> <p>Livestock use (PE 1) is likely to result in measurable water temperature increases for certain stream reaches. These impacts are expected to be generally confined to low gradient stream channels less than 10 feet wide with grass/grass-like vegetation providing shade. In particular, this impact is expected in Slide Creek in the York On/Off Allotment. Increases in water temperature are not expected along the Middle Fork John Day River because of the general exclusion of cattle and because it is a wide stream. Measurable increases in water temperature are not expected in Mosquito Creek in the Bear Allotment because the stream is largely confined by hillslopes and well-shaded with alder, dogwood, and other hardwoods. The effect to this indicator by livestock use in the York On/Off Allotment is <b>negative and meaningfully measured</b>. Water temperatures typically are below concern thresholds when spawning, incubation and larval development of MCR Steelhead occurs, as flows are greater than later in the year. PE 2 (permittee livestock management and infrastructure maintenance) and PE 5 (monitoring) activities will not remove vegetation that provides shade nor affect channel-forming processes that might widen stream channels. Consequently, there is no mechanism for PEs 2 and 3 to affect water temperature and the</p>

PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
				<p>effect of the PE for the indicator is neutral.</p> <p>Livestock grazing on federal land in the ESA action area is managed to attain the endpoint indicators, which were developed to meet PACFISH grazing standards and guidelines as well as water quality BMPs. The assumption is that meeting these endpoint indicators would move key riparian and stream channel elements (bank stability, W/D ratio, woody species regeneration) towards their Desired Conditions and meet Riparian Objectives, and thereby maintain water temperatures. If monitoring fails to show this trend, adaptive management would be implemented and endpoint indicators would be modified to minimize adverse impacts to this element of the PCE.</p>
			Sediment/ Turbidity	<p>The monitoring results presented above indicate that livestock use (PE 1), as well as use by wild ungulates, results in trampled and grazed riparian vegetation, and altered stream banks to some degree. Livestock also use trails to access streams for water. Livestock occasionally will concentrate their use in certain areas, potentially creating patches of relatively bare soil. Some of these areas may be adjacent to stream sections potentially used by MCR Steelhead for spawning, incubation and larval development, such as Slide Creek. Bare soil is prone to erosion and can result in fine sediment entering stream channels and resultant increases in turbidity. Habitat impacts are likely to include areas of exposed streambank up to a few feet wide where livestock access streams to drink or cross, and areas of bank disturbance where livestock graze in riparian areas. Exposed areas and other bank disturbances that occur are likely to result in a slight increase in turbidity for a short distance downstream during rainstorms or runoff events. However, given background levels of turbidity during runoff events it will be difficult to distinguish between turbidity resulting from these grazing impacts and background turbidity. A slight increase in fine sediment deposition for a short distance downstream of exposed and disturbed areas is also likely to occur.</p> <p>Endpoint indicators were developed in order to meet PACFISH grazing standards and guidelines as well as water quality BMPs. The assumption is that meeting these endpoint indicators would move key riparian and stream channel elements (bank stability, w/d ratio, woody species regeneration) towards their Desired Conditions and meet our Riparian Objectives. If monitoring fails to show this trend, adaptive management would be</p>

PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
				<p>implemented (Section 4.1.5) and endpoint indicators would be modified to minimize adverse effects to critical habitat.</p> <p>However, livestock grazing will increase the amount of sediment entering streams by the mechanisms described above. These impacts are expected to be localized and short-term. Consequently, the effect to this indicator in Slide Creek in the York On/Off Allotment by PE1 (livestock use) <b>is negative and expected to be measurable</b> using bank alteration methods described above as a proxy for turbidity and sedimentation. In the Bear Allotment, limited cattle crossings occur in Mosquito Creek. Due to the well armored banks and relatively well vegetated riparian zone, the amount of sediment expected to be generated from these localized crossings, the effects to this indicator from PE1 in the Bear Allotment are expected to be <b>negative and not meaningfully measured</b>.</p> <p>PE 2 involves use of vehicles on and off roads, as well as infrastructure maintenance. There is the potential for fine sediment to be transported from unpaved roads to stream channels, primarily at road crossings, during rainstorms or runoff events. However, it is impossible to determine the proportion of the suspended sediment attributable to road use by permittees, given the use of the roads for other purposes. In addition, background levels of suspended sediment in streams will be high during rainstorms and runoff events, and the contribution by permittee use of roads to increased turbidity cannot be meaningfully measured. Use of off-road vehicles should not result in measurable effects due to use of PDCs. Range riding with horses will not cause any meaningfully measured increases in streambed sediment or turbidity. Maintenance activities are typically distant from designated CH, disturb little to no soil, and are not hydrologically connected to stream channels. There is no mechanism for maintenance activities to affect the indicator. Overall, the effects of PE 2 to the indicator are <b>negative and not meaningfully measured</b>.</p> <p>Monitoring (PE 3) activities such as pebble counts and measuring cross-sections involve wading in stream channels. Other monitoring activities involve walking or riding horses in riparian areas. The timing of these activities is typically after spawning, incubation and larval development of MCR Steelhead, although there may be some overlap in timing. Spawning surveys also involve wading. Wading may result in very small increases in turbidity downstream for a short distance (a few feet) that will quickly dissipate. Walking</p>

PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
				and riding horses in riparian areas should not result in fine sediment delivery to stream channels. However, there may be very small and transient increases in turbidity when a stream is being crossed. The monitoring PE effect to the indicator is <b>negative, but not meaningfully measured</b> .
			Chemical Contamination/ Nutrients	<p>Urine and feces from livestock use (PE 1) in riparian areas increases the likelihood that nitrogen and phosphorous will enter streams. Increased nutrients will likely increase stream productivity at the source of nutrients and for a short distance downstream. It is anticipated that livestock grazing will have slight negative impacts to the indicator, but they are <b>not expected to be meaningfully measured</b>.</p> <p>PE 2 (permittee management and infrastructure maintenance) includes vehicle use. The risk of chemical contamination to streams will be minimized by use of PDC. Maintenance activities are typically distant from designated CH, and at locations not hydrologically connected to stream channels. Therefore, there is no mechanism for petroleum products spilling from power tools to affect CH. Use of horses for range riding will have similar effects (but much smaller scale) than that of PE 1, above. Maintenance activities are typically distant from stream channels. The overall effect of PE 2 is for <b>slight negative effects</b> to the indicator that are <b>not expected to be meaningfully measured</b>.</p> <p>Monitoring (PE 5) does not involve the use of chemicals and does not have the potential to affect nutrients in streams. PE 5 will have a neutral effect to the indicator.</p>

PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
	Suitable Substrate	Habitat Elements	Substrate Embeddedness	The analysis of effects to the sediment/turbidity indicator, above, determined that use of riparian areas by livestock is expected to increase the amount of sediment entering streams. A slight increase in fine sediment deposition for a short distance downstream of exposed and disturbed areas is likely to occur. There is the potential for fine sediment to slightly increase embeddedness within gravels suitable for spawning when the gravel is located immediately downstream from exposed and disturbed streambank areas. The effect to this indicator by livestock use (PE 1) is <b>negative and meaningfully measurable</b> , in Slide Creek of the York On/Off Allotment using bank alteration survey protocol as a proxy for embeddedness. In the Bear Allotment, embeddedness is expected to be <b>negative but insignificant</b> because most of Mosquito Creek in the allotment is well armored and vegetated with woody vegetation. The analysis for sediment/turbidity determined that PE 2 would have a slightly negative, but <b>not meaningfully measured effect</b> to the indicator. Therefore, the same conclusion is made for the substrate embeddedness indicator. As described above, monitoring (PE 5) would not introduce fine sediment into stream channels. The monitoring PE will have a <b>neutral effect</b> to the indicator.
(2) Freshwater rearing sites with:  (i) Water quantity and floodplain connectivity to form and maintain physical habitat	Water Quantity	Flow/ Hydrology	Changes in Peak/Base Flows	See discussion above.
			Increase in Drainage Network	See discussion above.
	Water Quality	Water Quality	Temperature	See discussion above. The rearing period includes the summer months when elevated water temperatures are most concerning for juvenile salmonids, and the sun's position in the sky results in the greatest potential for increased solar radiation to streams. It is this time period when the small, <b>but measurable</b> , increases to water temperature in Slide Creek of the York Allotment would take place.



PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
conditions and support juvenile growth and mobility;  (ii) Water quality and forage			Sediment/ Turbidity	See discussion above.
			Chemical Contamination/ Nutrients	See discussion above. The conclusion was for a <b>slight negative effect</b> to the indicator from livestock use (PE 1) and permittee management and infrastructure maintenance (PE 2) since there would be an increase in nutrients into streams. However, the introduction of nutrients may lead to small increases in contamination from fecal coliform.

PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
<p>supporting juvenile development; and</p> <p>(iii) Natural cover such as shade, submerged and overhanging large wood, log jams and beaver dams, aquatic vegetation, large rocks and boulders, side channels, and undercut banks</p>	Flood-plain Connect-ivity	Channel Condition and Dynamics	Floodplain Connectivity	<p>Channel entrenchment is the main concern for loss of floodplain connectivity. Indirect effects of livestock use (PE 1), including trailing and watering, on things such as bank stability, undercut banks, width depth ratio, shrub recruitment, and green line plant vigor have limited some streams' ability to access their floodplains, thus concentrating energies within confined channels and causing additional erosion. Many of these streams are still experiencing this phenomenon.</p> <p>Channel entrenchment as a result of livestock use (PE 1) will be prevented by use of endpoint indicators to meet PACFISH grazing standards and guidelines as well as water quality BMPs. The conclusion is that the effect to the indicator by livestock use is negative but not meaningfully measured. The assumption is that meeting these endpoint indicators would move key riparian and stream channel elements (bank stability, w/d ratio, woody species regeneration) towards their Desired Conditions and meet Riparian Objectives. If monitoring fails to show this trend, adaptive management would be implemented and endpoint indicators would be modified to minimize negative effects to floodplain connectivity.</p> <p>PE 2 (permittee management and infrastructure maintenance) includes on and off road vehicle use. Road use has no mechanism to affect floodplain connectivity. PDC for off-road use will prevent channel downcutting. Range riding with horses will occasionally cross a stream but effects to streambanks and beds will be so minimal as to not affect the indicator. Infrastructure maintenance actions do not affect streambanks or riparian vegetation adjacent to CH, and will therefore not affect floodplain connectivity. The overall effect of PE 2 is a neutral affect to the indicator.</p> <p>Monitoring (PE 5) does not remove riparian vegetation or otherwise have mechanisms to destabilize stream channels. PE 5 will have a neutral effect to the indicator.</p>

PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
			Width/Depth Ratio	Three level II stream surveys reported bankfull width-depth ratios ranging from 11.3 to 19.5, with 1 stream reach meeting the NMFS MPI criterion for AR (10-12) and 2 meeting the criteria for NPF (>12, see Appendix J for PIBO and stream survey monitoring data). Data from the one PIBO integrator site shows a substantial decrease in W/D ratio between 2005 and 2010 at the Slide Creek site (12 to 8). Livestock use (PE 1) is anticipated to have a <b>negative</b> , but not <b>meaningfully measured effect</b> to the indicator. The potential for increases in width-depth ratio is reduced because of implementation of endpoint indicators for livestock grazing (which includes use by wild ungulates) and adaptive management. McDowell and Mowry (2002) found width to depth ratios were higher in unfenced areas of Camp Creek (close to the action area) compared to exclosed areas. PE 2 (permittee management and infrastructure maintenance) includes on and off road vehicle use. Road use has no mechanism to affect W/D ratio. PDC for off-road use will prevent bank damage and effects to W/D ratio. Range riding with horses will occasionally cross a stream but effects to streambanks and beds will be so minimal as to not affect the indicator. Infrastructure maintenance actions do not affect streambanks or riparian vegetation adjacent to CH, and will therefore not affect W/D ratio. The overall effect of PE 2 is a <b>neutral effect</b> to the indicator. PE 5 (monitoring) does not remove vegetation or destabilize stream banks. There is no potential for it to increase W/D ratio. The monitoring PE will have a <b>neutral effect</b> to the indicator.
	Forage	Habitat Elements	Substrate Embeddedness	See discussion above for this indicator for the suitable substrate PCE habitat feature. The conclusion for livestock use (PE 1) was that a slight increase in fine sediment deposition for a short distance downstream of exposed and disturbed areas is likely to occur. This would result in small areas of increased embeddedness. Increased embeddedness may result in a decrease in the potential for production of aquatic macroinvertebrates (a forage item for rearing salmonids) in small, isolated patches. The conclusion is that PE 1 will have a <b>slight negative effect</b> on substrate embeddedness with respect to the production of forage. Consistent with the analysis for the suitable substrate PCE habitat feature, the effect of PCE 2 to the indicator is <b>negative but not meaningfully measured</b> , and the effect of PCE 5 is <b>neutral</b> .

PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
			Large Woody Debris	<p>Livestock grazing does not affect this indicator in conifer-dominated riparian forests. Livestock use can negatively affect this indicator when grazing occurs within hardwood stands such as aspen, alder, birch, and cottonwoods that could contribute larger pieces of wood to small streams. In sites in the action area that would be naturally dominated by cottonwood gallery riparian forests, such as the Middle Fork John Day River, livestock use (PE 1) could result in altering the level of cottonwood stocking and future large tree (and subsequent large woody debris) recruitment (Kaufmann et al. 1983, Case and Kaufmann 1997, Beschta and Ripple 2005). However, keeping cattle out of the riparian zone with electric fencing and/or resting the riparian pastures along the MFJDR will minimize loss of future cottonwood recruitment and effects to the large wood indicator are expected to be <b>negative but insignificant</b> in the Bear allotment. In Slide Creek in the York On/Off Allotment, grazing has reduced the recruitment and regeneration of woody vegetation along the stream corridor. Some conifers are still present and low to moderate levels of instream large wood exist. However, it is anticipated that livestock will continue to graze young woody species at levels <b>meaningfully measured</b> with respect to the future production of large woody debris. Thus, the effect to this indicator by livestock use (PE 1) is <b>negative and meaningfully measured</b> in the York On/Off Allotment. Using BMPs, end point indicators, and adaptive management will result in discouraging browse on existing hardwoods and willows but may not promote regeneration of new cottonwoods. PE 2 and PE 5 do not affect trees and associated LWD in any way. Therefore there is no mechanism for an effect and the <b>effect is neutral</b> to the indicator for both PEs</p>

PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
			Pool Frequency	<p>Pool frequency at the PIBO monitoring sites and stream survey reaches within the allotments are not currently meeting the DFC as described within Amendment 29 of the MNF LRMP and would be NPF using NMFS MPI criteria. See Appendix J for PIBO and stream survey results.</p> <p>Indirect effects of livestock grazing (including trailing and watering), on bank stability, undercut banks, width-depth ratio, shrub recruitment, green line plant composition and vigor have the potential to affect this indicator. The use of BMP's for livestock management, end point indicators and adaptive management, should result in an overall effect by PE 1 (livestock use) to pool frequency that is <b>not meaningfully measured</b> and unlikely to occur.</p> <p>PE 2 (permittee management and infrastructure maintenance) includes on and off road vehicle use. Road use has no mechanism to affect pool frequency. PDC for off-road use will prevent bank damage and effects to pool frequency. Range riding with horses will occasionally cross a stream but effects to streambanks will be so minimal as to not affect the indicator. Infrastructure maintenance actions do not affect streambanks or riparian vegetation adjacent to CH, and will therefore not affect pool frequency. The overall effect of PE 2 is a <b>neutral effect</b> to the indicator.</p> <p>PE 5 (monitoring) does not have any mechanisms to affect plants or bank and channel features that would impact pool frequency. The monitoring PE has a <b>neutral effect</b> to the indicator.</p>

PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
			Pool Quality	<p>PIBO data within the allotments (n=2) indicate that undercut bank frequency is less than the desired condition. Residual pool depth is generally low with very few to no pools greater than 1 meter deep in surveyed stream reaches (this parameter has changed somewhat along the Middle Fork John Day River engineered reach in the Bear Allotment). See Appendix J for PIBO and stream survey results. Based upon the PIBO and stream survey data, pool quality would be considered to be NPF using NMFS MPI criteria.</p> <p>Indirect effects of livestock grazing (including trailing and watering), on bank stability, undercut banks, width-depth ratio, shrub recruitment, green line plant composition and vigor have the potential to affect this indicator. The use of BMPs for livestock management, end point indicators (which are inclusive of wild ungulate use), and adaptive management, should result in an overall effect by PE 1 (livestock use) to pool quality that is <b>negative and not meaningfully measured</b>.</p> <p>PE 2 (permittee management and infrastructure maintenance) includes on and off road vehicle use. Road use has no mechanism to affect pool quality. PDC for off-road use will prevent bank damage and effects to pool quality. Range riding with horses will occasionally cross a stream but effects to streambanks will be so minimal as to not affect the indicator. Infrastructure maintenance actions do not affect streambanks or riparian vegetation adjacent to CH, and will therefore not affect pool frequency. The overall effect of PE 2 is a <b>neutral effect to the indicator</b>.</p> <p>PE 5 (monitoring) does not have any mechanisms to affect plants or bank and channel features that would impact pool quality. The monitoring PE has a <b>neutral effect</b> to the indicator.</p>

PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
			Off-Channel Habitat	<p>There is very little off-channel habitat within streams draining the action area. Of 3 level II stream surveys that reported side-channel information, all reported less than 4% of the stream reaches contained side channels or off-channel area. The use of BMP's for livestock management, end point indicators (which are inclusive of wild ungulate use), and adaptive management, should result in an overall effect by PE 1 (livestock use) to off-channel habitat that is <b>negative and not meaningfully measured</b>.</p> <p>PE 2 (permittee management and infrastructure maintenance) includes on and off road vehicle use. Road use has no mechanism to affect off-channel area. PDC for off-road use will prevent bank damage and effects to off-channel habitat. Range riding with horses will occasionally cross a stream but effects to streambanks will be so minimal as to not affect the indicator. Infrastructure maintenance actions do not affect streambanks or riparian vegetation adjacent to CH, and will therefore not affect off-channel areas. The overall effect of PE 2 is a <b>neutral effect</b> to the indicator.</p> <p>PE 5 (monitoring) does not have any mechanisms to affect off-channel habitat. Therefore, the monitoring PE has a <b>neutral effect</b> to the indicator.</p>

PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
			Refugia	<p>The availability of refugia is a limiting factor identified in the recovery plan for the Oregon steelhead population of the MCR Steelhead distinct population segment (NMFS 2009). The NMFS MPI (NMFS 1996) defines the Refugia indicator as: “important remnant habitat for sensitive aquatic species.” All of the habitat indicators in this crosswalk table are potential components of Refugia. Analysis for previous indicators has determined that PE 1 (livestock use) will have negative and meaningfully measured effects to several of them. This may occur in areas that meet the definition of Refugia. Therefore, PE 1 (livestock use) will have <b>negative and meaningfully measured</b> or evaluated effects to the Refugia indicator in Slide Creek of the York Allotment. In the Bear Allotment this indicator would have a <b>slight negative but insignificant effect</b> for reasons described above in the temperature, sediment, and large woody debris section.</p> <p>The highest level of effect to previous indicators by PE 2 (permittee management and infrastructure maintenance) was “negative but not meaningfully measurable.” This level of effects will not impact the function of Refugia to provide important remnant habitat. Therefore, the effect conclusion is <b>neutral</b> for PE 2.</p> <p>The highest level of effect to previous indicators by PE 5 (monitoring) was “negative but not meaningfully measurable” for small and transient increases in turbidity by wading in stream channels or crossing streams on foot or by horse. This level of effects will not impact the function of Refugia to provide important remnant habitat. Therefore, the effect conclusion is <b>neutral</b> for the monitoring PE.</p>



PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
		Watershed Condition	Riparian Reserves	<p>As described above, PE 1 (livestock use) will result in negative effects to indicators within riparian areas. A negative effect to Riparian Management Areas (RMA) (east-side analog of Riparian Reserves) is indicated. However, the negative effects should not rise to the level that impacts to the processes and functions of RMAs are meaningfully measurable. Endpoint indicators were developed with seral class in mind to meet PACFISH grazing standards and guidelines, enclosure B of the LMRP and water quality BMPs. The assumption is that meeting these endpoint indicators would move key riparian and stream channel elements (bank stability, w/d ratio, woody species regeneration) towards their Desired Conditions and meet Riparian Objectives. If monitoring fails to show this trend, adaptive management would be implemented and endpoint indicators would be modified to minimize adverse effects to Riparian Reserves. Therefore the determination of effects to the riparian reserves indicator is <b>negative but not meaningfully measurable</b>.</p> <p>The highest level of effect to previous indicators by PE 2 (permittee management and infrastructure maintenance) was “negative but not meaningfully measurable.” This level of effects will not impact the processes and functions of RMAs. Therefore, the effect conclusion is <b>neutral</b> for PE 2.</p> <p>The monitoring PE does not have any mechanisms to affect the processes and functions of RMAs. The monitoring PE has a <b>neutral</b> effect to the indicator.</p>

PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
	Natural Cover	Habitat Elements	Substrate	<p>This indicator focuses on the composition of streambed substrate rather than its embeddedness, which was evaluated earlier. Historic stream survey data indicated that Mosquito Creek within the allotments has substrate that is dominated by sand and over 30% embedded. See Substrate Embeddedness indicator above. Based on the stream survey data the Substrate indicator would be classified as NPF using NMFS MPI criteria. However, observations from summer 2011 suggested the substrate was gravel dominated with low levels of embeddedness (Reid unpub. observations).</p> <p>The analysis of effects to the sediment/turbidity indicator for PCE 1, above, determined that use of riparian areas by livestock is expected to increase the amount of sediment entering streams. A slight increase in fine sediment deposition for a short distance downstream of exposed and disturbed areas is likely to occur in Mosquito Creek in the Bear Allotment. However, this is not expected to measurably change the composition of existing substrate with regard to its function as cover for juvenile or adult MCR Steelhead. Therefore, the effect to this indicator by PE 1 (livestock use) is <b>negative and not meaningfully measurable</b>. The use of BMP's for livestock management, end point indicators (which are inclusive of wild ungulate use), and adaptive management, should further minimize the magnitude of potential negative effects by PE 1. In Slide Creek, in the York On/Off Allotment, fine sediment is prevalent in pools and riffle margins and has smothered gravel and cobble substrates reducing natural cover. Therefore, effects to this indicator are expected to be <b>negative and meaningfully measurable</b> in the York On/Off Allotment.</p> <p>The analysis of effects to the sediment/turbidity indicator for PCE 1, above, determined that the effect of PE 2 (permittee management and infrastructure maintenance) was "negative and not meaningfully measured." This level of effects is not expected to measurably change the composition of existing substrate with regard to its function as cover for juvenile or adult MCRS Steelhead. Therefore, the effect to this indicator by PE 1 (livestock use) is <b>negative and not meaningfully measurable</b>.</p> <p>As described above, PE 5 (monitoring) would not introduce fine sediment into stream channels. The monitoring PE will have a <b>neutral effect</b> to the indicator.</p>

PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
			Large Woody Debris	See Above.
			Pool Frequency	See Above.
			Pool Quality	See Above
			Off-Channel Habitat	See Above
			Refugia	See Above
		Watershed Condition	Riparian Reserves	See Above

PCE	PCE Habitat Feature	Matrix Pathway	Matrix Indicator	Rationale
(3) Freshwater migration corridors free of obstruction and excessive predation with water quantity and quality conditions and natural cover such as submerged and overhanging large wood, aquatic vegetation, large rocks and boulders, side channels, and undercut banks supporting juvenile and adult mobility and survival;	Migration Corridors Free of Obstruction	Habitat Access	Physical Barriers	No barriers will be created or removed by the actions of any PE. All PEs have a <b>neutral effect</b> on the physical barriers indicator.

TABLE 23. SUMMARY OF EFFECTS OF THE PROPOSED ACTION BY THE PROJECT ELEMENTS OF LIVESTOCK GRAZING IN THE BEAR AND YORK ON/OFF PEAK ALLOTMENTS TO THE INDICATORS ASSOCIATED WITH HABITAT FEATURES OF EACH PRIMARY CONSTITUENT ELEMENT OF MCR STEELHEAD CRITICAL HABITAT.

Primary Constituent Element	PCE Habitat Feature	Indicator	Effect Conclusion by Project Element		
			PE1: Livestock Use	PE2: Permittee Management of Livestock and Infrastructure Maintenance	PE 5: Monitoring
1. Freshwater spawning sites with water quantity and quality conditions and substrate supporting spawning, incubation and larval development	Water quantity	Changes in Peak/Base Flows	NNMM <sup>1</sup>	NNMM	Neutral
		Increase in Drainage Network	Neutral	Neutral	Neutral
	Water quality	Temperature	<i>NMM<sup>2</sup> (York)</i> NNMM (Bear)	Neutral	Neutral
		Sediment/Turbidity	<i>NMM (York)</i> NNMM (Bear)	NNMM	NNMM
		Chemical Contamination/Nutrients	NNMM	NNMM	Neutral
	Suitable substrate	Substrate Embeddedness	<i>NMM (York)</i> NNMM (Bear)	NNMM	Neutral
2. Freshwater rearing sites with:  (i) Water quantity and floodplain connectivity to form and maintain physical habitat conditions and support juvenile growth and mobility;  (ii) Water quality and forage supporting juvenile development; and	Flow/hydrology	Changes in Peak/Base Flows	NMM	NNMM	Neutral
		Increase in Drainage Network	Neutral	Neutral	Neutral
	Water quality	Temperature	<i>NMM (York)</i> NNMM (Bear)	Neutral	Neutral
		Sediment/Turbidity	<i>NMM (York)</i> NNMM (Bear)	NNMM	NNMM

Primary Constituent Element	PCE Habitat Feature	Indicator	Effect Conclusion by Project Element		
			PE1: Livestock Use	PE2: Permittee Management of Livestock and Infrastructure Maintenance	PE 5: Monitoring
(iii) Natural cover such as shade, submerged and overhanging large wood, log jams and beaver dams, aquatic vegetation, large rocks and boulders, side channels, and undercut banks		Chemical Contamination/Nutrients	NNMM	NNMM	Neutral
	Floodplain connectivity	Floodplain Connectivity	NNMM	Neutral	Neutral
		Width/Depth Ratio	NNMM	Neutral	Neutral
	Forage	Substrate Embeddedness	<i>NMM (York)</i> NNMM (Bear)	NNMM	Neutral
		Large Woody Debris	<i>NMM (York)</i> NNMM (Bear)	Neutral	Neutral
		Pool Frequency	NNMM	Neutral	Neutral
		Pool Quality	NNMM	Neutral	Neutral
		Off-Channel Habitat	NNMM	Neutral	Neutral
		Refugia	<i>NMM (York)</i> NNMM (Bear)	Neutral	Neutral
		Riparian Reserves	NNMM	Neutral	Neutral
	Natural cover	Substrate	<i>NMM (York)</i> NNMM (Bear)	NNMM	Neutral
		Large Woody Debris	<i>NMM (York)</i> NNMM (Bear)	Neutral	Neutral
		Pool Frequency	NNMM	Neutral	Neutral

Primary Constituent Element	PCE Habitat Feature	Indicator	Effect Conclusion by Project Element		
			PE1: Livestock Use	PE2: Permittee Management of Livestock and Infrastructure Maintenance	PE 5: Monitoring
		Pool Quality	NNMM	Neutral	Neutral
		Off-Channel Habitat	NNMM	Neutral	Neutral
		Refugia	<i>NMM (York)</i> NNMM (Bear)	Neutral	Neutral
		Riparian Reserves	NNMM	Neutral	Neutral
3. Freshwater migration corridors free of obstruction and excessive predation with water quantity and quality conditions and natural cover such as submerged and overhanging large wood, aquatic vegetation, large rocks and boulders, side channels, and undercut banks supporting juvenile and adult mobility and survival	Migration corridors free of obstruction	Physical Barriers	Neutral	Neutral	Neutral

<sup>1</sup>NNMM = Negative, not meaningfully measured

<sup>2</sup>*NMM* = Negative, meaningfully measured

TABLE 24. ANALYSIS OF EFFECTS TO BULL TROUT MPI INDICATORS CORRESPONDING TO PCEs OF DESIGNATED CRITICAL HABITAT FOR COLUMBIA RIVER BULL TROUT WITHIN THE BEAR ALLOTMENT.

PCE	PCE Habitat Feature	MPI Pathway	MPI Indicator	Rationale
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PCE	PCE Habitat Feature	MPI Pathway	MPI Indicator	Rationale
1) Springs, seeps, groundwater sources, and subsurface water connectivity (hyporheic flows) to contribute to water quality and quantity and provide thermal refugia	Floodplain Connectivity	Channel Condition and Dynamics	Floodplain connectivity	<p>Channel entrenchment is the main concern for loss of floodplain connectivity. Indirect effects of livestock use (PE 1), including trailing and watering, on things such as bank stability, undercut banks, width depth ratio, shrub recruitment, and green line plant vigor have limited some streams' ability to access their flood plains, thus concentrating energies within confined channels and causing additional erosion. Many of these streams are still experiencing this phenomenon.</p> <p>Channel entrenchment as a result of livestock use (PE 1) will be prevented by use of endpoint indicators to meet PACFISH / INFISH grazing standards and guidelines as well as water quality BMPs. No bull trout critical habitat is present in the York On/Off allotment so this allotment would be No Effect for all PEs and MPI indicators. For the Bear Allotment, bull trout critical habitat is present along the Middle Fork John Day River only. In this allotment, fences and scheduled non-use of the riparian pastures during the term of this consultation will prevent cattle use along the MFJDR greenline, which could lead to channel entrenchment, reducing floodplain connectivity. Therefore, there is a <b>neutral effect</b> on the floodplain connectivity indicator from PE1.</p> <p>PE 2 (permittee management and infrastructure maintenance) includes on and off road vehicle use. Road use has no mechanism to affect floodplain connectivity. PDC for off-road use will prevent channel downcutting. Range riding with horses will not cross the Middle Fork John Day River due to fences. Infrastructure maintenance actions do not affect streambanks or riparian vegetation adjacent to CH, and will therefore not affect floodplain connectivity. The overall effect of PE 2 is <b>neutral</b> to the indicator.</p> <p>Monitoring (PE 5) does not remove riparian vegetation or otherwise have mechanisms to destabilize stream channels. PE 5 will have a <b>neutral</b> effect to the indicator.</p>
	Water Quantity	Flow/ Hydrology	Change in peak/base flows	<p>Riparian vegetation has been linked to the water-holding capacity of streamside aquifers (Platts 1991). As riparian vegetation is removed by grazing and streamside soils are compacted by hooves, the ability of areas to retain water is decreased. Decreased evapotranspiration and infiltration increase and hasten surface runoff, resulting in a more rapid hydrologic response of streams to rainfall. When this occurs, high flows in the spring tend to increase in volume, leading to bank damage and erosion, and channel downcutting. Summer and fall base flows are decreased, often resulting in flows that are insufficient to provide suitable rearing habitat for juvenile salmonids. If aquifers lose their capacity to hold and slowly deliver water to the stream, differences between peak and base discharge rates increase dramatically (EPA 1993). Some streams that typically flowed perennially may experience periods of no flow in the summer or fall. Li <i>et al.</i> (1994) found that flow in a heavily grazed eastern Oregon stream became intermittent during the summer, while a nearby, well-vegetated reference stream in a similar-sized watershed had permanent flows. They suggested that the difference in flow regimes was due to diminished interaction between the stream and floodplain with resultant lowering of the water table.</p>



PCE	PCE Habitat Feature	MPI Pathway	MPI Indicator	Rationale
				Indirect effects of historic livestock grazing in the ESA action area for bull trout (including trailing and watering), on channel and bank features such as bank stability, undercut banks and width to depth ratio, as well as impacts to shrub recruitment and green line plant vigor, have likely affected peak and base flows on some streams. However, it is anticipated that the proposed action of PE 1 (livestock use) will have <b>neutral effects</b> to this indicator, because of fences enclosing bull trout critical habitat along the MFJDR and because of non-use of riparian pastures during this consultation period. No riparian vegetation is affected by vehicle use, range riding or maintenance activities in proximity of bull trout critical habitat. Therefore, PE 2 will have <b>neutral effects</b> to the peak/base flow indicator. PE5 (monitoring) will not increase compaction or remove vegetation, and therefore does not have a mechanism to affect peak/base flow. The effect to the indicator is <b>neutral</b> .
2) Migration habitats with minimal physical, biological, or water quality impediments between spawning, rearing, overwintering, and freshwater and marine foraging habitats, including but not limited to permanent, partial, intermittent, or seasonal barriers.	Obstruction	Water Quality	Chemical contaminants/nutrients, temperature	<p><b>Chemical Contaminants/Nutrients:</b> Urine and feces from livestock use (PE 1) in riparian areas increases the likelihood that nitrogen and phosphorous will enter streams. Increased nutrients will likely increase stream productivity at the source of nutrients and for a short distance downstream. It is anticipated that livestock grazing will have slight negative impacts to the indicator, but they are <b>not expected to be meaningfully measured</b>.</p> <p>PE 2 (permittee management and infrastructure maintenance) includes vehicle use. The risk of chemical contamination to streams will be minimized by use of PDC. Maintenance activities are typically distant from designated CH, and at locations not hydrologically connected to stream channels. Therefore, there is no mechanism for petroleum products spilling from power tools to affect CH. Use of horses for range riding will have similar effects (but much smaller scale) than that of PE 1, above. Maintenance activities are typically distant from stream channels. The overall effect of PE 2 is for slight negative effects to the indicator that are not <b>expected to be meaningfully measured</b>.</p> <p>Monitoring (PE 5) does not involve the use of chemicals and does not have the potential to affect nutrients in streams. PE 5 will have a neutral effect to the indicator.</p> <p><b>Temperature:</b> Currently the majority of the temperature monitoring data for the Bear and York On/Off Allotments indicates that State water quality standards are not being met. Most of the streams monitored within the allotments would be considered to be functioning at unacceptable risk using the bull trout matrix criteria (see Section 6.1.3)</p>

PCE	PCE Habitat Feature	MPI Pathway	MPI Indicator	Rationale
				Bull trout use the MFJDR primarily as a migration corridor or when winter rearing. Temperature will be maintained by not removing shade producing vegetation (large cottonwoods) along the MFJDR and by encouraging future cottonwood regeneration/recruitment by fencing or resting riparian pastures for the life of the consultation. Therefore PE1 would have <b>neutral effects to this indicator</b> . PE 2 (permittee livestock management and infrastructure maintenance) and PE 5 (monitoring) activities will not remove vegetation that provides shade nor affect channel-forming processes that might widen stream channels. Consequently, there is no mechanism for PEs 2 and 5 to affect water temperature and the effect of the PE for the indicator is <b>neutral</b> .
		Flow/ Hydrology	Change in peak/base flows	See discussion on change in peak/base flows for PCE 1. The effects are the same.
3) An abundant food base, including terrestrial organisms of riparian origin, aquatic macroinvertebrates, and forage fish.	Forage	Water Quality, Habitat Elements, Channel Condition and Dynamics, Habitat Access	Temperature	See discussion on temperature for PCE 2. The effects are the same.
			Sediment	<p>The monitoring results presented above indicate that livestock use (PE 1), as well as use by wild ungulates, results in trampled and grazed riparian vegetation, and altered stream banks to some degree. Livestock also use trails to access streams for water. Livestock occasionally will concentrate their use in certain areas, potentially creating patches of relatively bare soil. Bare soil is prone to erosion and can result in fine sediment entering stream channels and resultant increases in turbidity. Habitat impacts are likely to include areas of exposed streambank up to a few feet wide where livestock access streams to drink or cross, and areas of bank disturbance where livestock graze in riparian areas. No such area along the MFJDR in the action area is a watergap in the C1/C2 pasture vicinity. Exposed areas and other bank disturbances that occur are likely to result in a <b>slight but insignificant increase</b> in turbidity for a short distance downstream during rainstorms or runoff events. However, given background levels of turbidity during runoff events it will be difficult to distinguish between turbidity resulting from these grazing impacts and background turbidity. A slight increase in fine sediment deposition for a short distance downstream of exposed and disturbed areas is also likely to occur.</p> <p>Endpoint indicators were developed in order to meet PACFISH / INFISH grazing standards and guidelines as well as water quality BMPs. The assumption is that meeting these endpoint indicators would move key riparian and stream channel elements (bank stability, w/d ratio, woody species regeneration) towards their Desired Conditions and meet Riparian Objectives. If monitoring fails to show this trend, adaptive management would be implemented and endpoint indicators would be modified to minimize adverse effects to CH.</p>

PCE	PCE Habitat Feature	MPI Pathway	MPI Indicator	Rationale
				<p>PE 2 involves use of vehicles on and off roads, as well as infrastructure maintenance. There is no potential for fine sediment to be transported from unpaved roads to stream channels because there are no unpaved roads that cross the MFJDR in the action area. Therefore the PE is <b>neutral to the indicator</b>.</p> <p>Monitoring (PE 5) activities such as pebble counts and measuring cross-sections involve wading in stream channels. Other monitoring activities involve walking or riding horses in riparian areas. Spawning surveys also involve wading. Wading may result in very small increases in turbidity downstream for a short distance (a few feet) that will quickly dissipate. Walking and riding horses in riparian areas should not result in fine sediment delivery to stream channels. However, there may be very small and transient increases in turbidity when a stream is being crossed. Within the allotment the monitoring PE effect to the indicator is negative, <b>but not meaningfully measured</b>.</p>
			Chemical contaminants/nutrients	See discussion on Chemical Contaminants/Nutrients for PCE 2. The effects are the same.
			Physical barriers	No barriers will be created or removed as part of the proposed action. The proposed action would have a <b>neutral effect</b> on physical barriers.
			Large woody debris	<p>Livestock grazing does not affect this indicator in conifer-dominated riparian forests. Livestock use can negatively affect this indicator when grazing occurs within hardwood stands such as aspen, alder, birch, and cottonwoods that could contribute larger pieces of wood to small streams. In sites in the action area that would be naturally dominated by cottonwood gallery riparian forests, such as the Middle Fork John Day River, livestock use (PE 1) could result in altering the level of cottonwood stocking and future large tree (and subsequent large woody debris) recruitment (Kaufmann et al. 1983, Case and Kaufmann 1997, Beschta and Ripple 2005). However, keeping cattle out of the riparian zone with electric fencing and/or resting the riparian pastures along the MFJDR will minimize loss of future cottonwood recruitment and effects to the large wood indicator are expected to be <b>negative but insignificant</b> in the Bear allotment. PE 2 and PE 5 do not affect trees and associated LWD in any way. Therefore there is no mechanism for an effect and the <b>effect is neutral</b> to the indicator for both PEs.</p>

PCE	PCE Habitat Feature	MPI Pathway	MPI Indicator	Rationale
			Pool frequency	<p>Pool frequency at the stream survey reach within the allotment is not currently meeting the DFC as described within Amendment 29 of the MNF LRMP and would be considered to be FAR using bull trout matrix criteria. See Appendix J for PIBO and stream survey results.</p> <p>Indirect effects of livestock grazing (including trailing and watering), on bank stability, undercut banks, width-depth ratio, shrub recruitment, green line plant composition and vigor have the potential to affect this indicator. Due to fencing and/or non-use of riparian pastures along the MFJDR in the Bear Allotment, the PE would have a <b>neutral effect</b> to this indicator.</p> <p>PE 2 (permittee management and infrastructure maintenance) includes on and off-road vehicle use. Road use has no mechanism to affect pool frequency. PDC for off-road use will prevent bank damage and effects to pool frequency. Range riding with horses will occasionally cross a stream but effects to streambanks will be so minimal as to not affect the indicator. Infrastructure maintenance actions do not affect streambanks or riparian vegetation adjacent to CH, and will therefore not affect pool frequency. The overall effect of PE 2 is a <b>neutral effect</b> to the indicator.</p> <p>PE 5 (monitoring) does not have any mechanisms to affect plants or bank and channel features that would impact pool frequency. The monitoring PE has a <b>neutral effect</b> to the indicator.</p>
			Pool quality	<p>Biologist observations indicate that undercut banks are less than the desired condition in the MFJDR in the action area, although some has been remedied by channel reconstruction in 2010. Residual pool depth was generally low with very few to no pools greater than 1 meter deep prior to the restoration project. Although Width/Depth ratio's at the PIBO sites within the John Day Basin seem to be improving most streams are still considered to be too wide. See Appendix J for PIBO and stream survey results. Based upon stream survey data in the MFJDR, pool quality would be considered to be FAR using bull trout matrix criteria.</p> <p>Indirect effects of livestock grazing (including trailing and watering), on bank stability, undercut banks, width-depth ratio, shrub recruitment, green line plant composition and vigor have the potential to affect this indicator. Due to fencing and/or non-use of riparian pastures along the MFJDR in the Bear Allotment, the PE would have a <b>neutral effect</b> to this indicator.</p>

PCE	PCE Habitat Feature	MPI Pathway	MPI Indicator	Rationale
				<p>PE 2 (permittee management and infrastructure maintenance) includes on and off road vehicle use. Road use has no mechanism to affect pool quality. PDC for off-road use will prevent bank damage and effects to pool quality. Range riding with horses will occasionally cross a stream but effects to streambanks will be so minimal as to not affect the indicator. Infrastructure maintenance actions do not affect streambanks or riparian vegetation adjacent to CH, and will therefore not affect pool frequency. The overall effect of PE 2 is a <b>neutral effect</b> to the indicator.</p> <p>PE 5 (monitoring) does not have any mechanisms to affect plants or bank and channel features that would impact pool quality. The monitoring PE has a <b>neutral effect</b> to the indicator.</p>
			Off-channel habitat	<p>Generally there is very little Off-Channel Habitat within the Middle Fork of the John Day. The use of BMP's for livestock management, end point indicators (which are inclusive of wild ungulate use), and adaptive management, should result in an overall effect by PE 1 (livestock use) to off-channel habitat that is <b>negative and not meaningfully measured</b>. This analysis assumes all off-channel habitats may not be fenced.</p> <p>PE 2 (permittee management and infrastructure maintenance) includes on and off road vehicle use. Road use has no mechanism to affect pool frequency. PDC for off-road use will prevent bank damage and effects to off-channel habitat. Range riding with horses will occasionally cross a stream but effects to streambanks will be so minimal as to not affect the indicator. Infrastructure maintenance actions do not affect streambanks or riparian vegetation adjacent to CH, and will therefore not affect pool frequency. The overall effect of PE 2 is a <b>neutral effect</b> to the indicator.</p> <p>PE 5 (monitoring) does not have any mechanisms to affect off-channel habitat. The monitoring PE has a <b>neutral effect</b> to the indicator.</p>
			Refugia	<p>Refugia is considered to be a limiting factor identified within the Draft Bull trout recovery plan (FWS 2002) and the John Day Sub-basin plan (NPPC 2005). All of the habitat indicators in this crosswalk table are potential components of Refugia.</p> <p>Analysis for previous indicators has determined that PE 1 (livestock use) will have negative but not meaningfully measured effects to several of them. This may occur in areas that meet the definition of Refugia. Therefore, PE 1 (livestock use) will have <b>negative but not meaningfully measured</b> in relation to the Middle Fork John Day River.</p>

PCE	PCE Habitat Feature	MPI Pathway	MPI Indicator	Rationale
				<p>The highest level of effect to previous indicators by PE 2 (permittee management and infrastructure maintenance) was “negative but not meaningfully measurable.” This level of effects will not impact the function of Refugia to provide important remnant habitat. Therefore, the effect conclusion is <b>neutral</b> for PE 2.</p> <p>The highest level of effect to previous indicators by PE 5 (monitoring) was “negative but not meaningfully measurable” for small and transient increases in turbidity by wading in stream channels or crossing streams on foot or by horse. This level of effects will not impact the function of Refugia to provide important remnant habitat. Therefore, the effect conclusion is <b>neutral</b> for the monitoring PE.</p>
			Width/Depth ratio	<p>Over-utilization of riparian vegetation, bank alteration and increases in sediment delivery are primary causes of increased W/D ratios. The MFJD within the allotment is still exhibiting over widened channel (See Appendix J. for PIBO and stream survey results.), although monitoring data is showing improvements for this indicator at the Basin scale. Legacy effects of livestock management in the allotment likely contributed to degradation of this indicator.</p> <p>Due to fencing and/or non-use of riparian pastures along the MFJDR in the Bear Allotment, the PE would have a <b>neutral effect</b> to this indicator. PE 2 (permittee management and infrastructure maintenance) includes on and off road vehicle use. Road use has no mechanism to affect W/D ratio. PDC for off-road use will prevent bank damage and effects to W/D ratio. Range riding with horses will occasionally cross a stream but effects to streambanks and beds will be so minimal as to not affect the indicator. Infrastructure maintenance actions do not affect streambanks or riparian vegetation adjacent to CH, and will therefore not affect W/D ratio. The overall effect of PE 2 is a <b>neutral effect</b> to the indicator. PE 5 (monitoring) does not remove vegetation or destabilize stream banks. There is no potential for it to increase W/D ratio. The monitoring PE will have a <b>neutral effect</b> to the indicator.</p>
			Floodplain connectivity	See discussion on Floodplain Connectivity for PCE 1. The effects are the same.
4) Complex river, stream, lake,	Complex Condition	Habitat	Large woody	See discussion on Large Woody Debris for PCE 3. The effects are the same.

PCE	PCE Habitat Feature	MPI Pathway	MPI Indicator	Rationale
reservoir, and marine shoreline aquatic environments and processes that establish and maintain these aquatic environments, with features such as large wood, side channels, pools, undercut banks and unembedded substrates, to provide a variety of depths, gradients, velocities, and structure.		Elements	debris	
			Pool frequency and quality	See discussion on Pool Frequency and Pool Quality for PCE 3. The effects are the same.
			Streambank condition	<p>Stream bank conditions are highly variable throughout the allotments. Historic land use practices (livestock grazing, mining, channelization, road building, timber production, and wildfire) have altered many of the streams as well as the Middle Fork of the John Day from its historic and or desired condition.</p> <p>Due to fencing and/or non-use of riparian pastures along the MFJDR in the Bear Allotment, the Livestock Use PE would have a <b>neutral effect</b> to this indicator.</p> <p>PE 2 (permittee management and infrastructure maintenance) includes on and off road vehicle use. Road use has no mechanism to affect streambank stability. PDC for off-road use will prevent bank damage and effects to streambank stability. Range riding with horses will occasionally cross a stream but effects to streambanks will be so minimal as to not affect the indicator. Infrastructure maintenance actions do not affect streambanks or riparian vegetation adjacent to CH, and will therefore not affect streambank stability. The overall effect of PE 2 is a <b>neutral effect</b> to the indicator. PE 5 (monitoring) does not remove vegetation or destabilize streambanks. There is no potential for it to reduce streambank stability.</p>

PCE	PCE Habitat Feature	MPI Pathway	MPI Indicator	Rationale
				The monitoring PE will have a <b>neutral effect</b> to the indicator.
			Large pools	<p>PIBO and stream survey data within the Allotments indicates that residual pool depth is generally low with very few to no pools greater than 1 meter deep. Historic land use practices (livestock grazing, mining, channelization, road building, timber production, and wildfire) have altered many of the streams as well as the Middle Fork of the John Day from its historic and or desired condition. Many of these streams are still in a state of recovery based on stream type, soil type, and vegetation seral stage.</p> <p>Mechanisms to affect large pools include impacts to large woody debris and increased sedimentation. Effects to the large woody debris indicator are described above (see PCE 3 analysis) as <b>negative, but not meaningfully measured</b> for PE 1 and neutral for PEs 2 and 5. The sediment analysis for PCE 3 concluded that there would be <b>negative, but not meaningfully measured</b> effects for PE1 and neutral and negative but not meaningfully measured effects for PEs 2 and 5 respectively. The amounts of sediment generated are expected to be slight and affect the streambed for a short distance downstream, and are not likely to measurably affect pool depth. This will result in <b>negative, but not meaningfully measured</b> effects to the Large Pools indicator.</p>
			Off-channel habitat	See discussion on Off-channel Habitat for PCE 3. The effects are the same.
			Refugia	See discussion on Refugia for PCE 3. The effects are the same.
5). Water temperatures ranging from 2 to 15° C (36 to 59 ° F),	Water Quality	Water Quality	Temperature	See discussion above.



PCE	PCE Habitat Feature	MPI Pathway	MPI Indicator	Rationale
with adequate thermal refugia available for temperatures that exceed the upper end of this range. Specific temperatures within this range will vary depending on bull trout life-history stage and form; geography; elevation; diurnal and seasonal variation; shading, such as that provided by riparian habitat; and local groundwater influence.				<p>Livestock use (PE 1) is not likely to result in measurable water temperature increases for the MFJD River due to its width. Fencing and/or non-use will prevent removal of shade-producing vegetation and consequently this PE is <b>neutral</b> for the temperature indicator.</p> <p>PE 2 (permittee livestock management and infrastructure maintenance) and PE 5 (monitoring) activities will not remove vegetation that provides shade nor affect channel-forming processes that might widen stream channels. Consequently, there is no mechanism for PEs 2 and 5 to affect water temperature and the effect of the PE for the indicator is <b>neutral</b>.</p>
6). In spawning and rearing areas, substrate of sufficient amount, size, and composition to ensure success of egg and embryo overwinter survival, fry emergence, and young-of-the-year and juvenile survival. A minimal amount of fine sediment, generally ranging in size from	Suitable Substrate	Water Quality	Sediment	<p>There are no PIBO sites within the MFJDR portion of the Allotment.</p> <p>Stream surveys conducted in the MFJD reaches of bull trout CH reported Gravel as the dominate substrate. Low pool frequency, poor pool quality, and high water temperatures may be a limiting factor for bull trout within the allotment.</p> <p>Analysis for the Sediment indicator for PCE 3 above concluded that there would be negative but not meaningfully measured effects for PE 1 and negative but not meaningfully measured effects for PEs 2 and 5. The amounts of sediment generated, if any, are expected to be slight and affect the streambed for a short distance downstream. This will result in <b>negative, but not meaningfully measured effects to the sediment indicator</b> with respect to suitable substrate for bull trout spawning, particularly as there is no spawning within the allotment in the MFJDR.</p>

PCE	PCE Habitat Feature	MPI Pathway	MPI Indicator	Rationale
silt to coarse sand, embedded in larger substrates, is characteristic of these conditions. The size and amounts of fine sediment suitable to bull trout will likely vary from system to system.				
		Habitat Elements	Substrate Embeddedness	<p>Amendment 29 of the LMRP sets a desired condition of &lt;20% embedded.</p> <p>Currently there are no data to discuss substrate embeddedness within the allotment.</p> <p>The analysis of effects to the sediment/turbidity indicator (see analysis for PCE 3) determined that use of riparian areas by livestock is expected to increase the amount of sediment entering streams. A slight increase in fine sediment deposition for a short distance downstream of exposed and disturbed streambank areas is likely to occur. There is the potential for fine sediment to slightly increase embeddedness within gravels suitable for spawning when the gravel is located immediately downstream from exposed and disturbed streambank areas. Therefore, the effect to this indicator by livestock use (PE 1) is <b>negative but not meaningfully measurable</b>.</p> <p>The analysis for sediment/turbidity determined that PE 2 would have a <b>slightly negative, but not meaningfully measured effect</b> to the indicator. Therefore, the same conclusion is made for the substrate embeddedness indicator. As described above, monitoring (PE 5) would not introduce fine sediment into stream channels. The monitoring PE will have a <b>neutral effect</b> to the indicator.</p>

PCE	PCE Habitat Feature	MPI Pathway	MPI Indicator	Rationale
7). A natural hydrograph, including peak, high, low, and base flows within historic and seasonal ranges or, if flows are controlled, minimal flow departure from a natural hydrograph.	Water Quantity	Flow/ Hydrology	Change in peak/base flows	See discussion for the Change in Peak/Base Flows indicator for PCE 1. The effects are the same.
8). Sufficient water quality and quantity such that normal reproduction, growth, and survival are not inhibited.	Water Quality	Water Quality	Temperature	See discussion for the Temperature indicator for PCE 2 and the Refugia indicator for PCE 2. The effects are the same.
			Sediment	See discussion for the Sediment indicator for PCE 3. The effects are the same.
			Chemical Contaminants/ Nutrients	See discussion for the Chemical Contaminants/Nutrients indicator sediment for PCE 2. The effects are the same.
	Water Quantity	Flow/ Hydrology	Change in Peak/Base Flows	See discussion for the Change in Peak/Base Flows indicator for PCE 1. The effects are the same.

PCE	PCE Habitat Feature	MPI Pathway	MPI Indicator	Rationale
9). Sufficiently low levels of occurrence of nonnative predatory (e.g., lake trout, walleye, northern pike, smallmouth bass); interbreeding (e.g., brook trout); or competing (e.g., brown trout) species that, if present, are adequately temporally and spatially isolated from bull trout.	Species	Subpopulation characteristics	Persistence and Genetic Integrity	There is no element of the federal action that will introduce, or alter habitat conditions to favor, non-native predatory fish species, fish species with the potential for inbreeding with bull trout, or fish species with the potential to compete with bull trout. Therefore there is no causal mechanism to affect this PCE and thus the <b>effect is neutral</b> .

TABLE 25. SUMMARY OF EFFECTS OF THE PROPOSED ACTION BY THE PROJECT ELEMENTS OF LIVESTOCK GRAZING IN THE BEAR ALLOTMENT TO THE INDICATORS ASSOCIATED WITH HABITAT FEATURES OF EACH PRIMARY CONSTITUENT ELEMENT OF BULL TROUT CRITICAL HABITAT.

Primary Constituent Element	PCE Habitat Feature	Indicator	Effect Conclusion by Project Element		
			PE1: Livestock Use	PE2: Permittee Management of Livestock and Infrastructure Maintenance	PE 5: Monitoring
1. Freshwater spawning sites with water quantity and quality conditions and substrate supporting spawning, incubation and larval development	Water quantity	Changes in Peak/Base Flows	Neutral	Neutral	Neutral
		Increase in Drainage Network	Neutral	Neutral	Neutral
	Water quality	Temperature	Neutral	Neutral	Neutral

Primary Constituent Element	PCE Habitat Feature	Indicator	Effect Conclusion by Project Element		
			PE1: Livestock Use	PE2: Permittee Management of Livestock and Infrastructure Maintenance	PE 5: Monitoring
		Sediment/Turbidity	NNMM	Neutral	NNMM
		Chemical Contamination/Nutrients	NNMM	NNMM	Neutral
	Suitable substrate	Substrate Embeddedness	NNMM	NNMM	Neutral
2. Freshwater rearing sites with:  (i) Water quantity and floodplain connectivity to form and maintain physical habitat conditions and support juvenile growth and mobility;  (ii) Water quality and forage supporting juvenile development; and  (iii) Natural cover such as shade, submerged and overhanging large wood, log jams and beaver dams, aquatic vegetation, large rocks and boulders, side channels, and undercut	Flow/hydrology	Changes in Peak/Base Flows	Neutral	Neutral	Neutral
		Increase in Drainage Network	Neutral	Neutral	Neutral
	Water quality	Temperature	Neutral	Neutral	Neutral
		Sediment/Turbidity	NNMM	Neutral	NNMM
		Chemical Contamination/Nutrients	NNMM	NNMM	Neutral
	Floodplain connectivity	Floodplain Connectivity	NNMM	Neutral	Neutral
		Width/Depth Ratio	NNMM	Neutral	Neutral
	Forage	Substrate Embeddedness	NNMM	NNMM	Neutral
		Large Woody Debris	NNMM	Neutral	Neutral
		Pool Frequency	Neutral	Neutral	Neutral

Primary Constituent Element	PCE Habitat Feature	Indicator	Effect Conclusion by Project Element		
			PE1: Livestock Use	PE2: Permittee Management of Livestock and Infrastructure Maintenance	PE 5: Monitoring
banks		Pool Quality	Neutral	Neutral	Neutral
		Off-Channel Habitat	NNMM	Neutral	Neutral
		Refugia	NNMM	Neutral	Neutral
		Riparian Reserves	NNMM	Neutral	Neutral
	Natural cover	Substrate	NNMM	NNMM	Neutral
		Large Woody Debris	NNMM	Neutral	Neutral
		Pool Frequency	Neutral	Neutral	Neutral
		Pool Quality	Neutral	Neutral	Neutral
		Off-Channel Habitat	NNMM	Neutral	Neutral
		Refugia	NNMM	Neutral	Neutral
		Riparian Reserves	NNMM	Neutral	Neutral
3. Freshwater migration corridors free of obstruction and excessive predation with water quantity and quality conditions and natural cover such as submerged and overhanging large wood, aquatic vegetation, large rocks and boulders, side	Migration corridors free of obstruction	Physical Barriers	Neutral	Neutral	Neutral

Primary Constituent Element	PCE Habitat Feature	Indicator	Effect Conclusion by Project Element		
			PE1: Livestock Use	PE2: Permittee Management of Livestock and Infrastructure Maintenance	PE 5: Monitoring
channels, and undercut banks supporting juvenile and adult mobility and survival					

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## 7.3 ANALYSIS OF EFFECTS TO LISTED SPECIES

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### 7.3.1 GENERAL EFFECTS

Effects to MCR steelhead and CR bull trout from livestock grazing can be in the form of direct impacts to individual fish or indirectly through habitat disturbance. Direct disturbance includes trampling on redds, resulting in injury or death to incubating embryos or alevins; disturbing holding or spawning adults, forcing them to alter their behavior and seek cover; or disturbing rearing juveniles, forcing them to alter their behavior and seek cover.

Grazing can have a number of detrimental effects on riparian and aquatic habitat (Platts 1991). When riparian habitat is adversely affected, the survival and growth of listed fish species may also be negatively affected. For example, if temperatures increase to critical levels due to reduced shade, salmonid survival can decrease and some habitat may be abandoned as fish migrate to seek cooler temperatures. Loss of overhead cover in the form of overhanging vegetation or undercut banks is likely to result in increased predation of juvenile salmonids. Increases in fine sediment are likely to increase turbidity that can alter salmonid behavior, and is also likely to increase fine sediment in spawning gravels that decreases egg-to-fry survival.

However, the livestock grazing end-point indicators were developed to meet PACFISH grazing standards and guidelines, enclosure B of the LMRP and water quality BMPs. The assumption is that meeting the endpoint indicators would move key riparian and stream channel elements (bank stability, width to depth ratio, woody species regeneration) towards their Desired Conditions and meet Riparian Objectives. This will allow recovery of degraded riparian habitat to occur. Recovery of riparian vegetation results in the development of more complex habitat. Riparian recovery allows roots to stabilize streambanks, and stems and foliage to slow water velocities, trap fine sediment, provide overhead cover for fish, provide shade that may aid in keeping stream temperatures cool, and provide surfaces for macroinvertebrates to inhabit. Stable stream banks and fine sediment trapping result in less fine sediment in spawning substrate that would improve egg-to-fry survival (Tappel and Bjornn 1983, Bjornn and Reiser 1991). Reduced water velocities along stream edges increase the amount of available habitat for young salmonids (Bjorn and Reiser 1991). Spawning salmonids appear to prefer spawning in areas in close proximity of overhead cover (Bjornn and Reiser 1991), and overhead cover protects juvenile salmonids from predation. Shade provided by vegetation can be important in keeping stream temperatures cool for salmonids. Li et al. (1994) found that trout abundance decreased as solar input and water temperature increased. Macroinvertebrates inhabiting overhanging vegetation provide forage for juvenile salmonids when they fall into the stream. Each of these benefits contributes to increasing the amount and quality of habitat available for all freshwater life stages of salmonids such as MCR steelhead, CR bull trout, and spring Chinook salmon.

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### 7.3.2 DIRECT EFFECTS TO SPECIES

Two of the three allotments (York On/Off and Bear) contain MCR Steelhead spawning and rearing habitat. The Bear allotment also contains bull trout Critical Habitat and spring Chinook Essential Fish Habitat. At certain times and under various conditions it is possible for livestock use (PE 1) to directly impact listed MCR steelhead. These effects could manifest themselves as direct impacts to individual fish, fry, or incubating embryos.



Direct impacts are likely to occur if livestock wade into a stream and disturb rearing juveniles or spawning adults, and/or step on redds. Juveniles in close proximity to stream crossings or watering sites are likely to move out of an area when livestock enter or approach the stream. Juveniles are likely to be at increased risk of predation. Livestock will have access to documented<sup>1</sup> spawning habitat in Slide Creek the York On/Off Allotment during the spawning period. Even though cattle use is relatively low in this allotment (79 AUMs) the probability cattle will interact with redds is not discountable. Effects if cattle did trample redds is not insignificant (e. g. Gregory and Gamett 2009), and therefore results in an LAA determination for direct effects to the species for the York On/Off Allotment. For the Bear allotment, potential for cattle-fish interactions is discountable due to the fencing or non-use of the riparian pastures along the MFJDR and the lack of documented steelhead spawning in Mosquito Creek (Bear Allotment). Thus direct effects to the species are discountable in this allotment (but see indirect effects to habitat below in 7.3.3).

It is possible that MCR steelhead spawning behavior could be interrupted, forcing adults to retreat to nearby cover, and that redds will be at risk of being stepped on by cattle, potentially causing direct mortality (Gregory and Gamett 2009). However, these risks will be minimized with the implementation of the *Malheur National Forest Strategy to Minimize Redd Trampling "Take" of Steelhead and Bull Trout* (Appendix F). Additionally, MNF staff (range and aquatic specialists) will take extra effort to monitor these sites when they are in the field. If active redds are located, mitigation actions will be taken to eliminate or significantly minimize the potential for redd trampling (PDC 5 and 6).

The potential for direct impacts from PE 2 (permittee management and infrastructure maintenance) is much smaller. Road use has no potential for direct impacts to the species. The PDCs do not allow off-road vehicles to cross streams except for use of existing fords on road crossings. Range riders on horses will occasionally cross streams, but redds will be identified and avoided, and any disturbance to adults or juveniles should be sufficiently brief to not result in significant disruption of normal behavioral patterns. Infrastructure maintenance actions are not located in stream channels, so there is no mechanism for direct impacts to the species.

Some monitoring activities (PE 5) involve walking in stream channels. Actions such as pebble counts and redd surveys will result in individuals walking across stream channels for time periods that may result in MCR steelhead being disturbed and moving out of the area, resulting in direct impacts to the species. Surveyors will search for steelhead redds when walking in streams and avoid if present.

Direct impacts to bull trout are unlikely in the action area. Due to warm temperatures in the MFJDR, bull trout are likely only present in the cooler seasons when cattle are absent from the allotments. The only stream bull trout may use in the action area is the MFJDR, primarily as a migration corridor or as winter rearing. Therefore PE 1, 2, and 5 would have a **neutral effect** to bull trout individuals.

<sup>1</sup> A field visit on 20 September 2011 revealed the presence of numerous juvenile steelhead/redband trout in the upper portion of Slide Creek in the York Allotment. Spawning habitat in this area is presumed.

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### 7.3.3 DIRECT AND INDIRECT EFFECTS TO AQUATIC AND RIPARIAN HABITAT

Use of the NMFS MPI to determine effects to listed fish species is based upon using the effects of the action on habitat indicators as a surrogate for effects to the species. The premise is that the

indicators and the range of environmental baseline conditions provided by the three classifications (PF, AR, NPF) depict the biological requirements of the listed fish species. Since there is a direct relationship between habitat condition and the growth and survival of individual fish at various life stages, the effects of the action on habitat variables can be linked to effects to individuals of the species, and ultimately to an ESA effect determination.

The analysis in the “Effects to Critical Habitat” section (Section 7.2) evaluated specific NMFS MPI indicators that correspond to the PCEs of CH. The PCEs are used to describe “those physical or biological features that are essential to the conservation of the listed species.” The same sub-set of NMFS MPI indicators evaluated for effects to PCEs also apply to the analysis of effects to the species. To eliminate redundancy, only those indicator/PE combinations for which a conclusion of effect to a component of a PCE was “negative and meaningfully measured” will be brought forward for further evaluation in this section, as they have the potential to adversely affect listed MCR Steelhead. These conclusions were only found for PE 1 (livestock use) and not for PE 2 (permittee management and infrastructure maintenance) or PE 5 (monitoring). The indicators for which “negative and meaningfully measured” effects were concluded are:

- Temperature
- Sediment/turbidity
- Substrate embeddedness
- Large woody debris
- Refugia

#### 7.3.3.1 EFFECTS ON WATER TEMPERATURE

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Water temperature is an important factor affecting distribution and abundance of salmonids within the action area. Water temperatures influence water chemistry, as well as every phase of salmonid life history. Optimal temperatures for steelhead are 50° to 61° F (10° to 16° C), and the lethal temperature is approximately 77° F (25° C) where as temperatures for bull trout are much less; typically ranging from 36 to 59° F (2 to 15° C). Water temperatures in excess of 59° F (15° C) are thought to limit the distribution of bull trout (Buchanan and Gregory 1997). Light et al. (1996) reported that temperatures above 59° F (15° C) may favor competitive species, while sustained temperatures above 64° F (18° C) may cause harmful effects to bull trout. They further report that water temperatures above 77° F (25° C) are expected to be lethal to bull trout. These temperature ranges may vary depending on bull trout life history stage and form, geography, elevation, diurnal and seasonal variation, shade, such as that is provided by riparian habitat, and local groundwater influence. Within the action area, high stream temperatures occur near the end of July or the beginning of August and coincide with low stream flows and warm daytime temperatures. By the end of August, stream temperatures are typically dropping. Criteria for anadromous salmonid freshwater temperatures are found in the NMFS/USFWS matrix above.

Temperature monitoring information for MNF lands within or directly adjacent to the allotments is derived from five sites that were monitored sporadically from 1998 to 2008. The MFJDR is on the Oregon Department of Environmental Quality 303(d) list for water temperature (ODEQ 2010). Analysis of the temperature monitoring information at all sites determined they were NPF under NMFS MPI criteria (see Section 6.1.3).

Stream temperature is driven by the interaction of site conditions, weather, riparian vegetation, and the input of radiant energy to a stream system. Energy exchange that affects a change in

water temperature may involve solar radiation, long wave radiation, evaporative heat transfer, convective heat transfer, conduction, and advection (Lee 1980; Beschta and Weathered 1984). Solar radiation is the most important source of radiant energy affecting stream temperature (Brown 1969; Beschta 1997). With the exception of solar radiation that only delivers heat energy, all the other processes are capable of both introducing and removing heat from a stream. While the process of introducing and removing heat from a stream is complex, certain processes are more important than others in determining how stream temperature is affected by solar inputs (Beschta et al. 1987). In terms of water temperature increases, the principal source of heat energy is solar radiation directly striking the stream (Brown 1972) (Figure 6.<sup>10</sup>)

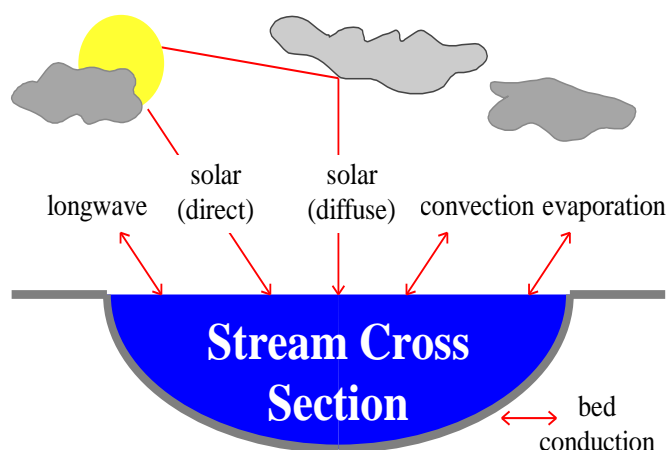


FIGURE 6. HEAT EXCHANGE BETWEEN A STREAM AND ITS ENVIRONMENT.

Canopy density and height are the dominant factors in the ability of streamside vegetation to intercept incoming solar radiation and reduce the rate of warming. Decline in the abundance and vigor of riparian plants in a floodplain may also cause streams to become shallow and wide, which increases the surface area that is exposed to solar radiation. Platts (1981) cited studies by Claire and Storch (1977) and others that found that removal of streamside vegetation contributed to increases in water temperatures in small headwater streams as well as influencing suspended sediment concentrations. Small streams (especially Rosgen C and E channels such as Slide Creek) are more susceptible to warming because they have a lower volume of water to absorb solar energy. They are also more susceptible to warming because grazing impacts herbaceous vegetation and shrubs that typically provide shade to the stream channel.

Effective shade is the total solar radiation blocked from reaching the stream over a twenty-four hour period, expressed as a percentage of the total solar radiation:

$$\text{Effective Shade} = \frac{\text{Total Solar Radiation} - \text{Total Solar Radiation Reaching the Stream}}{\text{Total Solar Radiation}}$$

<sup>10</sup> Stream temperature is an expression of heat energy per unit volume, which in turn is an indication of the rate of heat exchange between a stream and its environment.

### Total Solar Radiation

Effective shade is provided by features such as topography and vegetation (ODEQ 2010). Effective shade is influenced by slope steepness, vegetation species composition, tree height, vegetation density, tree distance from the stream bank, and stream width. Thus, although riparian vegetation is a physical barrier between the stream and incoming solar radiation, only a portion of the riparian canopy contributes to effective shade. The relationship of variables influencing effective shade can be simplified, to some degree, using geometry and computer models that simulate shade (Park 1993, Boyd 1996).

Figure 7 and Table 26 illustrate the relationship between shade and stream channel width. As stream channel width increases beyond the point where vegetation is not tall enough to cast a shadow across the stream channel, shade values decrease. The model analysis results in Table 26 are based on the shadow cast by vegetation at a distance of 1 foot and farther from the edge of the channel.

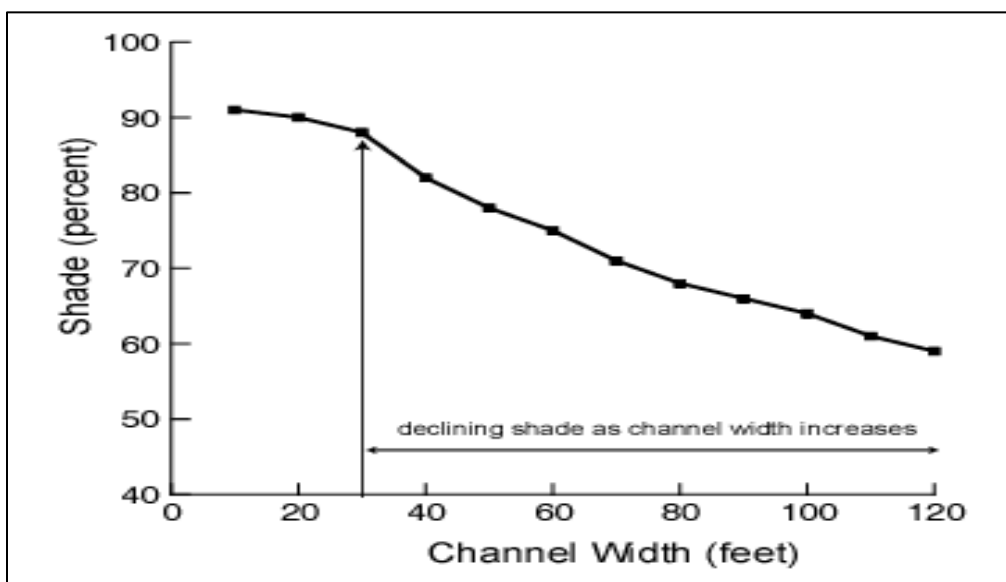


FIGURE 7. SHADE PROVIDED BY 150-FOOT TALL CONIFERS (PLATTS ET AL 1987), (PARK, 1993).

TABLE 26. EFFECTIVE SHADE PROVIDED BY THREE HEIGHTS OF GREENLINE VEGETATION AT VARYING ACTIVE STREAM CHANNEL WIDTHS.<sup>1</sup>

Active Channel Width (feet)	Percent Effective Shade at Varying Vegetation Heights		
	0.5 feet	2 feet	3 feet
1	0	46	57
3	0	22	41
7	0	9	18
10	0	7	12

12	0	5	10
14	0	5	9
16	0	4	8
18	0	4	7
20	0	3	6
22	0	0	3
24	0	0	1
26	0	0	0

<sup>1</sup>Effective shade values are based on the shadow cast by continuous vegetation growing at a distance of 1 foot and farther from the edge of the channel

Figure 7, Table 26 and the discussion above illustrate that decreasing vegetation height will reduce effective shade, and as a result increase solar radiation to the surface of the water in the stream. If the potential height of ungrazed vegetation is in the 2-3 feet range, then a considerable loss of effective shade takes place when the vegetation is grazed to an end-point of 0.5 feet (six inches) or less. The potential for stream warming is greatly increased, particularly for streams with narrow active channel widths.

Ungrazed height for 13 grass or grass-like plant species in the MNF ranged from 14 to 36 inches, with a mean of 26 inches (MNF 2007). The vegetation heights in Table 26 of two and three feet encompass the approximate mean ungrazed height and maximum ungrazed height of the 13 species. An end-point of 6 inch (0.5 feet) stubble height is used when livestock are grazed, which also reflects use by wild ungulates. In meadow streams with narrow channels, grass/grass-like species often are the plants that provide stream shade. Model results presented in Table 26 indicate that managing to a six-inch vegetation height will reduce effective shade to zero for channel widths that are 1 foot and greater when the model criterion is for vegetation beginning at 1 foot distance from the active channel, and to no more than 18 percent if modeled for vegetation at the edge of the active channel. This is considerably less than effective shade provided by potential vegetation heights of two to three feet for similar active channel widths.

A logical conclusion is that implementation of PE 1 (livestock use) will reduce effective shade for a bank distance sufficient to result in a measurable water temperature increase. These impacts are expected to be generally confined to narrow stream channels with grass/grass-like vegetation providing shade. In the action area, this would include portions of Slide Creek.

Platts (1991) states that grasses are too short to keep much solar radiation from reaching the water, except along very small streams (stream orders 1 and 2). Wright and Li (2002) measured wetted widths in late July and early August 1996-1998 for five 1<sup>st</sup> and 2<sup>nd</sup> order streams in the MFJD River drainage. The mean wetted width was 2.0 meters with a standard error of 0.5 meter (6.6 feet with a standard error of 1.6 feet). There is 90% confidence that the true mean wetted width is within 6.6 +/- 3.4 feet (3.2 to 10.0 feet). The greatest probability of measurable

temperature increases as a result of livestock grazing is therefore likely to occur in channels less than 10 feet wide. Portions of Slide and Mosquito creeks in the action area have summer wetted widths less than 10 feet wide even though these streams may be greater than second order.

Figure 7 and Table 26 also indicate that effective shade is reduced as channel width increases. Grazing by large hooved animals has the potential to increase channel width by bank alteration (Armour et al. 1991; Clary and Webster 1989; Kauffman and Kruger 1984). However, the analysis presented for effects to width-depth ratio in Table 22 and 28 for the Floodplain Connectivity habitat feature of PCE 2 (rearing critical habitat) concluded that the livestock use PE is anticipated to have a negative, but not meaningfully measured effect to the indicator. The potential for increases in width-depth ratio is mitigated because of implementation of endpoint indicators for livestock grazing (which includes all use by wild ungulates) and adaptive management. However, due to endpoint indicators and adaptive management, effective shade will not be measurably reduced as a result of increasing the width-to-depth ratio indicator.

**Conclusion.** The discussion above described negative effects to habitat and vegetation characteristics from the livestock use PE. Effects to these characteristics result in negative impacts (increases) to summer water temperature. It is probable that livestock use will result in small, but measurable increases in water temperature in streams with narrow channels (<10 feet) where grass/grass-like vegetation is providing stream shade such as portions of Slide Creek. This will occur as a direct result of reducing the height of shade-producing vegetation such as grasses and sedges by grazing.

These effects to water temperature will be minimized by use of endpoint indicators and PDC. As described earlier, if pre-season monitoring indicates that wild ungulate use is resulting in measurements near or exceeding an endpoint indicator, cattle will not be turned-out into that specific pasture. These indicators were developed to meet PACFISH grazing standards and guidelines as well as water quality BMPs. The assumption is that meeting these endpoint indicators would move key riparian and stream channel elements (bank stability, w/d ratio, woody species regeneration) towards their Desired Conditions and meet Riparian Objectives. If monitoring fails to show this upward trend, adaptive management and administrative actions would be implemented to continue to minimize adverse effects to CH and the listed MCR steelhead. Some impacts from past management activities (e.g. logging, roads, grazing) will persist over the life of this consultation and likely much longer in some cases.

#### 7.3.3.2 EFFECTS ON SEDIMENT/TURBIDITY, SUBSTRATE AND SUBSTRATE EMBEDDEDNESS

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Grazing by large herbivores such as domestic cattle can result in hoof shear to streambanks, and trampling and consumption of streamside vegetation. The result is a potential increase in the supply of fine sediment available for transport. This can occur when grazing results in compacted soils and bare areas; and when grazing results in decreased bank stability through mechanical damage to streambanks or reductions in rooting strength of streambank stabilizing vegetation. Both result in an increase in erosion rates and subsequent increases in fine sediment levels in streams.

Small amounts of fine sediment are also likely to enter streams where livestock access streams to cross or water. Small amounts of fine sediment are likely to become deposited in substrate that

can decrease egg-to-fry survival and slightly reduce available substrate cover for juveniles and macroinvertebrates.

Increased fine sediment is detrimental to MCR steelhead through increased turbidity and sediment deposition in the substrate (NMFS 2009). Increases in fine sediment lead to greater substrate embeddedness and a decrease in the interstitial spaces between gravel substrate important for salmonid spawning. Successful salmonid spawning requires clean gravels with low fine sediment content (Spence et al. 1996). Well-oxygenated water must be able to reach eggs and pre-emergent fry during incubation and emergence. Suffocation of these life stages may occur if redds become covered with fine sediment. Emerging fry may be physically blocked from escaping a redd. Increased sediment load is also detrimental to juvenile salmon by introducing suspended particulate matter that interferes with feeding and territorial behavior (Berg and Northcote 1985). Increased fine sediment deposition in the substrate is likely to decrease MCR steelhead egg-to-fry survival (Reiser and White 1988, Spence et al. 1996).

In addition, inputs of fine sediment resulting from livestock trampling banks can reduce benthic invertebrate abundance and lead to a shift from aquatic insects to mollusks, which are less palatable to salmonids. Studies have shown that sediment inputs resulting in substrate embeddedness of greater than one-third can result in a decrease in benthic invertebrate abundance and thus a decrease in food available for juvenile salmonids (e.g. Waters 1995). Other field studies have shown small amounts of fine sediment and corresponding embeddedness have measurable effects on juvenile steelhead growth, behavior, and survival (Suttle et al. 2004).

There are no streams in the proposed action area that have been identified on the 303(d) list for sedimentation. Fine sediment levels vary across the allotments depending on local geology, stream type, and management history.

**Conclusion.** The livestock use PE will likely result in fine sediment entering stream channels where MCR steelhead are present, particularly in Slide Creek of the York On/Off Allotment. It will likely also result in sediment entering areas where steelhead may not currently be present but its CH is designated. These areas could include Mosquito Creek in the Bear Allotment. The mechanisms include: 1) mechanical bank damage from hoof chisel and trampling; 2) trailing; and, 3) impacts to soil-holding vegetation by being eaten and trampled. These mechanisms negatively impact bank stability, resulting in increased erosion and sedimentation. The increases in fine sediment will negatively and measurably affect the Sediment/Turbidity and Substrate Embeddedness NMFS MPI indicators. This analysis understands that wild ungulate use may also significantly and negatively affect the sediment indicator in some of these streams.

These effects to the Sediment/Turbidity and Substrate Embeddedness indicators will be minimized by use of endpoint indicators and PDC. If pre-season monitoring indicates that wild ungulate use is resulting in measurements near or exceeding an endpoint indicator, cattle will not be turned-out into that specific pasture. These indicators were developed to meet PACFISH grazing standards and guidelines as well as water quality BMPs. The assumption is that meeting these endpoint indicators would move key riparian and stream channel elements (bank stability, w/d ratio, woody species regeneration) towards their Desired Conditions and meet Riparian Objectives. If monitoring fails to show this upward trend, adaptive management and administrative actions would be implemented to continue to minimize adverse effects to designated CH and the listed MCR steelhead. Some impacts from past management activities



(e.g. logging, roads, grazing) will persist over the life of this consultation and likely much longer in some cases.

#### 7.3.3.3 EFFECTS ON REFUGIA

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The concept of “Refugia” is not described in detail in the NMFS MPI (NMFS 1996) (see Table 11 earlier in this document). The definition provided therein is: “important remnant habitat for sensitive aquatic species.” The availability of various types of habitat refugia are described as limiting factors in the recovery plan for the Oregon steelhead populations of the MCR Steelhead DPS (e.g., loss of side-channels that provided high flow refugia; cold water refugia provided by Columbia River tributary streams such as the Deschutes River (NMFS 2009). In the action area, MSRAs may qualify as refugia because these areas have some of the highest steelhead habitat potential due to their low gradient and other geomorphic qualities.

The analysis of effects to PCEs of CH provided in Table 22 and 28, and summarized in Table 23 and 29, indicate that the livestock use PE will have negative and meaningfully measured effects to several of the MPI indicators that correlate to components of PCEs. Specifically, they are “Water Temperature,” “Sediment/Turbidity,” and “Substrate Embeddedness.” This may occur in stream reaches providing refugia conditions for one or more of these habitat characteristics (areas with cooler water temperatures, low levels of sediment in substrate or the water column, and low levels of substrate embeddedness). Therefore, PE 1 will have a negative effect to the Refugia indicator.

**Conclusion.** The livestock use PE will result in negative and meaningfully measured impacts to several habitat indicators associated with Refugia. Consequently, there will be negative and meaningfully measured, evaluated or detected impacts to the Refugia indicator. The effects are not expected to be distributed evenly across the ESA action area because stream reaches providing characteristics of refugia are not ubiquitous. The effects will likely be observed in areas where adequate cattle forage overlaps MSRAs or other refugia such as seeps or springs used by steelhead. Negative impacts to the Refugia indicator will be minimized by use of the endpoint indicators and PDC.

#### 7.3.3.4 EFFECTS ON LARGE WOODY DEBRIS

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Large woody debris (aka large wood) is one of the most important habitat components in many fish-bearing streams (Gurnell et al. 2002). Large wood helps provide cover, scour pools, stabilize banks, retain spawning gravels, create off-channel habitats, and provide habitat for macroinvertebrate production (Gregory et al. 2003).

In streams within the action area, large wood is usually provided by fallen conifers that have no effect from the project elements. However, in some areas where hardwoods—particularly black cottonwood and quaking aspen—play an important role in riparian species composition, ungulate grazing can prevent future large wood recruitment by limiting sapling regeneration and large tree recruitment. Young cottonwoods are desirable forage to both domestic and wild ungulates (Braatne et al. 1996).

Kaufmann et al. (1983) found late season riparian cattle grazing retarded regeneration of black cottonwood saplings in northeastern Oregon. Another study found when cattle were removed



from a riparian pastures, but wild ungulates were not exclosed, the number of black cottonwood seedlings/saplings increased 56% 3 years after livestock removal (Case and Kaufmann 1997). Clearly, cattle grazing can influence the abundance of black cottonwoods in a riparian area, which can have measurable and foreseeable future effects to riparian structure and future large wood recruitment. Beschta and Ripple (2005) surveyed a 40-mile reach of the Middle Fork John Day River in and adjacent to the action area for cottonwood abundance and stand structure and found very little cottonwood seedling/sapling regeneration or recruitment into large trees and described wild and domestic ungulate browsing as the primary causal factor.

The analysis of effects to PCEs of CH provided in Table 22 and 28, and summarized in Table 23 and 29, indicate that the livestock use PE will have negative and meaningfully measured effects to the “Large Woody Debris” MPI indicator that correlates to components of PCEs. Therefore, PE 1 will have a negative effect to the Large Woody Debris indicator.

**Conclusion.** The livestock use PE will likely result in negative effects to future large woody debris recruitment. The effects will likely be observed in areas where adequate cattle forage overlaps low-gradient stream sections such as MSRAs that have relatively open canopy and have potential to develop a cottonwood gallery forest, such as Slide Creel. The mechanisms include: 1) browsing on young cottonwoods seedlings/saplings, 2) retarding cottonwood succession and large tree recruitment; and, 3) reduction in future levels of instream large wood. These mechanisms will negatively and measurably affect the Large Woody Debris NMFS MPI indicators. Negative impacts to the Refugia indicator will be minimized by use of the endpoint indicators and PDC.

## 8 ESA EFFECT DETERMINATIONS

ESA effect determination summaries are presented in Table 2 by allotment.

### ***MCR Steelhead:***

#### ***Bear Allotment***

MCR steelhead, critical habitat, and MSRA are present within the allotment. Project design criteria such as fencing or rest of riparian pastures along the MFJDR will minimize cattle-fish/fish habitat interactions making effects insignificant or discountable. Consequently, the effect determinations for MCR steelhead are “May Affect, Not Likely to Adversely Affect.”

#### ***York On/Off Allotment***

MCR steelhead, critical habitat, and MSRA are present within the allotment. Although project design criteria such as indicator monitoring and move triggers will lead toward riparian recovery, effects to habitat or the species will still occur that are not insignificant or discountable. Particularly, meaningfully measurable impacts to the sediment, temperature, large woody debris, and substrate primary constituent elements are expected to occur. Consequently, the effect determinations for MCR steelhead are “May Affect, Likely to Adversely Affect.”

### ***CR bull trout:***

#### ***Bear Allotment***

At certain times of year, CR bull trout are expected to migrate through the allotment in the MFJDR, which also contains bull trout critical habitat. Project design criteria such as fencing or rest of riparian pastures along the MFJDR will minimize cattle-fish/fish habitat interactions

making effects insignificant or discountable. Consequently, the effect determinations for CR bull trout are “May Affect, Not Likely to Adversely Affect.”

#### *York On/Off Allotment*

No CR bull trout, critical habitat, or streams present within the Allotment. Consequently, the effect determinations for CR bull trout are “No Effect.”

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### 8.1 RATIONALE

The PCEs are the physical or biological features of critical habitat essential to the conservation of the species. For MCR steelhead PCE 1 (Freshwater spawning sites), the analysis determined that in the York On/Off Allotment there were **negative and measurable effects** to the *temperature* and *sediment* indicators corresponding to the *water quality* feature of the PCE, and the *substrate embeddedness* indicator corresponding to the *suitable substrate* feature of the PCE, as diagrammed below:

- PCE1: Freshwater spawning sites.
  - Water quality PCE feature
    - Temperature indicator
    - Sediment indicator
  - Suitable substrate PCE feature
    - Substrate embeddedness indicator

In addition, for PCE 2 (Freshwater rearing sites), the analysis determined that in the York On/Off Allotment there were **negative and measurable effects** to the *temperature* and *sediment* indicators corresponding to the *water quality* feature of the PCE, the *substrate embeddedness* indicator corresponding to the *forage* feature of the PCE, and the *refugia* and *large woody debris* indicator for both the *forage* and *natural cover* features of the PCE, as diagrammed below:

- PCE2: Freshwater rearing sites.
  - Water quality PCE feature
    - Temperature indicator
    - Sediment indicator
  - Forage PCE feature
    - Substrate embeddedness indicator
    - Large woody debris indicator
    - Refugia indicator
  - Natural cover PCE feature
    - Large woody debris indicator
    - Refugia indicator

Negative measurable effects do not meet the definition of “insignificant” effects and they are not “discountable” because the effects are likely to occur. Consequently, the effect determination for MCR Steelhead designated CH overall is “May Affect, Likely to Adversely Affect.”

The same NMFS MPI indicators determined to have negative, measurable effects during the PCE analysis were brought forward in the analysis of effects to the species. The mechanisms by which the livestock use PE would affect habitat characteristics that would in turn result in measurable increases in water temperature, increased sediment and turbidity, and increased substrate embeddedness were described in detail. The biological consequences to MCR

Steelhead were also described. The conclusion was that the effects to the indicators would result in negative effects to each indicator that were measurable, and therefore did not meet the definition of “insignificant” effects. They are not “discountable” because the effects are likely to occur.

In addition, there are likely to be direct effects to individual MCR steelhead from the livestock use PE, because livestock will have access to streams during the spawning, incubation, and rearing periods of the MCR steelhead life cycle. Direct impacts are likely to occur if livestock waded into a stream and disturb rearing juveniles or spawning adults, and/or step on redds. Juveniles in close proximity to stream crossings or watering sites are likely to move out of an area when livestock enter or approach the stream. Juveniles forced into open water could have increased risk of predation. It is likely that spawning behavior will be interrupted forcing adults to retreat to nearby cover and that redds will be at risk of being stepped on. However, these risks will be minimized with the implementation of the *Malheur National Forest Strategy to Minimize Redd Trampling “Take” of Steelhead and Bull Trout* (Appendix F), which incorporates PDC 5 and 6 (Section 4.1.2). Additionally, MNF staff (range and aquatic specialists) will take extra effort to monitor these sites when they are in the field. If active redds are located, mitigation actions consistent with PDC 6 (Section 4.1.2) will be taken to eliminate or significantly minimize the potential for redd trampling,

In summary, because the proposed action will result in measurable negative effects to components of MCR steelhead habitat, with indirect effects to the species, and it is likely there will be direct negative effects to adults, juveniles and possibly incubating embryos in redds, the effect determination is “May Affect, Likely to Adversely Affect” for the species.

For the CR bull trout analysis, there were no negative and measurable effects for any of the indicators for any of the PCEs: all effects were expected to be insignificant or discountable. Therefore, the effects determination for critical habitat and the species is “May Affect, Not Likely to Adversely Affect”.

## 9 ESA CUMULATIVE EFFECTS

ESA cumulative effects are those effects of future State, tribal, local or private activities that are reasonably certain to occur in the area of the Federal action subject to consultation. Future Federal actions that are unrelated to the proposed action are not considered in this section because they are subject to separate consultation pursuant to section 7 of the ESA. There are several future State or private activities that are reasonably certain to occur.

### 9.1 ODFW ELK AND DEER MANAGEMENT

The ODFW manages Rocky Mountain elk and mule deer populations in the action area. The action area is located entirely within the state of Oregon’s Northside Wildlife Management Unit (WMU).

Elk and mule deer utilize streamside vegetation differently. Both animals eat riparian vegetation, but have different forage preferences. The diets of elk, mule deer, and cattle are very different during early summer and become increasingly similar during late summer. Cattle diets have more grasses, deer diets have more shrubs and forbs, and elk diets are in between those of cattle and deer. (USDA 2006). There is overlap between what each species will eat dependent upon

season and availability. Additionally, Coe et al. (2005) found a cascading effect of larger ungulates displacing smaller ungulates. They found that the presence of livestock displaced smaller ungulates including mule deer and elk, and that livestock chose resources such as forage before smaller ungulates.

Table 27 presents Rocky Mountain elk and mule deer management objectives (MO) and population estimates from 2004-2010 for the Northside WMU that encompasses the York On/Off allotment. The mule deer population MO was obtained from ODFW (2003), available online at: [http://www.dfw.state.or.us/wildlife/management\\_plans/docs/MuleDeerPlanFinal.PDF](http://www.dfw.state.or.us/wildlife/management_plans/docs/MuleDeerPlanFinal.PDF). Mule deer population estimates, and Rocky Mountain elk MOs and population estimates, were obtained by MNF fish biologists or contractors from ODFW wildlife biologist Ryan Torland, Joel Hurtado, and Nick Myatt (pers. comm. 2011).

TABLE 27. ROCKY MOUNTAIN ELK AND MULE DEER MANAGEMENT OBJECTIVES AND WINTER POPULATION ESTIMATES FROM 2004-2010 FOR THE NORTHSIDE WILDLIFE MANAGEMENT UNIT IN OREGON, BEAR, AND YORK ON/OFF ALLOTMENTS.

Year	Northside Wildlife Management Unit	
	Elk Mgmt. Objective (2,000)	Deer Mgmt. Objective (15,500)
2004	2,000	7,950
2005	2,000	7,954
2006	2,000	8,137
2007	2,400	7,358
2008	2,500	7,325
2009	2,500	7,085
2010	2,500	7,228

ODFW has managed the elk population of the Northside WMU at the population MO of 2,000. Beginning in 2007, the elk population has exceeded the MO for four consecutive years within the Northside Unit. The mule deer population MO was not exceeded during 2004-2011 in either WMU and has remained below. It is not known what elk or mule deer population level in the WMU would result in detectable effects to the PCEs of CH for MCR steelhead or CR bull trout.

There are data available regarding combined wild ungulate use (i.e., deer and elk) in specific pastures nearby the allotments within the action area that were not grazed in 2009 and 2010. The MNF proposed grazing use endpoints were approached in a few instances. In 2010 in the Belshaw Riparian Pasture of the Mt. Vernon Allotment, the percent woody browse use was estimated at 37.6 percent when the endpoint is 40 percent. Similarly in 2010 for the Lower Creek Pasture of the John Day Allotment, the percent woody browse use was estimated at 38.6

percent. Percent streambank alteration was estimated at 15 percent at the Bear Creek Pasture in 2010, which was five percentage points below the proposed MNF endpoint of 20 percent. Had livestock been turned out in these allotments, their use would have been limited, otherwise the endpoints might have been exceeded.

Because these data were from areas nearby the Bear, and York On/Off allotments, it is likely wild ungulate use also occurs in these three allotments.

There is a potential for cumulative effects to MCR steelhead and CR bull trout designated CH from use by wild ungulates. Such effects are identical to those described in the effects to MCR steelhead CH section: (1) increased sediment in stream channels resulting in increased turbidity, substrate embeddedness, a reduction in macroinvertebrate production, and reduced quality of spawning gravel; (2) and an increase in water temperature as a result of shade loss along stream channels from grazing/browsing of riparian vegetation.

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## 9.2 UNAUTHORIZED LIVESTOCK GRAZING

Unauthorized livestock grazing has occurred in the allotments, and is reasonably certain to occur in the future. However, it is not a significant problem in these three allotments. Typically, it involves a few cow/calf pairs every few years. The unauthorized use has been resolved in a timely manner once MNF staff notified livestock owners. As long as the MNF takes timely action whenever trespass occurs, habitat degradation is likely to be minimized.

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## 9.3 ACTIONS ON PRIVATE PROPERTY

The ESA action area includes private property inholdings. There is the potential for these properties to be developed. However, there is no information on specific proposals at this time. The effects to PCEs of CH (if any) of activities on private property, such as cattle grazing, are expected to continue at the same rate as they have been.

# 10 ESSENTIAL FISH HABITAT FOR CHINOOK SALMON

The Magnuson-Stevens Act, as amended by the Sustainable Fisheries Act of 1996 (Public Law 104-267), established procedures designed to identify, conserve, and enhance Essential Fish Habitat (EFH) for those species regulated under a Federal fisheries management plan. The MSA requires Federal agencies to consult with NOAA Fisheries on all actions, or proposed actions, authorized, funded, or undertaken by the agency, that may adversely affect EFH.

The Pacific Fisheries Management Council (PFMC) is one of eight regional fishery management councils established under the Magnuson-Stevens Act. PFMC develops and carries out fisheries management plans for salmon, groundfish and coastal pelagic species off the coasts of Washington, Oregon, and California, and recommends Pacific halibut harvest regulations to the International Pacific Halibut Commission.

As required by the Magnuson-Stevens Act, the PFMC described and identified Essential Fish Habitat (EFH) in each of its fisheries management plans. The EFH includes “those waters and substrates necessary to fish for spawning, breeding, feeding, or growth to maturity.” All streams, lakes, ponds, wetlands, and other water bodies currently, or historically accessible to salmon in Washington, Oregon, Idaho, and California are designated as EFH for affected salmon stocks with management plans. The Middle Fork John Day 4<sup>th</sup> field HUC (HUC 17070203), which

encompasses the project area, has been designated as EFH for Chinook salmon (73 FR 200:60987 October 15, 2008). However, finer resolution of what constitutes waters “currently or historically accessible to salmon” is dependent upon local information.

The Magnuson-Stevens Act (MSA) also established an EFH consultation process. Federal agencies are required to consult with NMFS on all actions that may adversely affect EFH. The NMFS interprets the scope of these consultations to include actions by Federal agencies that occur outside designated EFH, such as upstream or upslope, but which nonetheless may have an adverse effect on habitat conditions necessary for the long-term survival of the species within EFH. NMFS must provide conservation recommendations for any Federal or State activity that may adversely affect EFH. Within 30 days of receiving EFH conservation recommendations from NMFS, Federal agencies must conclude EFH consultation by responding to NMFS with a written description of conservation measures the agency will use to avoid, mitigate or offset the impact of its action on EFH. If the Federal agency selects conservation measures, which are inconsistent with the conservation recommendations of NMFS, the Federal agency must explain in writing its reasons for not following NMFS recommendations.

The MNF searched for information to determine if the action under EFH consultation includes areas currently or historically accessible to Spring Chinook salmon. Regarding current use, an ODFW website provides access to maps titled *Spring Chinook Habitat: Susanville Quad*, *Spring Chinook Habitat: Wildcat Point Quad*. The Chinook salmon distribution maps depict use of the MFJD River, and the lower portion of Slide Creek on private land about 3 miles downstream of the York allotment boundary. The ODFW maps are accessible at: [http://nrimp.dfw.state.or.us/nrimp/default.aspx?pn=chs\\_dist](http://nrimp.dfw.state.or.us/nrimp/default.aspx?pn=chs_dist)).

The analysis of effects to designated CH for MCR steelhead concluded that there were measurable negative effects to several PCEs in the York Allotment. These effects would be a proxy for effects to Chinook salmon EFH as they have similar habitat requirements. However, there is no information to support that Chinook salmon historically or currently use Slide Creek in the York Allotment. The nearest Chinook salmon habitat in Slide Creek is about 3 miles downstream from the MNF boundary of the York allotment according to ODFW data. The issue at hand is whether meaningfully measurable habitat impacts to MCR steelhead in Slide Creek would be detectable 3 miles downstream where spring Chinook habitat occurs. Because 0.88 miles of Slide Creek exist in the York Allotment and because the proposed action is reducing stream shading through removal of herbaceous vegetation and likely increasing summer water temperature throughout much of these 0.88 miles (the lower portion of Slide Creek in the allotment is exclosed), it is not unreasonable to expect these increased water temperatures to be detected 3 miles downstream in spring Chinook salmon habitat. Therefore, the determination for spring Chinook EFH in the York Allotment is “May Adversely Affect”. In the Bear Allotment, the MNF concludes that the proposed action “Will Not Adversely Affect” EFH for MSA-managed Chinook salmon (Table 2) because of comparable effects to MCR steelhead due to riparian exclosures or non-use along the MFJDR.



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